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**INTERAÇÕES ENTRE SERES HUMANOS E PLANTAS COMO
DETERMINANTES PARA MONTAGEM DE COMUNIDADES VEGETAIS DA
CAATINGA**

Campina Grande, PB

2023

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CAATINGA**

Tese apresentada ao Programa de Pós-Graduação em
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Estadual da Paraíba, Federal de Pernambuco, como
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**INTERAÇÕES ENTRE SERES HUMANOS E PLANTAS COMO
DETERMINANTES PARA MONTAGEM DE COMUNIDADES VEGETAIS DA
CAATINGA**

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2023

Filha de agricultores, trabalhando com agricultores!

Dedico

Aos meus pais, Maria da Paz B. Ramos e João de Medeiros Ramos.
Em especial ao meu pai. Sua ausência deixou um vazio e a vida às vezes
parece já não ter tanto sentido.

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RESUMO

A complexa dinâmica dos ecossistemas, dada pela variação ao longo do tempo e espaço nas condições bióticas e abióticas, tem sido traduzida pela Ecologia, que ao longo do tempo tem ganhado força, complexidade e conectividade com diversas outras ciências, numa busca incessante em compreender os padrões que regem a abundância, composição e diversidade de espécies. No entanto, só recentemente, o papel dos seres humanos começou a ser incorporado nas explicações e previsões sobre essa dinâmica dos sistemas. Dentro desse contexto, o objetivo central desta tese foi compreender como ocorre a montagem de comunidades vegetais ao longo de gradientes climáticos submetidos a distintos distúrbios antrópicos crônicos no semiárido brasileiro. Para isso, buscamos também, integrar o conhecimento ecológico de populações humanas locais, sobre a utilização de espécies vegetais aos aspectos da Ecologia de comunidades vegetais, para determinar como características morfológicas e fisiológicas (traços funcionais) das plantas influenciam a seleção e uso de espécies por populações humanas locais. Nossos resultados revelam que as relações de uso e seleção de espécies tem severos efeitos na dinâmica dos sistemas sócio-ecológicos. Populações humanas locais selecionam as espécies vegetais de acordo com características importantes, como densidade da madeira, área foliar específica, espessura foliar. Tal seleção exerce uma forte pressão de uso nas espécies mais desejáveis, o que pode afetar negativamente a dinâmica e resiliência do sistema. Nesse sentido, as comunidades vegetais dos estratos adulto e regenerante encontram-se com o desenvolvimento comprometidos, o que por sua vez, afetará o fornecimento de bens e serviços ecossistêmicos para as sociedades humanas, especialmente, aqueles que dependem diretamente desses recursos. A dinâmica das comunidades vegetais da Caatinga é portanto, um complexo resultante da interação entre fatores abióticos juntamente com a influência dos seres humanos que são agentes capazes de influenciar o funcionamento do ecossistema.

Palavras chave: Florestas secas, montagem de comunidades, conhecimento ecológico local, distúrbio antrópico crônico, precipitação, características funcionais.

1. INTRODUÇÃO GERAL

1.1 Objetivos e Questionamentos

A Caatinga é sem dúvida um ecossistema fascinante, e aprendi a admirá-la desde muito cedo. Desde a infância estive imersa dentro da Caatinga, fazendo parte desse ecossistema tão diverso, sou filha de agricultores do Cariri paraibano. Vivo até aqui as experiências, não apenas como uma jovem mulher cientista, mas antes de tudo sou uma caririzeira, e essas são as raízes que me sustentaram até aqui e sustentarão até o fim do ciclo terrestre. Os conhecimentos, práticas e crenças que antecedem o meu eu cientista exerceram forte influência no rumo dos meus interesses de pesquisa. Desde a infância, acompanhei meus pais na lida com as plantações (que acompanham o período chuvoso) e com a criação de rebanhos. Os conhecimentos, práticas, e os motivos pelos quais meus pais usavam determinadas plantas, enquanto outras eram excluídas sempre foi algo que me chamou atenção. Talvez era algo inato à minha pessoa, talvez foi algo que aprendi a admirar, através da observação, da relação intrínseca que meus pais desenvolveram com tudo que os cercava.

Prever como as atividades humanas, incluindo a seleção de espécies por sociedades rurais, influencia a montagem de comunidades vegetais pode ser um campo complexo. Tendo em vista que diversos fatores estão atuando ao longo do tempo e espaço. Isso se torna ainda mais complexo em ambientes heterogêneos, como é a Caatinga. As pesquisas na Ecologia tem seguido esse caminho e, um arcabouço teórico tem se formado, mostrando que o caminho para as respostas dos distúrbios antrópicos e perdas de biodiversidade é o modelo de sistemas socioecológicos (BENNETT *et al.*, 2015; DÍAZ *et al.*, 2011; LOPES, 2017). Uma estrutura útil para entender, prever e abordar a dinâmica de um “clima humano”, no qual as sociedades humanas interagem diretamente com a biodiversidade e o ambiente em que vivem, em escala local e regional (HUDSON *et al.*, 2020). Há reconhecimento de que ao longo de sua evolução, as sociedades humanas atuaram moldando os ecossistemas em que vivem. A medida que algumas espécies são selecionadas enquanto outras são excluídas, os humanos atuam como agentes causadores de mudanças (ELLIS, 2011).

Por um lado, sabe-se que as alterações ambientais causadas por atividades antrópicas conduzem mudanças severas na biodiversidade em todo o planeta. O que conseqüentemente, resulta em alterações nas propriedades dos ecossistemas, que por sua vez, afetam os benefícios que os humanos podem obter dos ecossistemas. Por outro lado, tais alterações resultam também em uma maior desigualdade social, isso porque as populações humanas que dependem

35 diretamente da biodiversidade, têm o acesso e escolha desses recursos reduzidos, portanto, são
36 essas as populações as mais afetadas pela perda de biodiversidade (DÍAZ *et al.*, 2006). Nesse
37 cenário, há uma convergência para a necessidade de uma aliança entre distintas ciências, a fim
38 de traçar caminhos para entender as ligações mútuas entre sistemas naturais e sociais e, assim,
39 enfrentar os desafios da sustentabilidade (FIGUEROA; CALZADA; MEAVE, 2020). De fato,
40 abordagens interdisciplinares exigem um acentuado esforço de pesquisa e, por vezes, os
41 pesquisadores enfrentam uma série de desafios e barreiras para conduzir esse tipo de pesquisas
42 (LOPES, 2017). Entretanto, essas abordagens permitiram traçar estratégias para melhoria na
43 qualidade de vida de todos que compõem o planeta Terra (DÍAZ *et al.*, 2011, 2018).

44 É preciso considerar que os desejos, as percepções e capacidades para selecionar
45 componentes da biodiversidade variam dentro e entre os grupos humanos e, portanto, afetam
46 de maneira distinta a biodiversidade (ALBUQUERQUE *et al.*, 2018; CÁCERES *et al.*, 2015).
47 A abordagem através da teoria dos sistemas socioecológicos é uma ferramenta útil para
48 entender, prever e abordar a dinâmica dos ecossistemas (FOLKE, 2006). Uma vez que, através
49 dessa abordagem, é possível analisar como as sociedades humanas interagem diretamente com
50 os demais componentes dos ecossistemas, em escala local e regional (FERNÁNDEZ-
51 GIMÉNEZ *et al.*, 2018; FOLKE, 2006).

52 Dessa forma, na presente tese busquei compreender como ocorre a montagem de
53 comunidades vegetais ao longo de gradientes climáticos submetidos a distintos distúrbios
54 antrópicos crônicos no semiárido brasileiro. Para alcançar uma compreensão mais detalhada
55 sobre as interações entre seres humanos e plantas, integrei o conhecimento ecológico de
56 populações humanas locais, sobre a utilização de espécies vegetais, aos aspectos da Ecologia
57 de comunidades vegetais para determinar como características morfológicas e fisiológicas
58 (traços funcionais) das plantas influenciam a seleção e uso de espécies por populações humanas
59 da Caatinga. Dessa forma, busquei responder as seguintes questões: Como os distúrbios
60 antrópicos crônicos afetam a estrutura, composição e diversidade dos componentes (estratos)
61 da vegetação adulta e regenerante ao longo dos gradientes climáticos na Caatinga?; (2) Como
62 as variáveis climáticas e ambientais interagem com os distúrbios antrópicos crônicos? (3) Como
63 características funcionais das plantas influenciam a seleção de plantas por populações humanas
64 rurais?. A partir das respostas obtidas com esses questionamentos tornou-se possível elaborar
65 um cenário detalhado sobre as respostas das comunidades vegetais as distintas combinações
66 entre condições climáticas e distúrbios antrópicos crônicos. De modo que, ficaram evidentes
67 algumas ameaças aos sistemas socioecológicos, bem como a resiliência da Caatinga. Além
68 disso, chegamos a importantes achados sobre o papel das características funcionais na seleção

69 de plantas por populações humanas locais. Dessa forma, nesta tese estão apresentadas
70 discussões importantes sobre os processos de montagem de comunidades vegetais, resiliência
71 e a funcionalidade do ecossistema, todos sob a perspectiva da inclusão dos sistemas
72 socioecológicos.

73

74 **1.2 Estratégias de Pesquisa**

75

76 Para alcançar nossos objetivos de pesquisa, buscamos elaborar um desenho amostral
77 que atendesse o levantamento da vegetação, de variáveis ambientais, distúrbios antrópicos, bem
78 como o levantamento etnobotânico. Inicialmente para construção do desenho amostral da
79 pesquisa, foram elaborados mapas de precipitação e de cobertura vegetal. Para elaboração dos
80 mapas, coletamos dados climáticos e imagens de satélite que cobrissem toda porção da
81 microregião do Cariri. Para elaborar os mapas de cobertura vegetal, utilizamos o índice de
82 vegetação por diferença normalizada (NDVI), que tem como *Proxy* estimar a biomassa e
83 produtividade da vegetação, através do cálculo: $NDVI = (Infra\ Vermelho - Vermelho) / (Infra$
84 $Vermelho + Vermelho)$ (ROUSE *et al.*, 1973), dado pelo Software Argis. O NDVI foi obtido a
85 partir de imagens das bandas 4 e 5 do satélite Landsat 8, referentes a estação seca do ano de
86 2017 e obtidas no site earthexplorer.usgs.gov. As bandas correspondem a proporção de luz
87 refletidas pela vegetação detectadas por sensores, sendo designadas respectivamente como
88 vermelho e infravermelho próximo. As imagens utilizadas foram registros da estação seca, pois
89 é o momento mais favorável para mapear a vegetação de regiões secas. As imagens se referem
90 ao ano de 2017. Os dados climáticos para região foram obtidos pelo banco mundial de dados
91 climáticos *Word Clim*.

92 Após a elaboração dos mapas, foram definidas as áreas de interesse, incluindo áreas que
93 estivessem localizadas em distintos níveis de precipitação, bem como áreas com diferentes
94 níveis de cobertura vegetal. Buscamos considerar a variação ambiental em virtude da
95 heterogeneidade ambiental da Caatinga. Posteriormente, a seleção de todos os pontos foi
96 avaliada com visitas e observação de campo. Alguns pontos não condiziam com a cobertura
97 vegetal esperada, por exemplo, ou áreas que sofreram alguma perturbação aguda e que tiveram
98 toda cobertura vegetal retirada, foram descartados e selecionados novos pontos de interesse.
99 Como uma restrição para seleção dos pontos amostrais, estabelecemos uma distância mínima
100 de um raio de 2km entre os pontos. Consideramos esse tamanho do *buffer* para evitar a
101 sobreposição e assim aumentar a independência entre as áreas. As áreas que investigamos estão
102 todas localizadas na microregião do Cariri, porção central do estado da Paraíba, embora, em

103 escala macro, essa região tenha algumas características em comum, em nossas pesquisas e
104 viagens a campo observamos uma variação enorme nas formas vegetais, nos tipos de solo
105 (dados não publicados) nos níveis e formas de pressão antrópica, bem como nas diferentes
106 cosmovisões das comunidades locais da região estudada. Ao final, selecionamos um total de
107 21 áreas.

108 Para o levantamento da vegetação em cada área foram delimitadas parcelas de 50 por
109 20 metros. Realizamos o levantamento florístico e fitossociológico para o estrato adulto e
110 regenerante. Para o estrato adulto, registramos também os indivíduos mortos. Em cada parcela
111 sempre que possível, foram identificadas todas as espécies e coletado material vegetativo para
112 identificação botânica, que ocorreu com auxílio de literatura e especialista na área (Figura 1).
113 Afim de avaliar o efeito do aporte de serrapilheira no estrato regenerante, em todas as parcelas
114 foram mensuradas o estoque de serrapilheira a partir da coleta de toda necromassa presente em
115 cinco pontos de 1m x 1m (1m²) cada, distribuídos aleatoriamente em cada parcela.

116 Para obtermos um gradiente de intensidade de distúrbio antrópicos, em cada parcela
117 registramos métricas descritas como importantes indicadoras de distúrbios antropogênicos
118 crônicos em Florestas Tropicais Secas (MARTORELL e PETERS, 2005). Utilizamos algumas
119 métricas de medidas diretas que foram mensuradas na parcela e medidas indiretas baseadas no
120 contexto geográfico que foram obtidas com auxílio de imagens de satélite utilizando o Software
121 ArcGis 10.3. Com base nessas métricas, identificamos dois principais tipos de perturbação para
122 caracterizar índices de distúrbio antrópico crônico: (1) Pressão do gado. Este distúrbio está
123 relacionado à atividade dos herbívoros domésticos (gado, cavalos, burros, cabras e ovelhas) que
124 pastejam e causam danos físicos no solo através do pisoteio nas parcelas; (2) Atividades
125 humanas: este tipo de distúrbio se refere a marcas do uso da terra pelas pessoas, e são medidas
126 indiretas relacionadas ao contexto geográfico. Utilizamos imagens de satélite para estimar os
127 indicadores de pressão humana. Para coleta de dados sobre o conhecimento ecológico local
128 selecionamos cinco áreas, de modo que abrangessem distintos níveis de cobertura vegetal e
129 precipitação. Inicialmente mantivemos um primeiro contato com alguns moradores locais,
130 nesse momento explicamos para eles os objetivos e como estava estruturada nossa pesquisa.



131

132 **Figura 1.** Primeira etapa da pesquisa. (A, B) Mensuração de impactos antrópicos. (C, D) Levantamento florístico e fitossociológico. (E, F) Processo de herborização de material vegetal para identificação.

134 Recorremos às associações rurais e agentes de saúde das comunidades rurais para
135 fortalecer e garantir uma maior confiabilidade nas informações obtidas entre as populações
136 locais e os pesquisadores. Para coleta de dados sobre o conhecimento ecológico local utilizamos
137 um conjunto de técnicas complementares a fim de extrair o máximo das informações dos
138 moradores locais (Figura 2). Inicialmente, fizemos perguntas sobre quais plantas eles
139 conheciam e utilizavam para as diferentes categorias de uso que listamos. Em seguida,
140 utilizávamos as técnicas da nova leitura, turnê guiada e recurso visual (ALBUQUERQUE *et*
141 *al.*, 2014). A técnica de nova leitura consistiu em o pesquisador enunciar oralmente as plantas
142 citadas pelo participante, de forma a propiciar a recordação de mais plantas (ALBUQUERQUE
143 *et al.*, 2014). Utilizamos um recurso visual de fotografias em dois momentos diferentes:
144 primeiro, para aqueles participantes que afirmaram não conhecer as plantas; e segundo, para os
145 participantes que mencionaram esquecer no momento das entrevistas (MEDEIROS *et al.*,
146 2014). As fotografias tinham imagens de frutas, folhas, flores e caules de plantas com o hábito
147 de árvores e arbustos que foram registradas durante o levantamento da vegetação nas
148 comunidades. Também realizamos visitas guiadas para verificar a identidade botânica das
149 espécies mencionadas pelos participantes e para coletar material taxonômico
150 (ALBUQUERQUE *et al.*, 2014).



151
 152 **Figura 2.** Levantamento de dados etnobotânicos. (A, B, C) Lista livre e técnicas da nova leitura. (D) Uso de
 153 recurso visual (fotografias). (E, F, G, H) Turnê guiada. (I) Equipe de pesquisadores com agropastor da comunidade
 154 Curral do meio, município de São João do Cariri, estado da Paraíba.

155 1.3 Estrutura da Tese

156 A tese é composta por cinco capítulos. De acordo com o modelo adotado pelo Programa de
 157 Pós-Graduação em Etnobiologia e Conservação da Natureza, o primeiro deles denomina-se
 158 “Fundamentação Teórica”, onde apresento as principais temáticas envolvidas na tese. O
 159 capítulo II intitulado de “*Environmental factors modulate plant selection by local human*
 160 *populations in dry tropical forests*” corresponde a um capítulo de livro que foi publicado no
 161 livro *Ethnobotany of the Mountain Regions of Brazil* da editora Springer. O capítulo III que tem
 162 como título: “*Functional traits influencing selection and use of plant species by local human*
 163 *populations in the tropical dry Caatinga forest*” foi submetido a revista *Economic Botany*. O
 164 capítulo IV intitulado: “*The role of chronic anthropogenic disturbances in plant community*
 165 *assembly along a water stress gradient in Brazil’s semiarid Caatinga region*” que foi publicado
 166 na revista *Forest Ecology and Management*. O capítulo V, que aborda as principais
 167 considerações finais da presente tese. E por fim, nos apêndices, apresento o registro de
 168 ocorrência de uma espécie vegetal encontrada nos levantamentos fitossociológicos. O artigo

169 foi publicado na revista *Acta brasiliensis* e tem como título: “*First record of Cnidoscolus*
170 *obtusifolius Pohl (Euphorbiaceae) for Paraíba State , northeastern Brazil*”

171 No segundo capítulo buscamos contextualizar e reunir algumas evidências sobre os efeitos
172 de variáveis ambientais na seleção de espécies de plantas que populações humanas locais
173 realizam. No terceiro capítulo, consideramos que o uso e seleção de espécies vegetais é um
174 importante *driver* de mudanças nos ecossistemas, dessa forma compreender a dinâmica que
175 rege esse processo é fundamental para compreender a dinâmica dos sistemas sociais e
176 ecológicos. Dessa forma, buscamos integrar o conhecimento ecológico de populações humanas
177 sobre a utilização de espécies vegetais, aos aspectos da Ecologia de comunidades vegetais, para
178 compreender os fatores que influenciam a seleção de espécies de plantas pelas populações
179 humanas locais. Demonstramos que os agropastores selecionam e utilizam as espécies de
180 acordo com um conjunto de características mais adequadas ao fornecimento dos diferentes
181 serviços ecossistêmicos que eles desejam. Portanto, as características funcionais das espécies
182 de plantas são fatores importantes na seleção de espécies vegetais.

183 No quarto capítulo, buscamos verificar como os distúrbios antrópicos crônicos modulam
184 a estrutura, composição e diversidade de comunidades vegetais adultos e regenerantes
185 submetidas a distintas condições climáticas na Caatinga. Demonstramos que os distúrbios
186 antrópicos crônicos são mais severos nas áreas mais áridas. A limitação hídrica dificulta a
187 manutenção da dinâmica natural e restabelecimento das espécies após os distúrbios. O que
188 resulta em uma comunidade menos desenvolvida, com baixa diversidade. O índice de pressão
189 do gado foi negativamente relacionado ao diâmetro ao nível do solo, área basal, e diversidade
190 no estrato adulto. Para o estrato regenerante, o estoque de serapilheira teve efeito mais forte na
191 estrutura da comunidade regenerante, que é composto, principalmente, por espécies mais
192 resistentes a distúrbios. Esses resultados revelam que a interação entre os distúrbios antrópicos
193 crônicos e a baixa disponibilidade de água pode ter consequências drásticas na dinâmica da
194 Caatinga.

195 E por fim, embora não esteja nos objetivos centrais da tese, apresento aqui um novo
196 registro de ocorrência de uma espécie vegetal para o estado da Paraíba. Esse artigo surgiu a
197 partir do levantamento fitossociológico, uma das etapas fundamentais desta tese. A partir desse
198 achado fica evidente a importância dos levantamentos florísticos e fitossociológicos na região.
199 A Caatinga compreende uma vasta área, com território equivalente a alguns países europeus, e
200 elevada heterogeneidade ambiental, isso assegura que ainda temos muito a descobrir sobre a
201 Caatinga, esse ecossistema tão heterogêneo e diverso, no contexto ecológico e social.

202

203 **Abaixo descrito estão as referências bibliográficas dos produtos publicados da tese:**

204

205 **Capítulo II:** Ramos, M.B., Pedrosa, K.M., de Faria Lopes, S., 2022. Environmental Factors
206 Modulate Plant Selection by Local Human Populations in Dry Tropical Forests. Springer,
207 Cham, pp. 1–13. https://doi.org/10.1007/978-3-030-47254-2_115-2.

208

209 **Capítulo IV:** Ramos, M.B., Maciel, M.G.R., Cunha, S.S. da, de Souza, S.M., Pedrosa, K.M.,
210 de Souza, J.J.L.L., González, E.J., Meave, J.A., Lopes, S. de F., 2023. The role of chronic
211 anthropogenic disturbances in plant community assembly along a water availability gradient in
212 Brazil's semiarid Caatinga region. For. Ecol. Manage. 538, 120980.
213 <https://doi.org/10.1016/J.FORECO.2023.120980>

214

215 **Nota:** Ramos, M.B., Maciel, M.G.R., De Melo, J.I.M., Lopes, S. de F., 2020. First record of
216 *Cnidoscolus obtusifolius* Pohl (Euphorbiaceae) for Paraíba State , northeastern Brazil. Acta
217 Bras. 4, 187–190.

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231 2. CAPÍTULO I: FUNDAMENTAÇÃO TEÓRICA

232

233 As comunidades ecológicas são definidas como uma organização no espaço e no tempo,
234 compostas por um conjunto de populações que estão interconectados através das suas
235 interações. Tais interações determinam os padrões de abundância, riqueza, composição, e por
236 fim, o funcionamento das comunidades (VELLEND, 2010). A maneira como as espécies estão
237 distribuídas no ambiente está diretamente ligada aos mecanismos de utilização dos recursos, e
238 adaptação as condições ambientais (SILVERTOWN, 2004), além da influência de fatores
239 biológicos, tais como predação, competição e a interferência antropogênica (DORNELAS,
240 2010). Todos esses fatores variam no espaço e no tempo e atuam sinergicamente na distribuição
241 das espécies e na estruturação das comunidades vegetais (MEYNARD *et al.*, 2013; SANDE,
242 VAN DER *et al.*, 2017; SILVA e SOUZA, 2018).

243 Apesar do vasto arcabouço teórico e empírico acumulado a respeito dos padrões de
244 estruturação das comunidades, muitas lacunas permanecem abertas. A exemplo, do papel dos
245 seres humanos como agentes de mudanças, que só recentemente passou a ser discutido, e
246 apontado como importante covariável em todos os processos doravante apresentados
247 (BOWLER *et al.*, 2020). Além disso, grande parte dos esforços se concentraram em florestas
248 úmidas, o que permitiu a construção de um entendimento sobre os mecanismos e padrões que
249 governam desde a dinâmica e crescimento das populações, até o desenvolvimento e estrutura
250 das comunidades. No entanto, muitos desses padrões não podem ser aplicados efetivamente a
251 outros ecossistemas, por exemplo as Florestas Tropicais Sazonalmente Secas (FTSS). Por essas
252 razões, compreender os padrões de distribuição das espécies nos mais variados ambientes
253 continua sendo uma questão fundamental na Ecologia vegetal.

254 As Florestas Tropicais Sazonalmente Secas (FTSS) são consideradas importantes
255 componentes da vegetação tropical, possuem um conjunto de características particulares, a
256 exemplo dos baixos regimes hídricos (<1.000 mm), acentuada sazonalidade, com estação seca
257 que dura de 3 a 8 meses, além de altas taxas de evapotranspiração potencial (DEXTER *et al.*,
258 2018; MORO *et al.*, 2015), o que resulta em um conjunto de particularidades na estrutura,
259 composição de espécies e dinâmica do sistema. Ao mesmo tempo em que a dinâmica hídrica
260 exerce efeito modulador nesses ambientes, os distúrbios antrópicos crônicos atuam
261 potencializando o stress hídrico (GIBB *et al.*, 2015) tornando o ecossistema vulnerável, devido
262 as lentas taxas de recrutamento e crescimento dos vegetais (CALZADA *et al.*, 2018). No
263 contexto de mudanças climáticas globais, até 2100 as Florestas Secas passarão por um aumento
264 na evaporação e temperatura médias e, conseqüentemente, secas mais longas e severas que

265 podem reduzir a produção líquida primária, induzir mudanças no crescimento anual da
266 população e aumentar a probabilidade de extinção de espécies (SARMENTO; FRANCA,
267 2018).

268 Dentre as FTSS, a Caatinga é uma das maiores regiões semiáridas tropicais do mundo,
269 abrangendo um polígono com aproximadamente 912,500 km² do território brasileiro (MORO
270 *et al.*, 2015). Compreende um mosaico vegetacional dotado de uma grande variedade de
271 fisionomias, condicionadas por uma série de fatores físicos e microclimáticos (APGAUA *et al.*,
272 2015; MORO *et al.*, 2016). A forte sazonalidade, com altas taxas de radiação solar,
273 temperaturas elevadas, altos índices de evapotranspiração e baixos e irregulares índices
274 pluviométricos caracterizam os principais determinantes para os padrões de crescimento e
275 reprodução dos organismos vegetais (SILVA e SOUZA, 2018), atuando diretamente nos
276 processos de regeneração e podendo limitar o recrutamento de árvores (RIBEIRO *et al.*, 2015).

277 Outra importante característica da Caatinga, é que apesar da severidade climática, cerca
278 de 27 milhões de pessoas vivem nessa região, dependendo direta ou indiretamente dos recursos
279 naturais, o que conseqüentemente remete no alto grau de degradação da região
280 (ANTONGIOVANNI *et al.*, 2020). Considerando que as populações humanas estão espalhadas
281 nos mais diversos ambientes e paisagens do mundo, é importante compreender que o modo de
282 vida das pessoas que ocupam um determinado espaço natural pode fornecer informações
283 imprescindíveis para estratégias de conservação da biodiversidade (HUNTINGTON 2011) e,
284 nesse contexto, o conhecimento ecológico local (CEL) tem sido cada vez mais reconhecido
285 como um componente importante para direcionar ações conservacionistas (BERKES;
286 COLDING; FOLKE, 2000; LOPES, 2017). O conhecimento ecológico local se refere ao corpo
287 de conhecimentos, práticas e crenças, que as populações humanas adquiriram partir da
288 observação, uso e manejo dos recursos (BERKES *et al.* 2000). Tais conhecimentos podem ser
289 transmitidos culturalmente através das gerações e pode ser utilizado para entender como as
290 sociedades humanas interagem com a natureza, levando em conta suas estratégias e interesses
291 (FOLKE, 2000). Conectar o CEL ao Ecologia pode portanto, ajudar a desenvolver, de modo
292 mais eficiente, soluções para o manejo dos recursos e garantir a sustentabilidade dos sistemas
293 sócio ecológicos (BERKES; COLDING; FOLKE, 2000; DÍAZ *et al.*, 2018; FOLKE, 2006).

294 Algumas pesquisas têm apontado que as populações humanas são capazes de reconhecer
295 e selecionar indiretamente grupos de plantas com características similares, a depender do fim
296 de sua utilização (BROWN *et al.*, 2011; CÁMARA-LERET *et al.*, 2017; PEDROSA *et al.*,
297 2021). Nota-se uma preferência por grupos de plantas com atributos funcionais que favorecem
298 o uso da espécie para determinado fim (CÁMARA-LERET *et al.*, 2017). Espécies de interesse

299 para uso madeireiro, podem ser selecionadas por sua maior densidade, fato que reflete em um
300 maior potencial energético (FARIAS *et al.*, 2007). Além disso, destaca-se a seleção feita pelos
301 animais domesticados que acabam forrageando grupos de vegetais com maior palatabilidade,
302 reduzindo a diversidade da mesma (CUNHA *et al.*, 2022; NIU *et al.*, 2016; NUNES *et al.*,
303 2016).

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305 ***Montagem de comunidades vegetais em escala local***

306 Do ponto de vista da montagem de comunidades, compreender os fatores que regem as
307 comunidades vegetais é a base para o entendimento do funcionamento desses sistemas e, num
308 cenário de mudanças climáticas globais, é possível prever como os ecossistemas responderão
309 frente ao exponencial desenvolvimento das atividades antropogênicas (DORNELAS, 2010).
310 Alguns fatores abióticos como, o clima, atuam em macro escala determinando padrões gerais
311 de diversidade (VELLEND, 2010), por outro lado, em escala local, fatores como a altitude,
312 topografia, relevo, profundidade, fertilidade, quantidade de matéria orgânica, umidade e textura
313 do solo, atuam selecionando um conjunto de espécies com características morfofisiológicas
314 adaptadas a um conjunto específico de condições ambientais (MÉNDEZ-TORIBIO *et al.*, 2016;
315 RAMOS *et al.*, 2020)

316 Dentre as variáveis abióticas, o solo é um sistema dinâmico e complexo resultante de
317 processos históricos e da ação combinada de fatores físicos, químicos e biológicos (JENNY;
318 AMUNDSON, 1941). O solo desempenha um importante papel no crescimento, no sucesso
319 reprodutivo de plantas, no desempenho relativo de plantas coexistentes, na composição, e
320 produtividade da comunidade (GUREVITCH; SCHEINER; FOX, 2009). Por exemplo, a
321 topografia, as propriedades físicas e químicas do solo são responsáveis por controlar as
322 concentrações de nutrientes do solo, a retenção de umidade, além de apoiar ou limitar o
323 crescimento da raiz, e influenciar a quantidade de água, oxigênio, e na drenagem do solo
324 (MÉNDEZ-TORIBIO *et al.*, 2016; PUTTEN, VAN DER *et al.*, 2013). Em FTSS, o regime
325 hídrico é considerado o principal filtro ambiental, refletindo não apenas na vegetação, mas
326 também nas características dos solos, que são predominantemente pouco desenvolvidos,
327 cascalhentos, rasos, ácidos à alcalinos e eutróficos (ARAÚJO FILHO *et al.*, 2017). Como
328 resultado da baixa pluviosidade que minimiza a perda de nutrientes por lixiviação e
329 proximidade do material de origem, ocasionando maior fertilidade natural dos solos (ARAÚJO
330 FILHO *et al.*, 2017; ARRUDA *et al.*, 2015). Em FTSs, um dos atributos do solo que mais
331 influenciam as plantas é a estrutura, que está relacionada ao arranjo físico de partículas do solo.
332 Entre os atributos da estrutura do solo, a porosidade e a textura se destacam em FTSs por o

333 influenciar o fluxo de água e ar dentro do solo, bem como a retenção de água e relações
334 nutricionais (GUREVITCH; SCHEINER; FOX, 2009).

335 O entendimento do ambiente abiótico como modelador da distribuição de espécies é um
336 dos conceitos mais primários e duradouros dentro da Ecologia (KRAFT *et al.*, 2015). O
337 conceito de filtragem ambiental que temos atualmente é oriundo de estudos com vegetais e
338 postulam que apenas as espécies que possuem fenótipos específicos para se estabelecer no
339 ambiente vão persistir, enquanto outros menos adaptados a pressões do ambiente tem maior
340 probabilidade de serem excluídas (LA RIVA, DE *et al.*, 2016). Como estratégias para
341 coexistência, as plantas variam em suas características funcionais e, portanto, em sua
342 preferência às condições ambientais (DUBUIS *et al.*, 2013; LAUGHLIN *et al.*, 2015). Dentre
343 essas características funcionais, os teores de compostos fenólicos, teores de fibras,
344 concentrações de nutrientes foliares, área foliar específica, massa seca foliar, densidade de
345 madeira, altura média, diâmetro médio, respondem de maneira significativa as variações nas
346 condições abióticas, e podem, portanto, fornecer informações valiosas sobre os processos de
347 montagem das comunidades, bem como o funcionamento dos ecossistemas (MCGILL *et al.*,
348 2006; PETCHEY; GASTON, 2006).

349 Associado aos filtros ambientais e atuando em menor escala, fatores bióticos como as
350 interações ecológicas (competição, mutualismo, facilitação e herbivoria), e pressão humana são
351 determinantes na montagem das comunidades, estando diretamente relacionados a
352 diferenciação de nicho e coexistência das espécies (LA RIVA, DE *et al.*, 2016; MIGUEL *et al.*,
353 2018). No caso das interações, o consumo de vegetais por parte dos herbívoros, por um lado,
354 induz os vegetais a desenvolverem diversas estratégias de defesas, tais como, variações na
355 morfologia, composição química e resistência física (DÍAZ *et al.*, 2007). Por outro lado, a
356 pastagem por grandes herbívoros pode controlar o crescimento e a dominância de algumas
357 espécies através do pastejo (NIU *et al.*, 2016). Dessa forma, competidores mais fracos
358 conseguem coexistir com um conjunto de outras espécies, controlando e mantendo os padrões
359 de riqueza da área (NIU *et al.*, 2016). Em regiões semiáridas, as espécies perenes, investem em
360 compostos de carbono e menores concentrações de nitrogênio, o que lhes confere uma menor
361 palatabilidade (SILVA; ESPÍRITO-SANTO, MORAIS, 2015). A herbivoria por animais
362 domesticados está entre as métricas para mensuração de distúrbios antrópicos crônicos e tem
363 efeitos severos nas comunidades vegetais (MARTORELL; PETERS, 2005). Na Caatinga, esses
364 animais têm sido relatados como uma forte ameaça à riqueza e diversidade vegetal, em
365 detrimento de seu intenso forrageio sobre espécies nativas (SCHULZ *et al.*, 2016, 2019).

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369 Compreender a dinâmica das interações entre os seres humanos e meio ambiente tem sido
370 campo de investigação de diversas ciências e, requer esforço combinado e integrado dos mais
371 diversos campos científicos (ALBUQUERQUE; FERREIRA JÚNIOR, 2017; MILLENNIUM
372 ECOSYSTEM ASSESSMENT, 2005). Nessa perspectiva, a Etnobiologia emerge como a
373 ciência, que busca compreender as relações dinâmicas entre os povos, biota e ambientes
374 (ALBUQUERQUE e MEDEIROS, 2013) e que pode conectar cientistas ambientais, gestores
375 e grupos humanos que lidam diretamente com a biodiversidade (SMITH, 2009). Baseada em
376 ciências como a Ecologia, Antropologia, e Psicologia, a Etnobiologia reflete sobre as inter-
377 relações que populações humanas estabeleceram ao longo do espaço e do tempo com os
378 animais, as plantas, e o ambiente físico, levando em consideração os vieses, crenças, cultura e
379 religião dos povos em questão (ANDERSON *et al.*, 2011; WOLVERTON, 2013).

380 É sabido que ao longo de sua evolução os seres humanos manejaram diferentes formas
381 de vida e recursos naturais, moldando o ambiente e utilizando os serviços ecossistêmicos de
382 acordo com as variações sazonais, variabilidade dos preços, legislação e conflitos de interesse
383 (DÍAZ *et al.*, 2011). Dessa forma, atuam como construtores de nicho, modificando o ambiente
384 em que vivem (nicho) e o de outros organismos através de suas ações, podendo alterar também
385 os processos de seleção natural (ELLIS, 2015; ODLING-SMEE; LALAND; FELDMAN,
386 2013). O que no princípio era um meio de subsistência, hoje expressa a capacidade humana de
387 utilizar a biodiversidade, conduzindo a um crescimento exponencial nos desequilíbrios
388 ambientais, o que conseqüentemente levou a uma diminuição dos recursos naturais (ELLIS;
389 RAMANKUTTY, 2008).

390 Os distúrbios antropogênicos crônicos representam um perigo eminente às FTSS,
391 levando um decréscimo drástico da biodiversidade e conduzindo ao colapso ecossistêmico. A
392 coleta de lenha, exploração de produtos florestais não madeireiros e o pastoreio extensivo de
393 gado representam excelentes exemplos de fatores que contribuem para as alterações negativas
394 na função, composição e diversidade das FTSS, além de serem difíceis de detectar, tornando o
395 problema ainda mais crítico. Os impactos antrópicos, podem ainda levar a perda de espécies
396 com características funcionais específicas, que integram grupos funcionais responsáveis pela
397 prestação de importantes serviços ecossistêmicos (DÍAZ *et al.*, 2011; LEBRIJA-TREJOS *et al.*,
398 2010). Em adição, a remoção contínua de biomassa florestal pode alterar a estrutura da
399 vegetação (por exemplo, redução da densidade do caule e aumento da abertura do dossel), bem
400 como condições microclimáticas (por exemplo, os ambientes mais quentes com solos mais

401 degradados) (RIBEIRO *et al.*, 2015; SCHULZ *et al.*, 2016). Em conjunto, estas condições
402 podem limitar o recrutamento / reprodução de alguns grupos funcionais de plantas, enquanto
403 outros grupos podem experimentar aumentos no recrutamento e desempenho em todos os
404 estágios ontogenéticos, resultando em substituição do tipo vencedor / perdedor (TABARELLI;
405 PERES; MELO, 2012).

406 Da mesma forma que outras regiões brasileiras, a Caatinga historicamente tem sido
407 espaço para a criação de rebanhos, agricultura e extração de madeira para diversos fins, o que
408 transformou a vegetação em um conjunto de manchas altamente modificadas e com áreas em
409 processos de desertificação (ANTONGIOVANNI *et al.*, 2020; RIBEIRO *et al.*, 2015).
410 Considerando que populações humanas locais detém um vasto conhecimento acerca da
411 dinâmica dos sistemas, o entendimento das ameaças ao funcionamento desses ecossistemas
412 requer a integração do conhecimento ecológico local com as questões da Ecologia das
413 comunidades vegetais (LOPES, 2017). Diversos trabalhos foram realizados relatando uma
414 variedade de plantas utilizadas pra distintas categorias de uso (ALBUQUERQUE, Ulysses
415 Paulino *et al.*, 2015; ALMEIDA, C. F. C. B. R. DE *et al.*, 2005; C *et al.*, 2018; ESTOMBA;
416 LADIO, A.; LOZADA, 2006; FARIAS *et al.*, 2007; GAOUE *et al.*, 2016; LESO *et al.*, 2017;
417 LETCHER *et al.*, 2015; MEDEIROS, DE *et al.*, 2017; NUNES *et al.*, 2016; RAMOS *et al.*,
418 2008; SILVA OLIVEIRA, DA *et al.*, 2015).No entanto, se faz necessário uma compreensão
419 mais ampla, que mescle o conhecimento ecológico local e a Ecologia de comunidades
420 (BENNETT *et al.*, 2015; ISBELL *et al.*, 2017; LOPES, 2017; PERIAGO *et al.*, 2017).

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438 **3. CAPÍTULO II: Environmental factors modulate plant selection by local human**
439 **populations in dry tropical forests**

440 (Manuscrito publicado como capítulo do livro: Ethnobotany of Mountain Regions.
441 Ethnobotany of the Mountain Regions of Brazil – Springer

442

443 **Link de acesso para o artigo:** [https://link.springer.com/referenceworkentry/10.1007/978-3-](https://link.springer.com/referenceworkentry/10.1007/978-3-030-47254-2_115-2)
444 [030-47254-2_115-2](https://link.springer.com/referenceworkentry/10.1007/978-3-030-47254-2_115-2)

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Environmental Factors Modulate Plant Selection by Local Human Populations in Dry Tropical Forests

[Maiara Bezerra Ramos](#), [Kamila Marques Pedrosa](#) & [Sérgio de Faria Lopes](#)

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Abstract

The intensification of use of natural resources has promoted changes in ecological and evolutionary processes around the world, becoming the main threat to balance of ecosystems and socio-ecological systems. As a consequence, it is necessary to understand how human populations are interrelated with biodiversity patterns worldwide. The way in which human populations select the enormous diversity of plants for multiple uses has become the main focus of studies in ethnobiology; for this, some ecological and evolutionary theories and hypotheses have been included in ethnobiological investigations that seek to understand what mechanisms influence the use of plants. In this chapter, we contextualize a general approach to the main strategies that explain the use and selection of plants by local populations inserted in areas of seasonally dry tropical forests in Brazil.

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468 **Introduction**

469 Natural resource exploitation by human beings dates from the most remote times,
470 especially for the survival and perpetuation of our species (Clement et al. [2015](#)). However, the
471 way in which resources are exploited and selected is what marks human evolution and
472 distinguishes us from other species (Albuquerque et al. [2015](#); Ellis [2015](#)). By altering the
473 environment, human beings have acted as niche builders, modifying the environment in which
474 they live (niche) and that of other organisms, and may also alter natural selection processes
475 (Sterelny [2007](#); Odling-Smee et al. [2013](#); Ellis [2015](#)).

476 Throughout time and space, evolutionary forces have conditioned human beings to
477 adapt to different ways of life and use of natural resources (Albuquerque et al. [2018](#)). For
478 example, the domestication of plants is a phenomenon which has enabled human groups to
479 gain control over nature and predictability in provision of ecosystem services arising from
480 plants (Ellis and Ramankutty [2008](#); Anderson et al. [2011](#); Smith [2011](#); Ellis [2015](#)). However,
481 what in the beginning was a means of subsistence today expresses the human capacity to

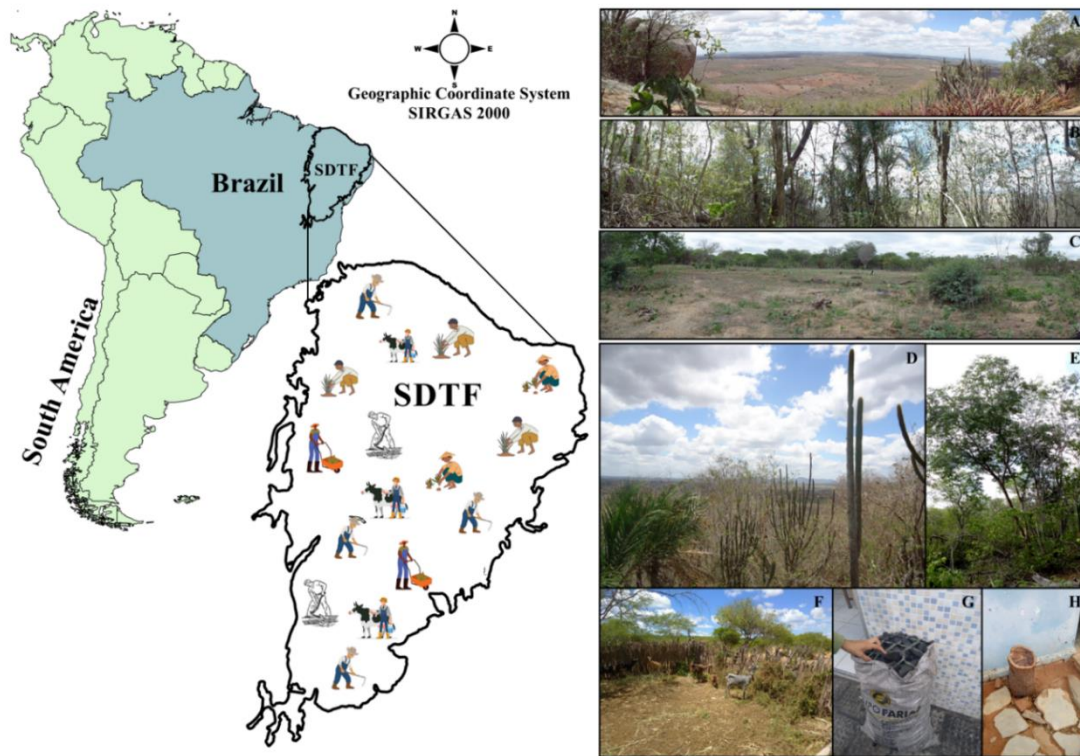
482 dominate nature, causing these relationships to increase exponentially over time, which has
483 consequently led to a continuous decrease in natural resources and changes in ecosystems
484 (Clement et al. [2015](#); Folke et al. [2021](#)).

485 Human populations are spread across the most diverse environments and landscapes in
486 the world, from areas with high biodiversity to slightly different environments (Ellis [2015](#)).
487 This suggests that local populations can exercise different forms of coexistence, use, and
488 selection of plants in view of the different existing environmental factors (Naah and Guuroh
489 [2017](#); Folke et al. [2021](#)). Ethnobiology has emerged as a science in order to understand the
490 dynamics of these interrelations, supported by biological and anthropological bases
491 (Wolverton [2013](#); Gaoue et al. [2017](#)), and which has devoted great efforts over the last few
492 years in order to understand the criteria for species selection and structuring of human behavior
493 through ecological and evolutionary bases (De Almeida et al. [2005](#); Albuquerque et al. [2015](#);
494 Cáceres et al. [2015](#); Savo et al. [2015](#); Nunes et al. [2016](#); Cámara-Leret et al. [2017](#); Naah and
495 Guuroh [2017](#); Periago et al. [2017](#); Teixidor-Toneu et al. [2018](#); Molina-Venegas et al. [2020](#);
496 Pedrosa et al. [2021](#)).

497 New hypotheses have emerged with the inclusion of mechanisms and processes for the
498 use and exploitation of resources, considering biases such as the environment and culture, as
499 well as aspects such as biological and cultural evolution to understand the dynamics of human
500 behavior in plant selection (Ellis and Ramankutty [2008](#); Anderson et al. [2011](#); Smith [2011](#);
501 Ellis [2015](#)). Environmental factors such as vegetation cover have been used as a predictor to
502 explain how people know and use plants, and it is shown to be positively correlated to the
503 degree of knowledge and use of woody trees (Lucena et al. [2012](#); Gonçalves et al. [2016](#)), since
504 knowledge and the selection criteria can be shaped depending on the supply of plants in the
505 environment (Gaoue et al. [2016](#); Cunha [2020](#)). On the other hand, cultural factors such as the
506 perception of organoleptic properties (odor and taste) reflect strategies which determine plant
507 choice for a given use (Johns [1998](#); Cámara-Leret et al. [2017](#); Dai et al. [2017](#)).

508 Thus, environmental and cultural factors may reflect a cost/benefit ratio in the behavior
509 of human groups (Ellis [2015](#)), especially in environments with certain environmental
510 restrictions, as is the case of Dry Forests which are marked by the strong spatial and temporal
511 irregularity of the rains. Faced with these conditions, human populations tend to develop a set
512 of adaptive strategies to ensure their coexistence and use of plants in these environments. The
513 strategies range from the choice of plants which have morphological attributes available
514 throughout the year, to the consumption of similar plants for a specific use (Brown et al. [2011](#);
515 Nascimento et al. [2016](#); Pedrosa et al. [2021](#)). These relationships provide human behavior

516 development which emphasizes strategies to maximize energy expenditure when acquiring
 517 the resource. The study of these phenomena proposes thinking about the assumptions which
 518 influence the strategies for selecting and using plants. Thus, this chapter serves as a general
 519 introduction to the main information which comprises understanding strategies for managing
 520 and using plants by local populations located in dry tropical forests in Brazil.



521
 522 **Fig. 1** Location map of Brazilian semiarid. Highlight the semiarid region with merely
 523 illustrative figures that represent different activities carried out by local populations. On the
 524 left side of figure, different types and uses of Caatinga vegetation. A, B, C, D, and E = images
 525 of vegetation, including areas of anthropized Caatinga. F = goats feeding, and fences built; G
 526 = coal production; H = production of technological artifacts. All uses from Caatinga species.
 527 (Photo: Humberto A. Almeida and Maiara B. Ramos)

528
 529
 530 **Human Populations in the Caatinga – Human-Environment Interactions: The Past and**
 531 **the Present**

532
 533 The semi-arid region of Brazil is inserted within the domain of Seasonally Dry Tropical
 534 Forests (SDTF), called Caatinga (Fig.1). The Caatinga is considered one of the largest and
 535 most populous semi-arid regions in the world, occupying about 912,500 km² of Brazilian

536 territory (Moro et al. [2016](#); Silva et al. [2018](#)). Its vegetation is strongly influenced by
537 environmental filters which determine the assembly of plant communities (Pennington et al.
538 [2018](#); Fernandes et al. [2020](#)). The plant species have a set of morphophysiological adaptations
539 to survive the conditions imposed by the low and irregular precipitation level which has the
540 capacity to limit the development and distribution of plant species (Souza [2020](#); Fernandes et
541 al. [2020](#)). Chronic anthropic disturbances linked to the precipitation gradient also have a strong
542 influence on the availability and richness of plants (Rito et al. [2017](#); Silva and Souza [2018](#)).
543 The Caatinga region historically had unrestrained extraction of its natural resources (Silva et
544 al. [2018](#)). The anthropization process suggests that the anthropic impacts have been massive
545 since the first colonial settlements (Souza and Souza [2016](#)). As the European conquest
546 advanced in Brazil's semi-arid region, land donations through *sesmaria* letters promoted the
547 settlement of these regions and the expansion of agricultural production (Melo [1998](#)).

548 The usage and occupation patterns of environments in the last four centuries in the
549 Brazilian semi-arid region have stimulated the reproduction of inadequate agricultural and
550 livestock techniques among the local populations for the semi-arid region, and which provide
551 strong environmental changes. Historical evidence shows that the majority of requests for
552 *sesmaria* letters were sent to farmers to allocate cattle in drier regions (Pereira [2008](#)). Cattle
553 migration increased fires to remove the tree vegetation to create space for grass growth which
554 served as potential forage (Pereira [2008](#)).

555 It is important to consider that the current species richness and distribution patterns are
556 not only a result of climatic and environmental factors such as the relief and soil
557 characteristics, but above all the result of these factors acting in conjunction with the influence
558 of human activities (Silva and Souza [2018](#); Antongiovanni et al. [2020](#); Ramos et al. [2020](#)).
559 Overgrazing and agriculture are currently considered the main agents causing biodiversity loss
560 in these regions (Sobrinho et al. [2016](#); Schulz et al. [2019](#); Antongiovanni et al. [2020](#)). An
561 aggravating factor for this scenario is that the synergistic effect between anthropic impacts
562 and low rainfall regimes negatively contributes to the resilience of these ecosystems (Rito et
563 al. [2017](#)). This happens through a delay in the establishment and recruitment of regenerating
564 species which synchronously develop with the rainy season, which leads to a decrease in
565 diversity (Marinho et al. [2016](#); Souza [2020](#)).

566 The environmental characteristics of the semi-arid region generate several
567 socioeconomic imbalances (Silva et al. [2018](#)). For example, the rain regime is considered a
568 problem for the economic development and survival of the human population which depends
569 on natural capital. As a consequence of this irregular rainfall, there is a shortage in the drinking

570 water supply to some places in the semi-arid region (Montenegro and Ragab 2012; Andrade
571 et al. 2017). The situation became even more aggravated by the failure to implement public
572 policies for different reasons. First, government programs were developed with a lack of
573 alignment between environmental and social issues (Silva et al. 2018). Several adopted
574 strategies failed due to a lack of articulation which considered environmental characteristics,
575 for example, the insertion of *Prosopis juliflora* (Sw). D.C. as an alternative forest resource for
576 fodder and wood production, but which further problematized the socio-environmental issues
577 of the Brazilian semi-arid region (Andrade et al. 2009). Second, most of these government
578 programs made it possible for large landowners to benefit, which has culminated in growing
579 social inequality in the region which has persisted for many years (Silva et al. 2018). In
580 addition, a lack of continuity in social programs which are compromised according to the
581 historical-political moment (Dantas et al. 2020). Some public policies have been implemented
582 to tackle the drought phenomenon and assist the most vulnerable populations, but it is not a
583 reality for the whole semi-arid and does not help to structure the population's economic
584 independence (Mattos and May 2020).

585

586 **The Influence of the Environment on Plant Selection**

587

588 The environmental context is one of the most tested predictor variables in ethnobiology;
589 and, among the aspects related to the environment, the availability of plants has remained as
590 one of the main hypotheses which guide research on the differential use of plants in socio-
591 ecological systems. The first evidence in ethnobiology which verified the influence of the
592 environment on knowledge emerged in the seminal study by Phillips and Gentry (1993), which
593 was influenced by ecological ideas and explained how the appearance (abundance) of plants
594 influenced herbivory (Fenny 1976). The idea formulated in 1993 was that plants which are
595 most used by human populations would be those most apparent in the environment. The
596 appearance hypothesis suggested by these authors confirmed inductive observations that more
597 abundant plants were more useful (Johns 1998). The hypothesis suggested by Phillips and
598 Gentry (1993) explained the use of quantitative techniques to test whether the use of plants by
599 local populations was related to availability.

600 Thus, the usage value model proposed in the appearance hypothesis has become an
601 important measure for characterizing plant use. The technique was also corroborated by
602 successive research (Lucena et al. 2012; Maldonado et al. 2013), and in most studies the
603 assumptions indicate that apparent species tend to be more used by local populations as they

604 are more likely to be experimented with, while less-apparent plants are less experimented with.
605 This relationship started to be used by researchers as an indicator of important plants for
606 people (Ribeiro et al. 2014), and it shows that the frequency with which people keep using it
607 is directly affected by the availability of plants (Gonçalves et al. 2016).

608 An evidence pattern in dry forest regions assumes that the ecological parameters,
609 relative dominance, and abundance of plants recorded in plant communities influence socio-
610 ecological systems (Lucena et al. 2012; Guerra et al. 2015; Ribeiro et al. 2014). Some studies
611 select two distinct areas as predictor variables to access the plant communities, characterizing
612 them as conserved areas (where there is no plant collection) and anthropic areas (plant
613 collection area) (Lucena et al. 2012; Ribeiro et al. 2014). However, measuring conserved
614 environments is a difficult task, since most environments have some level of anthropic impact
615 (Menezes and Baldauf 2021). Furthermore, as far as we know, studies reporting the use of
616 these samples do not declare the level of human impact in environments classified as
617 anthropic.

618 On the other hand, the use of spatial tools which can improve the validation of
619 anthropogenic disturbance levels in environments is increasing. The changes in the vegetation
620 cover have been used as a model to evaluate the anthropic impacts on the use of plants by rural
621 communities (Naah and Guuroh 2017). In a study with populations of sacred areas in Nepal,
622 Kunwar et al. (2020) used remote-sensing images to analyze how the lost vegetation cover
623 influenced the use of medicinal plants over the years. The authors realized that sacred practices
624 were changing, especially when they noticed that areas with declining vegetation cover made
625 it impossible to collect medicinal plants. (Kunwar et al. 2020). Perhaps, providing direct
626 information (i.e., biomass of domestic animals, collection of forest resources) and indirect
627 information (i.e., satellite images, history of past uses) about anthropic characteristics present
628 in ecological samples is an alternative to demonstrate how the ecological community
629 influences knowledge with greater precision.

630 Human disturbances in Brazil's seasonally dry forest are a continuing reality. Timber
631 extraction, collection of non-wood products, and livestock breeding are common practices
632 among local human populations and represent chronic anthropic disorders, punctual and
633 continuous disturbances in the environment (Martorell and Peters 2005; Antongiovanni et al.
634 2018), which lead to decreases in the taxonomic, functional, and phylogenetic diversity
635 patterns of plant communities (Coelho et al. 2020; Rito et al. 2017; Zorger et al. 2019), and
636 can cause direct and indirect impacts to local human populations who depend on forests (Díaz
637 et al. 2011).

638 Human-induced environmental changes have exposed diverse environments to high
639 levels of degradation due to intense human activities (Leal et al. 2018; Antongiovanni et al.
640 2020). The amount of fragmented landscapes without functional connectivity has had
641 profound impacts on the biodiversity of dry forests (Calzada et al. 2018; Antongiovanni et al.
642 2018). The criteria for plant selection occur in detriment for several reasons. Human
643 populations have developed several strategies when faced with changes in the landscape.
644 Evidence of the influence of plant availability on local knowledge has served as a model to
645 explain human behavior (Clement et al. 2015; Cámara-Leret et al. 2017). This assumption
646 made the scientific community investigate how human behavior can be shaped according to
647 environmental fluctuations (Díaz et al. 2011, 2018). Evidence suggests that the greater the
648 heterogeneity of the environment in terms of plant availability leads to people demonstrating
649 more specialized plant use behavior, since the environment enables a greater supply of plants
650 to choose from. In contrast, people tend to exhibit generalist behavior in environments with
651 less plant availability due to biological homogenization, as they concentrate their collections
652 for available species (Albuquerque et al. 2013).

653

654 **Human Populations in the Caatinga and Its Dynamics Over Time**

655

656 It is necessary to consider that people use different strategies throughout the year to
657 ensure safety in the use of a resource due to the climatic seasonality of the Brazilian semi-arid
658 region (Albuquerque 2006; Cunha 2020). These strategies were verified and tested through
659 some hypotheses in ethnobiology. The hypothesis of climatic seasonality strongly
660 demonstrates how the knowledge of medicinal plants is shaped by the Caatinga vegetation
661 seasonality (Albuquerque 2006). Studies show that native plants which are available
662 throughout the year are more widely used by local populations. On the other hand, herbaceous
663 plants are less chosen for use because of their availability and because they are ephemeral and
664 in line with precipitation periods (Albuquerque 2006; Naah 2018).

665 Some research has pointed out that human populations are able to recognize and select
666 groups of plants with similar characteristics, depending on the purpose of their use (Brown et
667 al. 2011; Nunes et al. 2016; Cámara-Leret et al. 2017; Pedrosa et al. 2021). Species of interest
668 for wood use can be selected for their higher wood density, a fact which is reflected in greater
669 energy potential (Lucena et al. 2007). In addition to the wood density, the efficiency offered
670 by the plant is a positive criterion in selecting a species for combustible uses (Hora et al. 2021).
671 Likewise, Ramos et al. (2015) demonstrated that plants used for the production of

672 technological artifacts had an undetermined replacement time. This may indicate a criterion
673 in selecting plants with a certain functional characteristic to meet a specific use. In addition,
674 that local populations perceive that domestic animals tend to forage groups of plants with
675 greater palatability (Niu et al. [2016](#); Cunha [2020](#)). This taxonomic selectivity enables us to
676 observe that local populations select plants with specific functional characteristics (Cámara-
677 Leret et al. [2017](#); Cunha [2020](#)). Non-random plant selection is one of the most tested ideas in
678 ethnobotany (Moerman [1979](#), [1991](#); Medeiros et al. [2015](#); Ford and Gaoue [2017](#); Muleba et
679 al. [2021](#)), and it has contributed to pertinent answers on the usage pattern of plants. For
680 example, the literature has shown that the chemical component is an attribute which can help
681 to justify selecting medicinal plants and is perhaps due to the bioactive characteristics shared
682 among themselves (Johns [1998](#); Yessoufou et al. [2015](#); Cámara-Leret et al. [2017](#); Ford and
683 Gaoue [2017](#)). This can refer to Moerman's explanations where he explains that species of the
684 same family share characteristics due to their evolutionary kinship, which may in turn
685 influence their physiology (Moerman [1991](#); Pedrosa et al. [2021](#)). Selecting plants for use may
686 be due to selecting a set of specific functional characteristics which present a phylogenetic
687 convergence of species with specific traits (Brown et al. [2011](#); Pedrosa et al. [2021](#)). This
688 relationship has caused the scientific community to use phylogenetic inferences to understand
689 the evolutionary history of plants which are useful for local populations (Cámara-Leret et al.
690 [2017](#); Garnatje et al. [2017](#)).

691 Predictions coined in evolutionary history are emerging as a tool to test whether the
692 taxonomic signal in the use of plants translates into phylogenetic selectivity (Grace et al.
693 [2015](#)). Phylogenetic methods enable exploring usage patterns of plants which are similar to
694 each other (Lei et al. [2020](#)). Furthermore, the use of phylogenetic hypotheses in ethnobiology
695 can overcome the limitation on investigations regarding plant selection. Despite little use until
696 now, the use of phylogenies is relatively new in ethnobiology. However, it has recently gained
697 momentum in the agreement that plant selection is not random (Dáaz et al. 2007; Brown et al.
698 [2011](#); Cámara-Leret et al. [2017](#); Dai et al. [2017](#); Leso et al. [2017](#); Naah and Guuroh [2017](#);
699 Guzman and Molina [2018](#); Lei et al. [2018](#); Cunha [2020](#); Lei et al. [2020](#); Hora et al. [2021](#);
700 Molina-Venegas et al. [2021](#); Pedrosa et al. [2021](#)).

701

702 **Final Considerations**

703

704 In summary, in this chapter we bring a general approach highlighting evidence of
705 environmental factors that modulate the use and selection of plant species in the Caatinga.

706 This approach highlights the importance of ecological and evolutionary integration to social
707 and cultural systems in order to understand how the dynamics of use and selection of plants
708 occurs in local human communities. We believe that by integrating different views, a better
709 understanding of the dynamics of ecosystems is possible, which is essential to ensure the
710 current and future provision of ecosystem services for human populations, especially in a
711 scenario of global change.

712

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714

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720

721 **References**

722

723 Albuquerque UP. Re-examining hypotheses concerning the use and knowledge of medicinal
724 plants: a study in the caatinga vegetation of NE Brazil. *J Ethnobiol Ethnomed.* 2006; 2:1–10.
725 <https://doi.org/10.1186/1746-4269-2-30>.

726 Albuquerque UP, Andrade LHC. Uso de plantas em uma comunidade rural no semi-árido do
727 estado de Pernambuco, município de Alagoinha (Nordeste do Brasil). *Interciencia.* 2002;27:
728 336–46.

729 Albuquerque UP, Soldati GT, Ramos MA, De Melo JG, De Medeiros PM, Nascimento ALB,
730 et al. The influence of the environment on natural resource use: Evidence of apparency.
731 *Evolutionary Ethnobiology.* 2015;131–147. https://doi.org/10.1007/978-3-319-19917-7_10

732 Albuquerque UP, Ferreira-Júnior WS. What do we study in evolutionary ethnobiology?
733 Defining the theoretical basis for a research program. *Evol Biol.* 2017;44:206–15.
734 <https://doi.org/10.1007/s11692-016-9398-z>.

735 Albuquerque UP, Gonçalves PHS, Ferreira-Júnior WS, Chaves LS, Oliveira RCS, Silva TLL,
736 Santos GC, Araújo EL. Humans as niche constructors: revisiting the concept of chronic
737 anthropogenic disturbances in ecology. *Perspect Ecol Conserv.* 2018;16:1–11.
738 <https://doi.org/10.1016/j.pecon.2019.11.004>.

739 Albuquerque UP, de Medeiros PM, Ferreira Júnior WS, Silva TC, Silva RRV, Gonçalves-Souza
740 T. Social-ecological theory of maximization: basic concepts and two initial models. *Biol*
741 *Theory*. 2019. <https://doi.org/10.1007/s13752-019-00316-8>.

742 Albuquerque UP, Nascimento ALB, Chaves LS, Feitosa IS, Moura JMB, Gonçalves PHS, Silva
743 RH, Silva TC, Ferreira-Júnior WS. The chemical ecology approach to modern and early human
744 use of medicinal plants. *Chemoecology*. 2020. <https://doi.org/10.1007/s00049-020-00302-8>.

745 Alvalá RCS, Cunha APMA, Brito SSB, Seluchi ME, Marengo JA, Moraes OLL, Carvalho MA.
746 Drought monitoring in the Brazilian Semiarid region. *An Acad Bras Cienc*. 2019.
747 <https://doi.org/10.1590/0001-3765201720170209>.

748 Anderson EN, Pearsall D, Hunn E, Turner N. *Ethnobiology*. *Ethnobiology*. 2011;6:3–22.
749 <https://doi.org/10.1002/9781118015872>

750 Antongiovanni M, Venticinque EM, Matsumoto M, Fonseca CR. Chronic anthropogenic
751 disturbance on Caatinga dry forest fragments. Biswas S, editor. *Journal of Applied Ecology*.
752 2020; 1365–2664.13686. <https://doi.org/10.1111/1365-2664.13686>

753 Andrade LA, Fabricante JR, Oliveira FX. Invasão biológica por *Prosopis juliflora* (Sw.) DC.:
754 Impactos sobre a diversidade e a estrutura do component arbustivo-arbóreo da caatinga no
755 estado do Rio Grande do Norte, Brasil. *Acta Bot Bras*. 2009;23:935–43.
756 <https://doi.org/10.1590/S0102-33062009000400004>.

757 Antongiovanni M, Venticinque EM, Fonseca CR. Fragmentation patterns of the Caatinga
758 drylands. *Landsc Ecol*. 2018;33:1353–76. <https://doi.org/10.1007/s10980-018-0672-6>.

759 Born J, Pluess AR, Burslem DFRP, Nilus R, Maycock CR, Ghazoul J. Differing life history
760 characteristics support coexistence of tree soil generalist and specialist species in tropical rain
761 forests. *Biotropica*. 2014; 46:58–68. <https://doi.org/10.1111/btp.12083>.

762 Brown KA, Flynn DFB, Abram NK, Ingram JC, Johnson SE, Wright P. Assessing natural
763 resource use by forest-reliant communities in Madagascar using functional diversity and
764 functional redundancy metrics. *PLoS One*. 2011.
765 <https://doi.org/10.1371/journal.pone.0024107>.

766 Cáceres DM, Tapella E, Quétier F, Díaz S. The social value of biodiversity and ecosystem
767 services from the perspectives of different social actors. *Ecology and Society*. 2015;20.
768 <https://doi.org/10.5751/ES-07297-200162>.

769 Calzada L, Meave JA, Bonfil C, Figueroa F. Lands at risk: land use/land cover change in two
770 contrasting tropical dry regions of Mexico. *Appl Geogr*. 2018;99:22–30.
771 <https://doi.org/10.3390/land9100364>.

772 Cámara-Leret R, Faurby S, Macía MJ, Balslev H, Gödel B, Svenning JC, Kissling WD,
773 Rønsted N, Saslis-Lagoudakis CH. Fundamental species traits explain provisioning services of
774 tropical american palms. *Nat Plants*. 2017;3:1–7. <https://doi.org/10.1038/nplants.2016.220>.

775 Cardoso MB, Ladio A, Dutrus SM, Lozada M. Preference and calorific value of fuelwood
776 species in rural populations in northwestern Patagonia. *Biomass Bioenergy*. 2015;81:514–20.
777 <https://doi.org/10.1016/j.biombioe.2015.08.003>.

778 Clement CR, Denevan WN, Heckenberger MJ, Junqueira B, Neves EG, Teixeira WG. Woods
779 the domestication of Amazonia before European conquest. *Proc R Soc B*. 2015.
780 <https://doi.org/10.1098/rspb.2015.0913>.

781 Coelho AJP, Magnago IFS, Matos FAR, Mota NM, Diniz ES, Meira-Neto JAA. Effects of
782 anthropogenic disturbances on biodiversity and biomass stock of cerrado, the brazilian savanna.
783 *Biodivers Conserv*. 2020. <https://doi.org/10.1007/s10531-020-02013-6>.

784 Cunha SS. Recursos forrageiros em Florestas Secas: como o conhecimento ecológico local pode
785 influenciar a dinâmica de uso frente a sazonalidade. Mestrado na pós-graduação em Ecologia e
786 Conservação (Dissertação). Universidade Estadual da Paraíba, Campus Campina Grande-PB.
787 2020.

788 Silva JMC, Barbosa LCF, Leal IR, Tabarelli M. The Caatinga: Understanding the challenges.
789 *Caatinga: The Largest Tropical Dry Forest Region in South America*. Springer International
790 Publishing; 2018. pp. 3–19. https://doi.org/10.1007/978-3-319-68339-3_1

791 Dai X, Zhang W, Xu J, Duffy KJ, Guo Q. Global pattern of plant utilization across different
792 organisms: does plant apparency or plant phylogeny matter? *Ecol Evol*. 2017;7:2535–45.
793 <https://doi.org/10.1002/ece3.2882>.

794 Dantas JC, Silva RM, Santos CAG. Drought impacts, social organization, and public policies
795 in northeastern Brazil: a case study of the upper Paraíba river basin. *Environ Monit Assess*.
796 2020. <https://doi-org.ez19.periodicos.capes.gov.br/10.1007/s10661-0208219-0>.

797 De Almeida CFCBR, De Lima E Silva TC, De Amorim ELC, Maia MBDS, De Albuquerque
798 UP. Life strategy and chemical composition as predictors of the selection of medicinal plants
799 from the caatinga (Northeast Brazil). *Journal of Arid Environments*. 2005;62:127–142.
800 <https://doi.org/10.1016/j.jaridenv.2004.09.020>.

801 de Andrade EM, Aquino D do N, Chaves LCG, Lopes FB. Water as Capital and Its Uses in the
802 Caatinga. *Caatinga: The Largest Tropical Dry Forest Region in South America*. 2017;281–302.
803 https://doi.org/10.1007/978-3-319-68339-3_10.

804 Díaz S, Quétier F, Cáceres DM, Trainor SF, Pérez-Harguindeguy N, Bret-Harte MS, et al.
805 Linking functional diversity and social actor strategies in a framework for interdisciplinary

806 analysis of nature's benefits to society. Proceedings of the National Academy of Sciences of
807 the United States of America. 2011;108: 895–902. <https://doi.org/10.1073/pnas.1017993108>

808 Díaz S, Pascual U, Stenseke M, Martín-López B, Watson RT, Molnár Z, et al. Assessing
809 nature's contributions to people: Recognizing culture, and diverse sources of knowledge, can
810 improve assessments. Science. 2018;359:270–272. <https://doi.org/10.1126/science.aap8826>

811 Ellis EC. Ecology in an anthropogenic biosphere. Ecol Monogr. 2015;85:287–331.
812 <https://doi.org/10.1890/14-2274.1>.

813 Ellis EC, Ramankutty N. Putting people in the map: Anthropogenic biomes of the world.
814 Frontiers in Ecology and the Environment. Ecological Society of America; 2008. pp. 439–447.
815 <https://doi.org/10.1890/070062>

816 Fenny PP. Plant apparency and chemical defense. In: Wallace JW, Mansell RL, editors. Recent
817 advances in phytochemistry. New York: Plenum Press; 1976. p. 1–40.

818 Fernandes MF, Cardoso D, de Queiroz LP. An updated plant checklist of the Brazilian Caatinga
819 seasonally dry forests and woodlands reveals high species richness and endemism. Journal of
820 Arid Environments. 2020; 174:104079. <https://doi.org/10.1016/j.jaridenv.2019.104079>.

821 Ferreira-Júnior WS, Medeiros PM, Albuquerque UP. Evolutionary ethnobiology. Encyclopedia
822 of life sciences (online). 2019. <https://doi.org/10.1002/9780470015902.a0028232>.

823 Folke C, Polasky S, Rockström J, Galaz V, Westley F, Lamont M, et al. Our future in the
824 Anthropocene biosphere. Ambio. 2021. <https://doi.org/10.1007/s13280-021-01544-8>

825 Ford J, Gaoue OG. Alkaloid-poor plant families, Poaceae and Cyperaceae, are overutilized for
826 medicine in Hawaiian pharmacopoeia. Econ Bot. 2017;71:123–32.

827 Gaoue OG, Coe MA, Bond M, Hart G, Seyler BC, McMillen H. Theories and major hypotheses
828 in ethnobotany. Econ Bot. 2017;71:269–87.

829 Garnatje T, Peñuelas J, Vallès J. Reaffirming 'ethnobotanical convergence'. Trends Plant Sci.
830 2017. <https://doi.org/10.1016/j.tplants.2017.06.001>.

831 Giraldi M, Hanazaki N. Uso e conhecimento tradicional de plantas medicinais no Sertão do
832 Ribeirão, Florianópolis/SC, Brasil. Acta Bot Bras. 2010;24:395–406.

833 Gonçalves PHS, Albuquerque UP, Medeiros PM. The most commonly available woody plant
834 species are the most useful for human populations: a meta-analysis. Ecol Appl. 2016;26:2238–
835 53.

836 Grace OM, Buerki S, Symonds MR Forest F, Wyk AV, Smith GF, Klopper RR, BJORÅ NS,
837 Demissew S, Simmonds MSJ, Rønsted N. Evolutionary history and leaf succulence as
838 explanations for medicinal use in aloes and the global popularity of *Aloe vera*. Evol Biol. 2015.
839 <https://doi.org/10.1186/s12862-015-0291-7>.

840 Gras A, Hidalgo, O, D'Ambrosio U, Parada M, Garnatje T, Vallés J. The Role of Botanical
841 Families in Medicinal Ethnobotany: A Phylogenetic Perspective. *Plants* 2021, Vol. 10, Page
842 163 10. Multidisciplinary Digital Publishing Institute: 163.
843 <https://doi.org/10.3390/plants10010163>.

844 Guerra NM, Carvalho TKN, Ribeiro JES, Ribeiro JPO, Barbosa AR, Lima JRF, Alves CAB,
845 Oliveira RS, Lucena RFP. Ecological apparency hypothesis and plant utility in the semiarid
846 region of Brazil. *Ethnobot Res Appl.* 2015;14:423–35.

847 Guzman E, Molina J. The predictive utility of the plant phylogeny in identifying sources of
848 cardiovascular drugs. *Pharm Biol.* 2018;1:154–64.
849 <https://doi.org/10.1080/13880209.2018.1444642>

850 Hora JSL, Feitosa IS, Albuquerque UP, Ramos MA, Medeiros PM. Drivers of species' use for
851 fuelwood purposes: a case study in the Brazilian semiarid region. *J Arid Environ.*
852 2021;185:104–324.

853 Johns T. *With bitter herbs they shall eat it: chemical ecology and the origins of human diet and*
854 *medicine.* Tucson: University of Arizona Press; 1990.

855 Johns. *The Origins of Human Diet and Medicine: Chemical Ecology.* *The Quarterly Review of*
856 *Biology.* 1998;73:118–118. <https://doi.org/10.1086/420162>

857 Kraft NJB, Adler PB, Godoy O, James EC, Fuller S, Levine JM. Community assembly,
858 coexistence and the environmental filtering metaphor. *Funct Ecol.* 2015;29:592–9.

859 Kunwar RM, Evans A, Mainali J, Ansari AS, Rimal B, Bussmann RW. Change in forest and
860 vegetation cover influencing distribution and uses of plants in the Kailash Sacred Landscape,
861 Nepal. *Environ Dev Sustain.* 2020;22:1397–412.

862 Ladio AH, Albuquerque UP. The concept of hybridization and its contribution to urban
863 ethnobiology. *Ethnobiol Conserv.* 2014. <https://doi.org/10.15451/ec2014-11-3.6-1-9>.

864 Laurance WF, Sayer J, Cassman KG. Agricultural expansion and its impacts on tropical nature.
865 *Trends Ecol Evol.* 2014;29:107–16.

866 Leal IR, Lopes AV, Machado IC, Tabarelli M. Plant-animal interactions in the Caatinga: current
867 overview and future perspectives. In: Silva JMC, Tabarelli M, Leal IR, editors. *Biodiversity,*
868 *ecosystem services and sustainable development in Caatinga: the largest tropical dry forest*
869 *region in South America.* Cham: Springer; 2018. p. 255–78.

870 Lei D, Wu J, Leon C, Huang L, Hawkins JA. Medicinal plants of Chinese Pharmacopoeia and
871 Daodi: insights from phylogeny and biogeography. *Chin Herb Med.* 2018;10:268–78.

872 Lei D, Thuraiya AJ, Teixidor-Toneu I, Saslis-Lagoudakis CH, Ghazanfar SA, Hawkins JA.
873 Comparative analysis of four medicinal floras: phylogenetic methods to identify cross- cultural
874 patterns. *Plants People Planet*. 2020;00:1–13. <https://doi.org/10.1002/ppp3.10131>.

875 Leso LK, Elansary HO, Mearns K, Yessoufou K. Ethnobotany at a local scale: diversity of
876 knowledge of medicinal plants and assessment of plant cultural importance in the Polokwane
877 local municipality, South Africa. *Bot Lett*. 2017.
878 <https://doi.org/10.1080/23818107.2016.1268064>.

879 Lins-Neto EMF, Peroni N, Casas A, Parra F, Aguirre X, Guillén S, Albuquerque UP. Brazilian
880 and Mexican experiences in the study of incipiente domestication. *J Ethnobiol Ethnomed*.
881 2014;10:33.

882 Lucena RFP, Medeiros PM, Araújo EF, Alves AGC, Albuquerque UP. The ecological
883 apparency hypothesis and the importance of useful plants in rural communities from
884 Northeastern Brazil: an assessment based on use value. *J Environ Manag*. 2012;96:106–115.
885 <https://doi.org/10.1016/J.JENVMAN.2011.09.001>.

886 Maldonado B, Caballero J, Delgado-Salinas A, Lira R. Relationship between use value and
887 ecological importance of floristic resources of seasonally dry tropical forest in the Balsas river
888 basin, México. *Econ Bot*. 2013;67:17–29.

889 Marinho FP, Mazzochini GG, Manhães AP, Weisser WW, Ganade G. Effects of past and
890 present land use on vegetation cover and regeneration in a tropical dryland forest. *J Arid*
891 *Environ*. 2016;132:26–33. <https://doi.org/10.1016/j.jaridenv.2016.04.006>

892 Martorell C, Peters E. The measurement of chronic disturbance and its effects on the threatened
893 cactus *Mammillaria pectinifera*. *Biol Conserv*. 2005;124:199–207.
894 <https://doi.org/10.1016/j.biocon.2005.01.025>

895 Mattos CL, May P. Two equivalent droughts with different social impacts in the Brazilian
896 Semiarid Region. Edição especial – Sociedade e ambiente no Semiárido: controvérsias e
897 abordagens. 2020;55:28–53.

898 Medeiros PM, Ladio AH, Albuquerque IP. Local criteria for medicinal plant selection. In:
899 Albuquerque UP, Medeiros P, Casas A, editors. *Evolutive ethnobotany*. Springer; 2015.
900 <https://doi.org/10.1007/978-3-319-19917-7-11>.

901 Melo JC. Lei de Terras de 1850: Esbulhos e conflitos na transição para o capitalismo. Ariús
902 Campina Grande. 1998;8:56–64.

903 Menezes J, Baldauf C. Multiple perspectives on a biocultural environment: landscape
904 ethnoecology in the Brazilian dry forest. *J Arid Environ*. 2021.
905 <https://doi.org/10.1016/j.jaridenv.2020.104387>.

906 Moerman DE. Symbols and selectivity: a statistical analysis of native American medical
907 ethnobotany. *J Ethnopharmacol.* 1979. [https://doi.org/10.1016/0378-8741\(79\)90002-3](https://doi.org/10.1016/0378-8741(79)90002-3).

908 Moerman DE. The medicinal flora of native North America: an analysis. *J Ethnopharmacol.*
909 1991;31:1–42.

910 Molina-Venegas R, Fischer M, Mollel NP, Hemp A. Connecting plant evolutionary history and
911 human well-being at mt. Kilimanjaro, Tanzania. *Bot J Linn Soc.* 2020. 1–13.
912 <https://doi.org/10.1093/botlinnean/boaa049>.

913 Molina-Venegas R, Rodríguez MÁ, Pardo-de-Santayana M, Ronquillo C, Mabberley DJ.
914 Maximum levels of global phylogenetic diversity efficiently capture plant services for
915 humankind. *Nat Ecol Evol.* 2021. <https://doi.org/10.1038/s41559-021-01414-2>.

916 Montenegro S, Ragab R. Impact of possible climate and land use changes in the semi arid
917 regions: A case study from North Eastern Brazil. *Journal of Hydrology.* 2012;434–435:55–68.
918 <https://doi.org/10.1016/J.JHYDROL.2012.02.036>

919 Moro MF, Lughadha EN, Araújo FS, Martins FR. A phytogeographical metaanalysis of the 420
920 Semiarid Caatinga Domain in Brazil. *Bot Rev.* 2016;82:91–148

921 Muleba I, Yessoufou K, Rampedi IT. Testing the non-random hypothesis of medicinal plant
922 selection using the woody flora of the Mpumalanga Province, South Africa. *Environ Dev*
923 *Sustain.* 2021;23:4162–73.

924 Naah J-BSN. Investigating criteria for valuation of forage resources by local agro-pastoralists
925 in West Africa: using quantitative ethnoecological approach. *Journal of Ethnobiology and*
926 *Ethnomedicine* 2018;14:1–14. *BioMed Central:* 1–16. [https://doi.org/10.1186/S13002-018-](https://doi.org/10.1186/S13002-018-0261-4)
927 [0261-4](https://doi.org/10.1186/S13002-018-0261-4)

928 Naah JBSN, Guuroh RT. Factors influencing local ecological knowledge of forage resources:
929 Ethnobotanical evidence from West Africa’s savannas. *Journal of Environmental Management.*
930 2017;188:297–307. <https://doi.org/10.1016/j.jenvman.2016.11.064>

931 Nascimento ALB, Lozano A, Melo JG, Alves RRN, Albuquerque UP. Functional aspects of
932 the use of plants and animals in local medical systems and their implications for resilience. *J*
933 *Ethnopharmacol.* 2016;194:348–57. <https://doi.org/10.1016/j.jep.2016.08.017>

934 Niu K, He JS, Zhang S, Lechowicz MJ. Grazing increases functional richness but not functional
935 divergence in Tibetan alpine meadow plant communities. *Biodivers Conserv.* 2016;25:2441–
936 52. <https://doi.org/10.1007/s10531-015-0960-2>.

937 Nunes FA, Cabral D, Santos MV, Amorim ELC, Albuquerque UP. Plants used to feed
938 ruminants in semi-arid Brazil: a study of nutritional composition guided by local ecological
939 knowledge. *J Arid Environ.* 2016;135:96–103. <https://doi.org/10.1016/j.jaridenv.2016.08.015>

940 Odling-Smee J, Erwin DH, Palkovacs EP, Feldman MW, Laland KN. Niche construction
941 theory: a practical guide for ecologists. *Q Rev Biol.* 2013;88:4–28.

942 Pedrosa KM, Almeida HA, Ramos MB, Faria SL. Plants with similar characteristics drive their
943 use by local populations in the semi-arid region of Brazil. *Environ Dev Sustain.*
944 2021;23:16834–16847. <https://doi.org/10.1007/S10668-021-01355-7>

945 Pennington RT, Lehmann CER, Rowland LM. Tropical savannas and dry forests. *Curr Biol.*
946 2018;28:R541–5. <https://doi.org/10.1016/j.cub.2018.03.014>.

947 PubMed

948 Pereira DD. Cariris Paraibanos: do sesmarialismo aos assentamentos de reforma agrária. raízes
949 da desertificação? Tese de doutorado (Doutorado em Recursos Naturais). Universidade Federal
950 de Campina Grande, Paraíba. 2008. p. 373.

951 Periago ME, Tamburini DM, Ojeda RA, Cáceres DM, Díaz S. Combining ecological aspects
952 and local knowledge for the conservation of two native mammals in the Gran Chaco. *Journal*
953 *of Arid Environments.* 2017;147:54–62. <https://doi.org/10.1016/j.jaridenv.2017.07.017>

954 Phillips O, Gentry AH. The useful plants of Tambopata, Peru: I. Statistical hypotheses tests
955 with a new quantitative technique. *Econ Bot.* 1993;47:15–32.

956 Pinho BX, Tabarelli M, Engelbrecht BMJ, Sfair J, Melo FPL. Plant functional assembly is
957 mediated by rainfall and soil conditions in a seasonally dry tropical forest. *Basic Appl Ecol.*
958 2019;40:1–11. <https://doi.org/10.1016/j.baae.2019.08.0021439-1791>.

959 Ramos MB, Diniz FC, de Almeida HA, de Almeida GR, Pinto AS, Meave JA, et al. The role
960 of edaphic factors on plant species richness and diversity along altitudinal gradients in the
961 Brazilian semi-arid region. *Journal of Tropical Ecology.* 2020;36:199–212.
962 <https://doi.org/10.1017/S0266467420000115>

963 Ramos MA, Lucena RFP, Albuquerque UP. What drives the knowledge and local uses of timber
964 resources in human-altered landscapes in the semiarid region of northeast Brazil? *Int J Sustain*
965 *Dev Ecol.* 2015. <https://doi.org/10.1080/13504509.2015.1091796>.

966 Ribeiro JES, Carvalho TKN, Ribeiro JPO, Guerra NM, Pedrosa KM, Alves CAB, Sousa-Júnior
967 SP, Souto JS, Nunes AT, Lima JRF, Oliveira RS, Lucena RFP. Ecological apparency
968 hypothesis and availability of useful plants: testing different use values. *Ethnobot Res Appl.*
969 2014;12:425–32.

970 Ribeiro EMS, Arroyo-Rodríguez V, Santos BA, Tabarelli M, Leal IR. Chronic anthropogenic
971 disturbance drives the biological impoverishment of the Brazilian Caatinga vegetation. *J Appl*
972 *Ecol.* 2015;52:611–20. <https://doi.org/10.1111/1365-2664.12420>

973 Rito KF, Arroyo-Rodriguez V, Queiroz RT, Leal IR, Tabarelli M. Precipitation mediates the
974 effect of human disturbance on the Brazilian Caatinga vegetation. *J Ecol.* 2017;105:828–38.

975 Santoro FR, Ferreira-Júnior WS, Araújo TAS, Ladio AH, Albuquerque UP. Does plant species
976 richness guarantee the resilience of local medical systems? A perspective from utilitarian
977 redundancy. *PLoS One.* 2015;10:e0119826.

978 Santoro FR, Nascimento ALB, Soldati GT, Ferreira-Junior WS, Albuquerque UP. Evolutionary
979 ethnobiology and cultural evolution: opportunities for research and dialog. *J Ethnobiol*
980 *Ethnomed.* 2018;14:1–14.

981 Saslis-Lagoudakis CH, Savolainenb V, Williamsond EM, Forestc F, Wagstaffe SJ, Baralf SR,
982 Watsong MF, Pendryg CA, Hawkinsa A. Phylogenies reveal predictive power of traditional
983 medicine in bioprospecting. *Proc Natl Acad Sci U S A.* 2012;109:15835–40.

984 Saslis-Lagoudakis CH, Klitgaard BB, Floresta F, Francis L, Savolainen V, Williamson EM,
985 Hawkins JA. The use of phylogeny to interpret cross-cultural patterns in plant use and guide
986 medicinal plant discovery: an example from *Pterocarpus* (Leguminosae). *PLoS One.* 2017.
987 <https://doi.org/10.1371/journal.pone.0022275>.

988 Savo V, Joy R, Caneva G, McClatchey C. Plant selection for ethnobotanical uses on the Amalfi
989 Coast (Southern Italy). *J Ethnobiol Ethnomed.* 2015. [https://doi.org/10.1186/s13002-015-0038-](https://doi.org/10.1186/s13002-015-0038-y)
990 [y](https://doi.org/10.1186/s13002-015-0038-y).

991 Schulz K, Guschal M, Kowarik I, Silva de Almeida-Cortez J, Valadares de Sá Barreto Sampaio
992 E, Cierjacks A. Grazing reduces plant species diversity of Caatinga dry forests in northeastern
993 Brazil. *Applied Vegetation Science.* 2019;22:348–359. <https://doi.org/10.1111/avsc.12434>

994 Sfair JC, De Bello F, De Frana TQ, et al. Chronic human disturbance affects plant trait
995 distribution in a seasonally dry tropical forest. *Environ Res Lett.* 2018;13:025005.
996 <https://doi.org/10.1088/1748-9326/aa9f5e>.

997 Silva AC, Souza AF. Aridity drives plant biogeographical sub regions in the Caatinga, the
998 largest tropical dry forest and woodland block in South America. *PLoS ONE.* 2018;13.
999 <https://doi.org/10.1371/journal.pone.0196130>

1000 Singh SP. Chronic disturbance, a principal cause of environmental degradation in developing
1001 countries. *Environ Conserv.* 1998;25:1–2.

1002 Smith BD. A Cultural Niche Construction Theory of Initial Domestication. *Biological Theory.*
1003 2011;6:260–271. <https://doi.org/10.1007/s13752-012-0028-4>

1004 Sobrinho MS, Tabarelli M, Machado IC, et al. Land use, fallow period and the recovery of a
1005 Caatinga forest. *Biotropica.* 2016;48:586–97. <https://doi.org/10.1111/btp.12334>.

1006 Sousa R, Feng X, Antonino A, Montenegro S, Souza E, Porporato A. Vegetation response to
1007 rainfall seasonality and interannual variability in tropical dry forests. *Hydrol Process.*
1008 2016;30:3583–359. <https://doi.org/10.1002/hyp.10953>

1009 Souza SM. Cobertura e regeneração vegetal como preditores do processo de desertificação na
1010 Caatinga. Mestrado na pós-graduação em Ecologia e Conservação (Dissertação). Universidade
1011 Estadual da Paraíba, Campus Campina Grande PB. 2020.

1012 Souza BI, Souza RR. The occupation process of the Cariri Velhos – PB and the effects in the
1013 vegetal cover: contribution to the cultural biogeography of the semiarid. *Caderno Geogr.*
1014 2016;26:229. <https://doi.org/10.5752/p.2318-2962.2016v26>.

1015 Sterelny K. Social intelligence, human intelligence and niche construction. *Philosophical*
1016 *Transactions of the Royal Society B: Biological Sciences.* 2007;362:719–730.
1017 <https://doi.org/10.1098/RSTB.2006.2006>

1018 Teixidor-Toneu TI, Jordan FM, Hawkins JA. Comparative phylogenetic methods and the
1019 cultural evolution of medicinal plant use. *Nat Plants.* 2018;4:754–61.
1020 <https://doi.org/10.1038/s41477-018-0226-6>

1021 Toneu IT, Jordan FM, Hawkins JA. Comparative phylogenetic methods and the cultural
1022 evolution of medicinal plant use. *Nat Plants.* 2018;4:754–61.

1023 Wolverton S. *Ethnobiology 5: interdisciplinarity in an era of rapid environmental change.*
1024 *Ethnobiol Lett.* 2013;4:21–5.

1025 Yessoufou K, Daru BH, Muasya AM. Phylogenetic exploration of commonly used medicinal
1026 plants in South Africa. *Mol Ecol Resour.* 2015. <https://doi.org/10.1111/1755-0998-12310>.

1027 Zorger BB, Tabarelli M, Queiroz RT, Rosado BHP, Pinho BX. Functional organization of
1028 woody plant assemblages along precipitation and human disturbance gradients in a seasonally
1029 dry tropical forest. *Biotropica.* 2019; 00:1–13. <https://doi.org/10.1111/btp.12721>.

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1040 **4. CAPÍTULO III: Functional traits influencing selection and use of plant species by**
 1041 **local human populations in the tropical dry Caatinga forest**

1042 (Manuscrito submetido ao periódico Economic Botany)

1043

History for Manuscript Number: ECBO-D-23-00040
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1058 **Functional traits drive the selection of plant species used by agropastoralists in the**
1059 **Brazilian semiarid region**

1060

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1071

1072 **Abstract**

1073 Biodiversity plays a key role in the well-being and sustenance of human life. Rural
1074 communities directly dependent on vegetation resources will likely be the most affected by
1075 the losses of goods and services provided by biodiversity. Here, we tested the hypothesis that
1076 plant functional traits drive the selection and use of plant species by agropastoralists in the
1077 Brazilian semiarid region. We approached agropastoralists from rural communities across a
1078 Caatinga area to gather information about local ecological knowledge. This information was
1079 contrasted with the functional trait attributes of plants used for fuel, construction and forage,
1080 which are the three most common categories of plant use in the study area. Our results
1081 confirm that agropastoralists select the plants they use in each use category according to
1082 functional trait attributes that they associate with desirable goods and services. For example,
1083 wood density emerged as a relevant plant functional trait for plant selection in the fuel and
1084 construction categories due to its relationship with durability. Our approach allowed us to
1085 identify the fundamental plant functional traits driving the selection of plants used by local

1086 communities in the Caatinga.

1087

1088 *Key words:* ecosystem provision services, selective use, specific leaf area, wood density, plant

1089 height, stem diameter, ethnobotany

1090

1091 **1. Introduction**

1092 The quality of life and well-being of the global human population largely depends on the

1093 timely and adequate provision of a variety of ecosystem goods and services by nature (Díaz et

1094 al. 2006; Cáceres et al. 2015). Ecosystem services are the benefits that people obtain from

1095 ecosystems, and are related to multiple biodiversity components (Díaz et al. 2006; Brauman et

1096 al. 2007). For example, vegetation provides a broad range of key ecosystem services (Díaz et

1097 al. 2007; Hassen et al. 2023), including food, fibers, fuel, biochemicals, medicines, regulation

1098 of climate and erosion, plant pollination, spiritual and religious values, recreation and

1099 ecotourism, cultural heritage values, photosynthesis and carbon uptake from the atmosphere,

1100 biomass, and nutrient cycling (Millennium Ecosystem Assessment 2005; Díaz et al. 2011).

1101 The humans have learned to use the diversity of organisms in a unique way, with an

1102 extraordinary capacity to alter ecological and evolutionary processes across the ecosystems

1103 (Ellis and Ramankutty 2008; Smith 2011; Ellis 2015), which today represents a threat to the

1104 persistence of the human species itself.

1105 The theoretical framework of functional ecology helps understand species' responses

1106 to ecological interactions and environmental modifications due to global change, as well as

1107 the interplay between biodiversity and ecosystem properties (Diaz and Cabido 1997; Albert et

1108 al. 2010). Alterations in the provision of ecosystem services have direct impacts on the quality

1109 of life of humans and compromises our survival; therefore, it is necessary to assess the

1110 perceptions of the different social actors about ecosystem services and the components of
1111 functional diversity (Díaz and Cabido 2001).

1112 The numerous plant species occurring in every region of the planet differ greatly in their
1113 functional trait attributes (de Bello et al. 2013; Luo et al. 2016). Such large functional diversity
1114 may play an important role in the selection process of species by local populations; at each time
1115 and place, the choice of sets of useful species, on the one hand, and undesirable species, on the
1116 other, may be guided depending on the type of use that people want to make of plants, and
1117 weighted by the abundance of the different species in the environment and their attractive or
1118 undesirable characteristics (Albuquerque and Oliveira 2007; Meuret and Provenza 2015).

1119 Previous studies have examined the relationships between functional traits and plant
1120 selection. For example, when focusing on food, people mainly select plants with large fruits
1121 (Cámara-Leret et al. 2017), whereas when it comes to feeding their herds they rather value
1122 plants with high nutritional contents (Meuret and Provenza 2015; Nunes et al. 2016; Cunha et
1123 al. 2022). Management decisions made by agropastoralists are often based on plant
1124 characteristics, for example their phenological patterns, plant dimensions (e.g., height) and its
1125 casted shade, leaf density, and rooting patterns (Martin and Isaac 2015; Valencia et al. 2015).
1126 Similarly, there is evidence suggesting that people value woody plants with high wood density
1127 for fuel or construction (Ticktin 2004; Brown et al. 2011; Gaoue et al. 2016; Gonçalves et al.
1128 2021a). Therefore, a functional trait-based approach can shed new light on the relationship
1129 between the functional trait attributes of useful plants and the fulfillment of different human
1130 needs, and to the dynamics of the provision and maintenance of ecosystem services (Brown et
1131 al. 2011; Díaz et al. 2011; Cámara-Leret et al. 2017; Kusumoto et al., 2020).

1132 The observation, use, and management of biodiversity and the physical environment
1133 over millennia have led human societies to develop an enormous body of knowledge,
1134 practices, and beliefs, which is known as Local Ecological Knowledge (LEK) (Berkes et al.

1135 2000). Therefore, the concept of LEK creates an important window of understanding about
1136 how human societies relate to biodiversity (Díaz et al. 2011; Lopes 2017). Connecting our
1137 knowledge on biodiversity and the provision of ecosystem services with societal interests is
1138 essential for the maintenance and sustainability of socio-ecological systems (Berkes et al.
1139 2000; Folke 2006; Saynes-Vásquez et al. 2013; Pascual et al. 2017; Díaz et al. 2018).

1140 The management of plants for various purposes such as subsistence agriculture and
1141 raising goats, sheep, and cattle reveals essential aspects of the LEK. This is true for the local
1142 populations of the Caatinga, which is the largest tropical semi-arid region in the New World
1143 and has a long history of colonization, changes in land use and interactions between humans
1144 and plants (Melo 2017; Antongiovanni et al. 2020). Despite previous efforts to investigate
1145 LEK of rural populations in Brazil's semiarid region (Farias et al. 2007; Ramos et al. 2008;
1146 Lucena et al. 2012; Gonçalves et al. 2016; Gonçalves et al. 2021b), important gaps persist
1147 regarding the selection criteria used by these populations, especially regarding the role of
1148 plant functional trait attributes (Cunha et al. 2022).

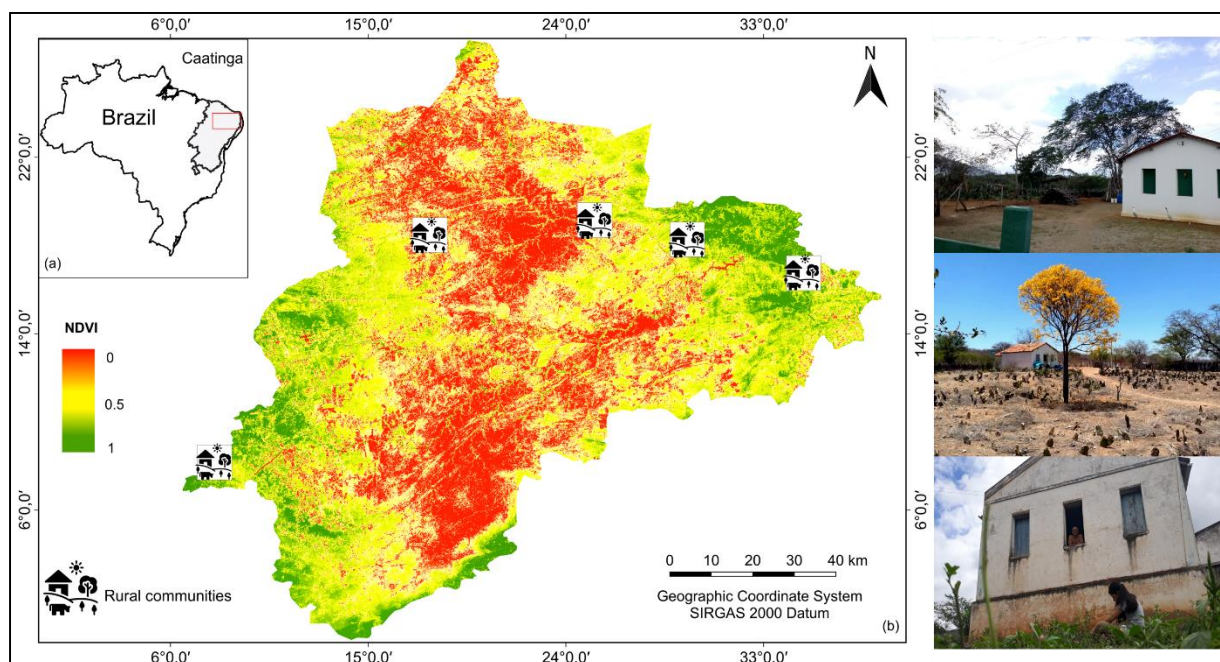
1149 In this study, we examined the relationship between the functional trait attributes of
1150 plant species (*sensu* Violle et al. 2007, i.e., the different values or states that functional traits
1151 can take in a given species, population or individual) and the selection and use of plants by
1152 local human populations. To this end, we gathered data on LEK and on functional traits to
1153 delineate a functional profile of the preferred species used for fuel, construction, and forage
1154 by agropastoralists from the Brazilian semiarid. We hypothesized that for each primary use
1155 that people give to plants (i.e., fuel, construction and forage), humans based their selection on
1156 certain functional trait attributes, which ultimately results in a set of species that best meet the
1157 needs for each type of use. Specifically, we hypothesized that for fuel and construction uses
1158 stem traits would be more important for plant selection, whereas for plants used as forage leaf
1159 traits would be more important.

1160 **2. Materials and methods**

1161

1162 *2.1. Study area*

1163 We conducted the study in the Brazilian semiarid region, specifically in the central
1164 portion of Paraíba State. The study involved local human populations from five rural
1165 communities distributed across the region (Fig. 1). Regional climate is Bsw, i.e., hot, semi-
1166 arid (Alvares et al. 2013), with two well-defined seasons: a rainy period (March to July), and
1167 a period of drought (August to February). Total annual precipitation varies greatly in space
1168 and time, with records as low as 300 mm (the lowest average total annual precipitation in
1169 Brazil) and no higher than 750 mm, and high evapotranspiration rates (< 0.65) (Moro et al.
1170 2015; Souza et al. 2020).



1171

1172 **Fig. 1.**

1173 The prevailing vegetation is seasonally dry tropical forest, known in Brazil as
1174 Caatinga. The flora of this ecosystem encompasses many species with morpho-physiological
1175 adaptations to water deficit, including leaf area reduction, thick cuticles, presence of wax and
1176 trichomes on leaves, presence of spines and thorns, water-storing tissues, stomata closure

1177 mechanisms, CAM metabolism, and deciduousness (de Queiroz et al. 2018; Fernandes et al.
1178 2020). The environmental variation across the region creates a vegetational mosaic with large
1179 differences in community structure and species richness and composition even over short
1180 distances (Moro et al. 2016; Silva and Souza 2018).

1181 The study region comprises 31 subdivisions (*municípios*) hosting a total population of
1182 190,367 inhabitants (15.45 inhabitants per km²), of which 90,030 people live in the rural
1183 environment. The human presence in the region dates to approximately 10,000 years BP
1184 (Souza et al. 2020). In the 19th century the region underwent strong modifications, when a
1185 wave of colonizers brought about the foundation of new population centers and the clearing of
1186 forest for agriculture, including cash crops such as cotton and sisal, and intensified the
1187 extraction of charcoal and firewood. During the same period extensive livestock farming
1188 expanded, using the native vegetation as a source of fodder (Silva et al. 2018).

1189

1190 *2.2. Selection of rural communities and interviewees*

1191 The study sites were selected based on satellite imagery, precipitation maps and field
1192 surveys. We selected rural communities having different levels of vegetation cover and annual
1193 rainfall in order to represent a resource availability gradient, thus maximizing the amount of
1194 information on the use and selection of plant species. The degree of vegetation cover was
1195 assessed through the Normalized Difference Vegetation Index (NDVI), obtained with ArcGIS
1196 (Fig. 1). Satellite images were acquired in the dry season of 2017 (downloaded from
1197 earthexplorer.usgs.gov; Rouse et al. 1974). Mean annual precipitation (MAP) data for each
1198 area were obtained from WorldClim (<http://www.worldclim.org/>), updated with monthly data
1199 and overall means for the 1970–2000 period, with a 30" (~1 km²) spatial resolution (Hijmans
1200 et al. 2005). These data were collected with the ‘raster’ package (Hijmans 2021) in R (R Core
1201 Team 2019). MAP ranged among sites from 379 to 690 mm.

1202 We approached local agropastoralists from the five rural communities for conducting
1203 interviews on plant use by them: Curral do Meio (São João do Cariri *minicípio*, 30
1204 interviewees), Caiçara (Cabaceiras *município*, 32), Alto dos Cardeiros (Barra de Santana
1205 *município*, 24), Viveiro (São José dos Cordeiros *município*, 23), and Olho d'Água (Monteiro
1206 *minicípio*, 11) (Fig.1). We defined agropastoralists as those people who make a living from
1207 family farming and raising small herds, and who possess ample experience with these
1208 activities. The rural economy of these communities has been traditionally dominated by cattle
1209 production (bovine, caprine and ovine) along with rain-fed subsistence agriculture, with
1210 maize, beans, sweet potatoes, watermelon, and squash being the main cultivated produce
1211 (Ramos et al. 2023). Cattle is raised in an extensive format and the animals often roam freely
1212 in the native vegetation. Even though most households have gas ranges, firewood and wood
1213 for charcoal production are still gathered as part of the traditions of many inhabitants.

1214

1215 *2.3. Data collection*

1216 The analyzes focused on plant use categories that are currently the most common in
1217 the region and are related to changes in vegetation patterns in our study area (Forage, Fuel,
1218 and Construction), and their respective use subcategories. During the interviews, the
1219 agropastoralists also mentioned characteristics (i.e., plant trait attributes) that, in their view,
1220 rendered some plants to be less preferred or even rejected by them; thus we were able to also
1221 gathered information on plants apparently not being used (and classified them as 'not used').
1222 By combining information on LEK and functional ecology, we were able to delineate a
1223 functional profile for plants undergoing high use pressure and for those suffering little
1224 pressure of use.

1225 Initially, we approached the study communities through informal conversations with
1226 the first people we had contact with. These people were usually at their work sites, either

1227 tending animals or their plantations. In these conversations we disclosed our research
1228 objectives and, following a snowball sampling technique (Bailey, 1994), asked them to
1229 recommend us other neighbors involved in agriculture and livestock-related activities that
1230 were considered equally experienced, and who had lived in the community for long times.
1231 The process was repeated until saturation or until reaching the desired sample size. In total we
1232 conducted 120 interviews with agropastoralists (men and women) over 12 months of field
1233 work (April 2019–February 2020). Information on plant use and knowledge was obtained by
1234 asking the interviewee to freely list as many plants as possible (Albuquerque et al. 2014) for
1235 each main use category (Fodder, Fuel, Construction). Next, we used semi-structured formats
1236 to collect information about local knowledge. Thereafter, we inquired about the characteristics
1237 of each species, which prompted them to remember and mention other useful species. The
1238 interviews were supplemented with other research techniques, such as participatory
1239 observation, walks through the forest around the homes and in places used to gathering
1240 resources, as well as the use of visual stimuli, i.e. using photographs of plants occurring in the
1241 region (Albuquerque et al. 2014; Medeiros et al. 2014). For each use category, the
1242 agropastoralists mentioned some visual and/or tactile attributes, such as leaf texture, size, and
1243 thickness, phenology, nutritional quality, aromas in leaves or stems, presence of oils or latex,
1244 trunk thickness, wood hardness, and the arrangement of wood fibers. In addition, interviewees
1245 highlighted their preferred plants as well as those they do not like, and justified their
1246 preferences and selection criteria. A plant species was classified as ‘preferred’ if at least 20%
1247 of respondents agreed with such classification; the remaining species were classified as
1248 ‘common’ (Valencia et al. 2015).

1249 When more than one part of the same plant was associated with a particular use, we
1250 made a separate record for each plant part. For example, interviewee 1 stated the use of
1251 leaves, fruits, flowers and bark of *Cenostigma pyramidale* as fodder (three different

1252 applications). Regarding unused plants, we asked which species were in this category and
1253 asked them to justify this judgment by listing characteristics that make it useless or
1254 unattractive.

1255 The walks through the forest with the interviewees allowed us to supplement the
1256 information on vernacular names and plant uses, and to collect specimens of species that
1257 could not be identified in the field. These specimens were deposited at the Manoel Arruda
1258 Camara Herbarium (ACAM, Universidad Estadual da Paraíba) for future reference. Initial
1259 taxonomic identifications were done with the help of local and regional floristic studies
1260 (Araujo et al. 2010; Cabral et al. 2013); later, names of plant taxa were verified using Flora do
1261 Brasil (The Angiosperm Phylogeny Group 2016).

1262

1263 *2.4. Functional trait selection*

1264 To analyze the association between LEK and the functional traits of plant species used
1265 by the agropastoralists, we integrated a dataset containing these two types of information.
1266 Based on these information, we first established broad associations with important functional
1267 traits recorded in the literature; this allowed us to select eight functional traits for which
1268 there were data available in the literature: maximum stem diameter, maximum plant height,
1269 wood density, leaf thickness, leaf area, specific leaf area, and leaf nitrogen content. For the
1270 few species that lacked information on these functional trait attributes, we assigned the mean
1271 value of all congeneric species. Trait values (i.e., attributes) were extracted from specialized
1272 literature (peer-reviewed articles, dissertations, and academic theses that have collected
1273 functional traits). In addition, we retrieved data available in the TRY database
1274 (<https://www.try-db.org/TryWeb/Home.php>). We used the standardized values given, which had
1275 been converted to uniform units of traits.

1276

1277 2.5. Data analysis

1278 First, we defined functional groups of the plant species recorded through the interviews
1279 through a cluster analysis using a matrix containing the eight selected plant functional traits for
1280 each species. The dendrogram was constructed based on a Gower dissimilarity matrix with the
1281 UPGMA (Unweighted Pair-Group Method using Arithmetic Averages) clustering method.
1282 Determination of the functional groups was made with the NbClust package (Charrad et al.
1283 2014), which offers 30 indices to determine the best number of classes and suggests to the user
1284 what the best clustering scheme is. The clusters so produced were used as vectors to be
1285 represented in a principal component analysis (PCA). Next, to evaluate each use category,
1286 indicated by agropastoralists, we constructed generalized linear models (GLM) by using the
1287 number of mentions for each use category as response variables and the plant functional traits
1288 as predictor variables. To prevent problems derived from multicollinearity among predictor
1289 variables, we calculated the variance inflation factor (VIF) for each of them with the *car*
1290 package (Sugawara and Nikaido 2014). All variables included in our models had VIF values <
1291 4, suggesting independence among them (Bondell and Reich 2008). We also tested correlations
1292 between all predictor variables. For count variables a negative binomial distribution was used
1293 (Crawley 2007). Finally, we applied a model selection procedure based on the sample corrected
1294 Akaike Information Criterion (AICc); models with $\Delta AICc < 2$ were considered equally
1295 supported and capable of explaining the variation of the response variable. A total of 147
1296 models were constructed for the total set of response variables (Fuel, Construction, Fodder, and
1297 Not Used). All analysis were performed in R ver. 3.6.0 (R Core Team, 2019).

1298

1299

1300

1301

1302 **3. Results**

1303

1304 *3.1. Classification of plants by agropastoralists*

1305 We recorded a total of 4,938 mentions of use for 82 plant species distributed among 71
1306 genera and 24 families (Appendix A). The largest number of mentions of use corresponded to
1307 the Forage category (2,673 mentions; 72 species), and this was followed by the Construction
1308 (1,103 mentions, 42 species) and Fuel (1,070 mentions, 50 species) categories; in turn, 32
1309 plant species were mentioned (84 mentions) as not having any use as they lacked useful
1310 characteristics for any use category (Fig. 2; Appendix A).

1311



1312

1313 **Fig. 2.**

1314

1315 *3.2. Preferred plants by agro-pastoralists*

1316 Agropastoralists select plants within each use category according to the traits whose
1317 attributes they associate with desirable goods and services, *e.g.*, leaf texture, size and wood
1318 hardness, deciduousness, protein content, tree height and diameter, and wood quality (Table
1319 1).

1320 **Table 1.** Attributes mentioned by interviewed agro-pastoralists from the semiarid region of
 1321 Brazil.

Use category	Attributes	
	Positive	Negative
Fuel	Abundant ember production	Toxic smoke
	Hot ember	Soft wood
	Little waste production	The fire dies out readily
	Combustion power	Produces much waste
	Lasting ember	Short longevity
	Dark wood	Noisy snapping
	Abundant crumbs	Unpleasant smell during
	Little smoke production	combustion
	Hard wood	Wet wood
	Pleasant smell	Produces much smoke
	Ignition	
	High fire durability	
High durability of the flame		
Construction	Fast growth	Low longevity and durability
	Dark color	Soft wood
	High longevity and durability	Light and light-colored wood
	Heavy wood	Twisted wood
Fodder	Leaves with high vitamin content	Bitter
	Increases milk production	Bad smell
	Helps animals to fatten	Produces helminthiasis (worm
	Sprouts	infestation)

Nice smell	Astringent (caustic) leaves
Soft leaves	Toxic foliage
Thin and tender leaves	
Velvety leaves	
Sweet	
Prevents helminthiasis	
Smooth leaves	

1322

1323 Among the total number of species mentioned in the interviews, there is a set of
1324 preferred species within each use category (Table 2). We recorded 29 indicators used by
1325 agropastoralists to describe good-quality plants in each category and 18 indicators to describe
1326 low-quality plants whose use is avoided. Some indicators for the Forage category were related
1327 to the intrinsic characteristics of the plants, other indicators highlighted the behavior of the
1328 cattle, and a last group highlighted the effects and impacts of Forage plants on cattle's health.

1329 A large proportion of the interviewees (83.2 %) expressed a preference for some plant
1330 species, and 12 species were mentioned as preferred in all use categories. Fifteen species were
1331 mentioned in each of the Construction and Fuel categories, and 29 species were pointed as
1332 preferred for Fodder. Regarding those species with a shared preference among categories,
1333 *Cenostigma pyramidale* and *Mimosa ophthalmocentra* stood out with 307 and 281 mentions,
1334 respectively (Table 2).

1335 Agropastoralists also stated that some species are less preferred, based on attributes
1336 such as hard and bitter leaves in the case of Fodder use, and too soft wood for Fuel and
1337 Construction uses (Table 1). Regarding the Construction category, specifically for the uses in
1338 the trough and hedge subcategories, the interviewees select species with low wood density,
1339 according to the literature, unlike the other uses within this category.

1340

1341 **Table 2.** Checklist of preferred species, with their numbers of mentions in each use category.

1342 Use categories: Fuel, Cons (Construction), and Fodder. The figures in the last column (%)

1343 represent the overall percentage preference of each species mentioned by the interviewees.

Species	Fuel	Cons	Fodder	%
<i>Cenostigma pyramidale</i> (Tul.) Gagnon & G.P.Lewis	62	13	26	68.08
<i>Croton blanchetianus</i> Baill.	1	18	20	52.13
<i>Mimosa tenuiflora</i> (Willd.) Poir.	28	15	21	45.75
<i>Astronium urundeuva</i> (M.Allemão) Eng.	21	10	18	38.30
<i>Mimosa ophthalmocentra</i> Mart. ex Benth.	20	24	12	37.23
<i>Prosopis juliflora</i> (Sw.) DC.	16	10	17	31.92
<i>Aspidosperma pyrifolium</i> Mart.	4	16	6	23.40
<i>Ziziphus joazeiro</i> Mart.	0	0	21	20.21
<i>Spondias tuberosa</i> Arruda	0	0	18	17.02
<i>Piptadenia retusa</i> P.G.Ribeiro, Seigler & Ebinger	5	2	12	15.96
<i>Cynophalla flexuosa</i> (L.) J.Presl	0	1	13	10.64
<i>Schinopsis brasiliensis</i> Engl.	6	1	4	10.64
<i>Anadenanthera colubrina</i> (Vell.) Brenan	8	2	1	9.57
<i>Bauhinia cheilantha</i> (Bong.) Steud	1	3	5	7.45
<i>Sideroxylon obtusifolium</i> (Roem. & Schult.) T.D.Penn.	0	0	6	6.38
<i>Combretum leprosum</i> Mart.	28	2	2	4.26
<i>Croton heliotropiifolius</i> Kunth	1	2	0	3.19
<i>Ceiba glaziovii</i> Kuntze k. Schum	0	0	2	2.13
<i>Lippia grata</i> Schauer	0	1	1	2.13
<i>Mimosa arenosa</i> (Willd.) Poir.	1	0	1	2.13
<i>Tabebuia aurea</i> (Silva Manso) Benth. & Hook.f. ex S.Moore	0	0	2	2.13
<i>Commiphora leptophloeos</i> (Mart.) J.B.Gillett	0	0	1	1.06

<i>Euphorbia tirucalli</i> L.	0	0	1	1.06
<i>Lantana camara</i> L.	0	0	1	1.06
<i>Leucaena leucocephala</i> (Lam.) de Wit	0	0	1	1.06
<i>Lachesiodendron viridiflorum</i> (Kunth) P.G.Ribeiro, L.P. Queiroz & Luckow	0	0	1	1.06
<i>Melochia tomentosa</i> L.	0	0	2	1.06
<i>Mimosa lewisii</i> Barneby	1	0	0	1.06
<i>Mimosa paraibana</i> Barneby	0	0	1	1.06
<i>Pseudobombax marginatum</i> (A.St.-Hil., Juss. & Cambess.) A.Robyns	0	0	1	1.06
<i>Senna spectabilis</i> (DC.) H.S.Irwin & Barneby	4	0	2	1.06

1344

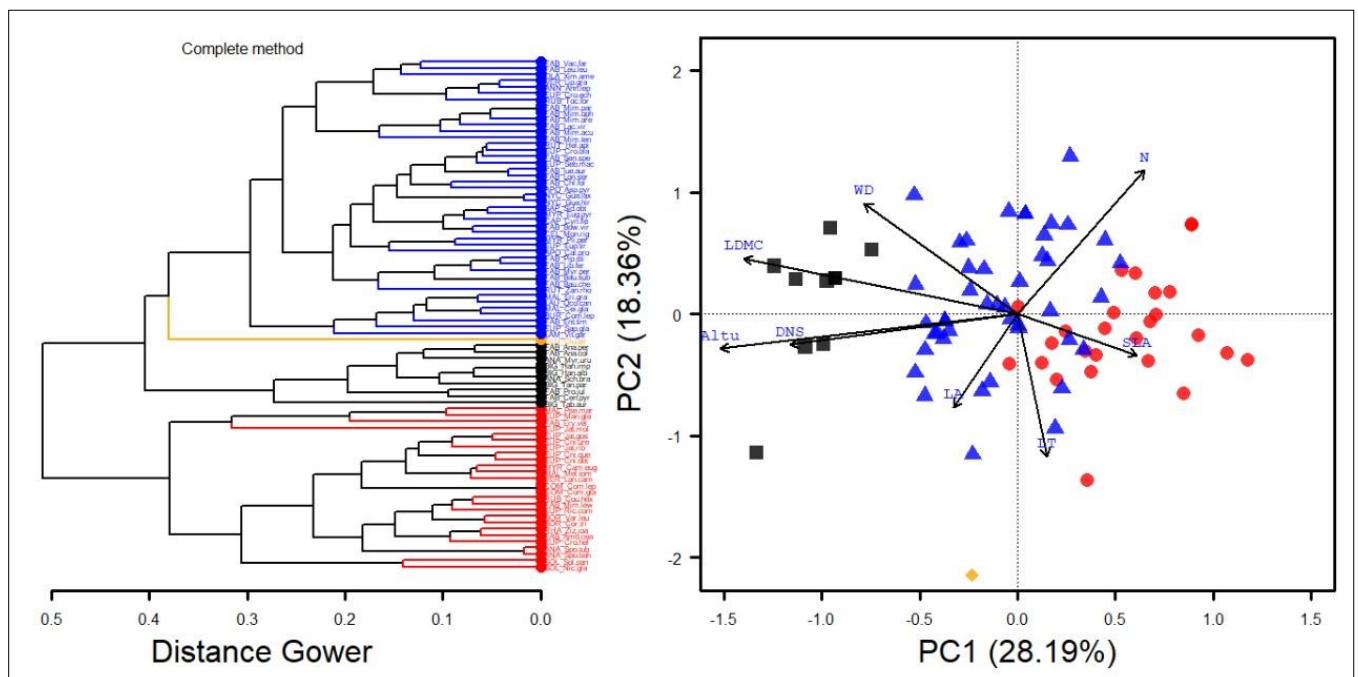
1345 *3.3. Relationships between local preferences and functional traits*

1346 The functional dendrogram rendered three major groups plus a minor one of species
1347 according to their functional trait attributes (Fig. 3a). At the opposite (upper and lower) ends
1348 of the dendrogram are species that cluster together based on the functional similarity in the
1349 measured plant traits. These two groups include species mentioned in the Fodder category,
1350 along with species that are not used. *Ziziphus joazeiro* and *Spondias tuberosa*, for example,
1351 are included in the red group, at the lower end of the dendrogram, and these are the two most
1352 frequently cited species in the Fodder category (Table 2), thus being highly preferred by the
1353 agropastoralists. However, species with low preference for Fuel and Construction uses were
1354 also present in the same group, such as *Manihot glaziovii* and *Jatropha mollissima* (Fig. 3a),
1355 both mentioned as having poor quality wood and other little desirable characteristics such as
1356 producing “excessive smoke” and “weak fire”. Importantly, all species included in the red
1357 cluster were mentioned as used for Fodder and are associated with high specific leaf area and
1358 leaf thickness values (Fig. 3b). At the upper end of the dendrogram (blue cluster) are
1359 concentrated most low-preferred species in all use categories (Fig. 3a); ten of them are shared

1360 across all use categories (Appendix A). In the central part of the dendrogram (black cluster)
1361 *Cenostigma pyramidale*, *Prosopis juliflora*, *Astronium urundeuva*, *Schinopsis brasiliensis*,
1362 *Anadenanthera colubrina* and *Tabebuia aurea* are present (Fig. 3a), all of them being among
1363 the 20 most frequently mentioned species in the study, mainly in the Fuel and Construction
1364 categories (Appendix A). This cluster is associated with height, stem diameter, wood density
1365 and leaf dry mass (Fig. 3 b) and encompasses species that differ considerably in terms of their
1366 functional profiles. Finally, the last cluster (yellow) only included one species (*Clusia*
1367 *paralicola*), which has high leaf thickness and was mentioned a single time in the interviews.

1368 In general, the dendrogram revealed that many of the plant species recorded in the
1369 interviews have multiple uses, in addition to a lower selectivity for the fodder category
1370 compared to the other uses.

1371



1372

1373 **Fig. 3.**

1374 The assessment of functional traits of the plant species mentioned in the interviews
1375 showed that despite being interrelated, as revealed by the PCA (Fig. 3b), the different traits
1376 tend to be associated with different uses. The best models support these results. In general,

1377 wood density (WD) and maximum diameter (Dmax) were strong predictors of plant selection
1378 for Fodder, Fuel and Construction uses (Table 3), as these traits were included in all best
1379 models (Appendix B).

1380 The selected models fitted to evaluate the role of plant traits associated to Fodder use
1381 revealed 16 equally plausible models (with leaf area, specific leaf area, leaf thickness and
1382 nitrogen concentration). The best model ($\Delta AICc = 0$) included three factors, namely maximum
1383 diameter and wood density (with positive effects), and specific leaf area (with a negative effect)
1384 on plant selection for Fodder, indicating that plants with high specific leaf area, values tend to
1385 be under-consumed or less preferred as forage.

1386 Notably, for wood-related uses, no model selected included mean height. We included
1387 this trait in the modeling because we had anticipated this to be an important predictor of wood
1388 use. Fuel, a use category that encompasses species in the firewood and charcoal use
1389 subcategories, was best described by three models. The first one included mean diameter, wood
1390 density (both functional traits being linked to wood quality) and leaf area, all with positive
1391 effects. The second-best model contained, in addition to wood density and mean diameter, leaf
1392 area and leaf nitrogen content, with positive and negative effects, respectively, suggesting that
1393 plants with high nitrogen contents are not preferred for Construction. For the Construction use
1394 category, three models were equally plausible, two of them including two terms (first model,
1395 maximum diameter and wood density; second model, maximum diameter and leaf dry matter
1396 content), whereas the last model included wood density, maximum diameter and leaf dry matter,
1397 all with positive effects.

1398 The condition of not being used recorded for some plants seems to be determined by
1399 high leaf area and specific leaf area values, both of them being predictor variables included in
1400 the two best models selected to explain the plants mentioned as Not used. The first model, in
1401 addition to these two traits (positive effects), included the mean diameter (with a negative

1402 effect), suggesting that low-diameter plants do not offer many benefits.

1403

1404 **Table 3.** Best-supported generalized linear models (GLM) fitted to analyze the relationship
1405 between functional traits of plants mentioned by interviewed agropastoralists and the use they
1406 make of them, in the semi-arid region of Brazil. Dmax, maximum diameter (cm); LA, leaf area
1407 (cm²); SLA, specific leaf area (mm² mg⁻¹); WD, wood density (g cm⁻³).

1408

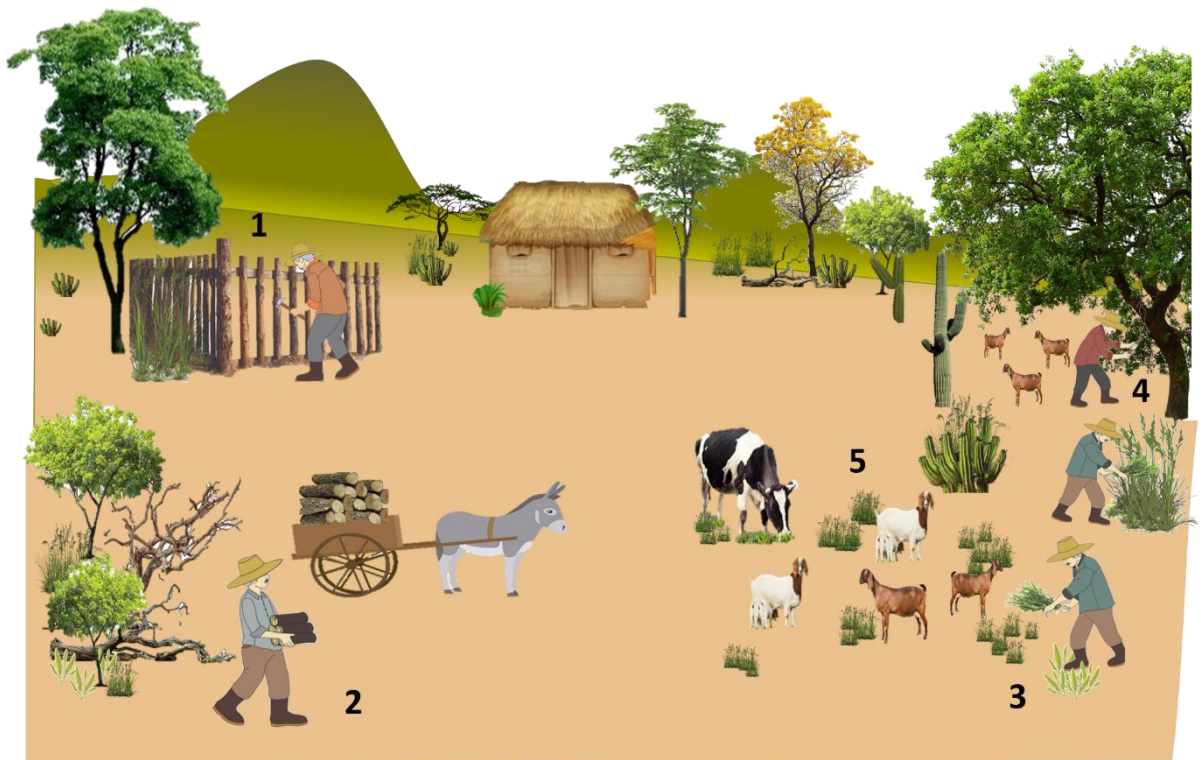
Response variable (Use category)	Predictor variable (Plant trait)	Estimate
Fuel	Dmax	0.483
	DM	0.708
	AF	0.348
Construction	Dmax	0.636
	DM	0.500
Forage	Dmax	0.649
	DM	0.208
	SLA	-0.442
Not used	Dmax	-0.282
	AF	0.293
	AFE	0.529

1409

1410 **4. Discussion**

1411 Our study provides valuable information about the patterns of use of plant species in

1412 the environmentally heterogeneous semiarid region of Brazil Our findings reinforce the
1413 intrinsic relationships between rural populations and the plant species of the Caatinga, and
1414 shed light on the relationships within the functional traits/ecosystem services/LEK triad. Our
1415 results confirm the importance of plant functional traits in decision making about the use and
1416 management of plant species. Selection by rural human populations always considers the
1417 benefit/resource use maximization ratio(Gaoue et al. 2017; Albuquerque et al. 2019; Cunha et
1418 al. 2022). Nevertheless, some relationships between the use of plant species and their
1419 functional traits are not so clear, and thus we explore here some paths to understand them.



1420

1421 **Fig. 4.**

1422 *4.1. Plant use and selection by rural communities*

1423 The largest number of mentions for the Fodder category reinforces the dependence of
1424 local populations on local biodiversity elements for the nourishment of livestock herds in the
1425 Caatinga. Also, it reveals the magnitude of LEK about the behavior and preferences of
1426 animals related to sources of forage and demonstrates the great versatility in the use of plant

1427 parts, apontada, como uma estratégia para otimização no uso dos recursos (Linstädter et al.
1428 2013). Among the species mentioned for Fodder, *Cenostigma pyramidalis* stands out, as it
1429 emerged as one of the most highly preferred species by ruminants (Cunha et al. 2022). This
1430 legume is characterized by high leaf protein content, an important indicator of nutritional
1431 quality (Chettri and Sharma 2009; Nunes et al. 2015). Moreover, the high regional abundance
1432 of this species appears to be related to its high frequency of mentions of use, since some
1433 ecosystem services largely depend on the characteristics of the dominant species (Santos et al.
1434 2009; Araujo et al. 2010; Díaz et al. 2011; Ramos et al. 2023).

1435 The numbers of plant species mentioned for the Construction and Fuel use categories,
1436 were similarly low. This may be a consequence of the gradual increase in the socioeconomic
1437 status of local populations, which seems to be leading to some abrupt changes in the
1438 dependence on plant use for those categories. For example, at present cooking gas is widely
1439 used in the region as a domestic source of energy, including the homes of all interviewees
1440 (personal observations). Yet, several cultural and spatial factors still require using native
1441 species for fuel (Hora et al. 2021; Pedrosa et al. 2022). *Cenostigma pyramidalis* was also
1442 highlighted in the Fuel category, which may be related to its high wood density, as well as the
1443 low moisture content reported by the interviewees (Table 1).

1444 The knowledge that agropastoralists have about herd behavior and plant traits related
1445 to their selection as fodder has been much emphasized (Molnár et al. 2020). The specificity
1446 and variety of this use promotes a growing body of knowledge with a high level of detail.
1447 People develop intrinsic relationships with their environment from experiences and
1448 observations; this relationship provides them with an ever growing body of information and
1449 peculiar understandings (Naah and Guuroh 2017), allowing them to maximize resource use
1450 based on plant functional traits of the species, weighted by their availability. Notably, the
1451 criteria defining the preferences are many and variable among use categories (Cámara-Leret

1452 et al. 2017).

1453

1454 4.2. Functional traits involved in plant species preferences of agropastoralists

1455 Based on a qualitative approach, Caatinga agropastoralists seem to be able to
1456 distinguish plant species according to their functional traits, relating them to the perceived
1457 quality of the used species, including high palatability, combustion power, and wood density
1458 (Table 1). For the Fodder category, leaf dry matter content was included in the second-best
1459 model; this is interesting, as this trait has been linked to a quality attribution to resources
1460 (Chettri and Sharma 2009). *Spondias tuberosa*, one of the preferred species for Fodder, stood
1461 out for the versatility of the parts used (leaves, branches, and especially fruits); according to
1462 the agropastoralists, the sweet flavor of its fruits is highly appreciated by ruminants.
1463 Interestingly, and considering all plant traits analyzed, this species belongs to the group of
1464 species positively associated with specific leaf area.

1465 Overall, agropastoralists mentioned their preference for plant species with bigger
1466 diameter and higher wood density but smaller leaf area and, by contrast, their tendency to
1467 reject plants with larger leaf area and specific leaf area, and thicker leaves. Native plant
1468 diversity provides human populations with a range of critical ecosystem goods and services
1469 (Díaz et al. 2007; Brown et al. 2011). However, the way people benefit from these services
1470 varies according to their perceptions, needs, and ability to access and select the different
1471 species (Díaz et al. 2011; Nunes et al. 2016; Cunha et al. 2022). In arid and semi-arid
1472 ecosystems in particular, maintaining the provision of ecosystem services requires a deep
1473 understanding of forest resource dynamics (Linstädter et al. 2013), due to the different
1474 strategies of plants species selection and use. In the case of Fodder use, the consumption of
1475 different plant parts reveals the choice of different plant traits for a single purpose.

1476 Several studies have reported people's preferences for some species for Fodder, Fuel

1477 and Construction, with quality and availability being the main drivers of such preferences
1478 (Chettri and Sharma 2009; Nunes et al. 2015; Cámara-Leret et al. 2017; Gonçalves et al.
1479 2021a; Hora et al. 2021). For Fodder use, for example, palatability was a significant criterion,
1480 as it relates to greater acceptability by ruminants. Regarding the Construction and Fuel
1481 categories, respectively, wood durability, combustion potential, and flame duration are the
1482 most relevant criteria (Santos et al. 2009; Marquez-Reynoso et al. 2017; Hora et al. 2021).

1483 Agropastoralists often select plants for Construction through the judgement of
1484 functional traits related to wood quality, such as wood density and stem diameter, for
1485 example. This means that, by adopting the strategy of selecting plants with greater durability,
1486 they avoid the need to frequently replenish their stocks. However, this judgement also results
1487 in plant use being more heavily concentrated on a small set of highly durable species, which
1488 can exert strong pressure on them, progressively leading to their population decline and
1489 slowly increasing the pressure on lower quality species to replace them. Therefore, the
1490 perception of durability is a major factor in the selection of species, and it plays an important
1491 role for burning fuel, as well as in the daily dynamics of the life of rural populations (Santos
1492 et al. 2009; Marquez-Reynoso et al. 2017). By using more durable species, it is likely that
1493 more time can be spent on other activities such as herding, housework, or farming.

1494 The positive relationship between wood density and all use categories examined here
1495 highlights how important this plant trait is for selecting plant species in the Caatinga,
1496 especially for rural constructions, which is the subcategory that received the largest number of
1497 mentions in the Construction category (Table 2). This relationship is also supported by the
1498 regional availability of high-wood density species, such as *Cenostigma pyramidale*, *Mimosa*
1499 *tenuiflora* and *M. ophthalmocentra* (Araujo et al. 2010). Considering that high wood density
1500 is a conservative water use attribute (Poorter et al. 2019; Romero et al. 2020; Romero et al.
1501 2022), in dry environments it is expected that there will be a dominance of these species and

1502 consequently a greater exploitation.

1503 The classification of several species as not used reinforces the specificity in selecting
1504 plants possessing desirable characteristics; plants with high specific leaf area, and high leaf
1505 area along with smaller stem diameters are clearly unattractive and thus do not seem to be
1506 useful for rural populations in the Caatinga; notable examples of this situation are *Jatropha*
1507 *molissima*, *J. ribifolia*, and *Cnidocolus urens*.

1508

1509 *4.3 Implications for conservation*

1510 Some agropastoralists expressed the concern that the availability of some preferred
1511 plant species has declined considerably over recent years, for example *Astronium urundeuva*
1512 and *Schinopsis brasiliensis*. These species were frequently reported as preferred (Table 3) for
1513 having characteristics suggested to be the best for their use as wood. Nevertheless, the current
1514 abundance of these two species is very low (Ramos et al. 2023), which has resulted in other
1515 plant species replacing them and becoming the preferred ones for this use. Such a shift is in
1516 agreement with the ecological appearance hypothesis that predicts that because the most
1517 apparent (visible) plants are more susceptible to herbivory, the most abundant useful plants
1518 experience greater pressure of use (Gaoue et al. 2016; Lucena et al., 2012; Gonçalves et al.,
1519 2016).

1520 Two contrasting opinions are currently debated in this regard. On the one hand, there
1521 is a view that the pressure derived from plant use has little effect on the abundance of
1522 dominant species, especially those with resprouting capacity, so that they would not be
1523 affected significantly, as is the case of *Croton blanchetianus*, *Cenostigma pyramidale* and
1524 *Aspidosperma pyriformium*, all of which reproduce successfully under the environmental
1525 conditions prevailing in some areas (Ramos et al. 2020; Gonçalves et al. 2021b; Ramos et al.
1526 2023). On the other hand, the use history of these plant species reveals severe and constant

1527 pressures (Singh 1998; Silva et al. 2018), and considering the region's climatic severity, many
1528 areas of the Caatinga have not been able to recover from the impacts (Schulz et al. 2019;
1529 Antongiovanni et al. 2020), which may be leading to a decline in in the populations of some
1530 species and, consequently, to the collapse of the ecosystem due to inadequate management
1531 (Molnár et al. 2020). In the global scenario, cattle grazing comprises the most common land
1532 use (Travers et al. 2018), and consequently it is one of the activities that most strongly affect
1533 both the land and the species of the local biodiversity. For these reasons, grazing can
1534 compromise economic activities, especially in arid and semi-arid regions (Schulz et al. 2019).

1535 The information produced in this study provides new insight into resource
1536 management; selective exploitation of plant resources is a growing problem that permeates
1537 across different forest systems (Ellis 2015). Intense and short-term pressures critically
1538 threaten resource regeneration and require urgent changes in the exploitation of ecosystems
1539 (Díaz et al. 2007; Brown et al. 2011).

1540 **Conclusion**

1541 Overall, the advantage of the integration of the information on plant traits with LEK
1542 allowed a better understanding of the complexity of human-biodiversity relationships of
1543 Caatinga. Our results lead to the conclusion that the functional plant traits are critical in
1544 decision-making for the selection of plant species for each primary use (i.e., fuel, construction
1545 and forage). Therefore, this approach represents a potentially powerful tool for understanding
1546 the provision of ecosystem services to human societies, allowing an exchange of information
1547 between researchers, resource managers and biodiversity conservation agencies. Will enhance
1548 our capacity to make a sound management of biodiversity and its conservation of Caatinga.

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1560

1561 **Ethics declarations**

1562 The study was approved by the Research Ethics Committee of Centro Universitário
1563 UNIFACISA (appraisal No: 010713/2019; certificate of presentation for ethical review No.
1564 07564918.2.0000.5175). Before starting data collection *in situ*, we explained our research
1565 objectives to the informants, and made clear to them that all information would be treated
1566 confidentially. We did this to get their consent before the interviews started and also to
1567 establish a relationship of trust with them. We obtained oral consent from those who agreed to
1568 participate, and later they signed the free and informed consent form.

1569

1570 **Competing Interests**

1571 The authors declare no competing interests.

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1577 **MBR:** Investigation, Conceptualization, Data acquisition (field work), Data analysis, Writing-
1578 Original draft preparation, Visualization, Writing - Review and Editing, Project administration.

1579 **HAA:** Investigation, Conceptualization, Data analysis, Writing- Original draft preparation,
1580 Visualization, Writing - Review and Editing

1581 **MGRM, SSC, SMS, and KMP** Investigation, Conceptualization, Data acquisition (field
1582 work), Data analysis, Writing - Review and Editing.

1583 **JAM:** Conceptualization, Data analysis, Visualization, Writing- Original draft preparation,
1584 Writing - Review and Editing.

1585 **SFL:** Conceptualization, Data analysis, Funding acquisition, Supervision, Project
1586 administration, Visualization, Writing - Review and Editing.

1587

1588 **Data Availability**

1589 N/A

1590

1591 **Literature Cited**

1592 Albuquerque, U. P., L. V. F. C. Cunha., R. F. P. Lucena., and R. R. N. Alves. 2014. *Methods*
1593 *and Techniques in Ethnobiology and Ethnoecology*. Edited by Romulo Romeu Nóbrega
1594 Alves Albuquerque, Ulysses Paulino De, Cunha, Luiz Vital Fernandes Cruz da, Lucena,
1595 Reinaldo Farias Paiva de, Alves. Springer Protocols Handbooks. Springer New York.
1596 doi:10.1007/978-1-4614-8636-7.

1597 Albuquerque, U. P., P. M. de Medeiros, W. S. Ferreira Júnior, T. C. da Silva, R. R. V. da
1598 Silva, and T. Gonçalves-Souza. 2019. Social-Ecological Theory of Maximization: Basic
1599 Concepts and Two Initial Models. *Biological Theory* 14. Springer: 73–85.
1600 doi:10.1007/s13752-019-00316-8.

1601 Albuquerque, U. P. de, and R. F. de Oliveira. 2007. Is the use-impact on native caatinga
1602 species in Brazil reduced by the high species richness of medicinal plants? *Journal of*

1603 *Ethnopharmacology* 113. Elsevier: 156–170. doi:10.1016/J.JEP.2007.05.025.

1604 Alvares, C. A., J. L. Stape, P. C. Sentelhas, J. L. De Moraes Gonçalves, and G. Sparovek.
1605 2013. Köppen’s climate classification map for Brazil. *Meteorologische Zeitschrift* 22:
1606 711–728. doi:10.1127/0941-2948/2013/0507.

1607 Antongiovanni, M., E. M. Venticinque, M. Matsumoto, and C. R. Fonseca. 2020. Chronic
1608 anthropogenic disturbance on Caatinga dry forest fragments. Edited by Shekhar Biswas.
1609 *Journal of Applied Ecology*. Blackwell Publishing Ltd: 1365-2664.13686.
1610 doi:10.1111/1365-2664.13686.

1611 Araujo, K. D., H. N. Parente, É. Éder-silva, C. I. Ramalho, R. T. Dantas, A. P. de Andrade,
1612 and D. S. da Silva. 2010. Levantamento florístico do estrato arbustivo-arbóreo em áreas
1613 contíguas de Caatinga no Cariri Paraibano. *Revista Caatinga* 23: 63–70.

1614 de Bello, F., S. Lavorel, S. Lavergne, C. H. Albert, I. Boulangeat, F. Mazel, and W. Thuiller.
1615 2013. Hierarchical effects of environmental filters on the functional structure of plant
1616 communities: A case study in the French Alps. *Ecography* 36: 393–402.
1617 doi:10.1111/j.1600-0587.2012.07438.x.

1618 Berkes, F., J. Colding, and C. Folke. 2000. Rediscovery of Traditional Ecological Knowledge
1619 as Adaptive Management. *Ecological Applications* 10. Wiley: 1251.
1620 doi:10.2307/2641280.

1621 Bondell, H. D., and B. J. Reich. 2008. Simultaneous regression shrinkage, variable selection,
1622 and supervised clustering of predictors with OSCAR. *Biometrics* 64: 115–123.
1623 doi:10.1111/j.1541-0420.2007.00843.x.

1624 Brauman, K. A., G. C. Daily, T. K. eo Duarte, and H. A. Mooney. 2007. The Nature and
1625 Value of Ecosystem Services: An Overview Highlighting Hydrologic Services.
1626 <https://doi.org/10.1146/annurev.energy.32.031306.102758> 32. Annual Reviews: 67–98.
1627 doi:10.1146/ANNUREV.ENERGY.32.031306.102758.

1628 Brown, K. A., D. F. B. Flynn, N. K. Abram, J. C. Ingram, S. E. Johnson, and P. Wright. 2011.
1629 Assessing natural resource use by forest-reliant communities in madagascar using
1630 functional diversity and functional redundancy metrics. *PLoS ONE* 6.
1631 doi:10.1371/journal.pone.0024107.

1632 Cabral, G. A. de L., E. V. de S. B. Sampaio, and J. S. de Almeida-Cortez. 2013. Estrutura
1633 Espacial e Biomassa da Parte Aérea em Diferentes Estádios Sucessionais de Caatinga,
1634 em Santa Terezinha, Paraíba. *Revista Brasileira de Geografia Física* 06: 566–574.

1635 Cáceres, D. M., E. Tapella, F. Quétier, and S. Díaz. 2015. The social value of biodiversity and
1636 ecosystem services from the perspectives of different social actors. *Ecology and Society*
1637 20. doi:10.5751/ES-07297-200162.

1638 Cámara-Leret, R., S. Faurby, M. J. Macía, H. Balslev, B. Gödel, J. C. Svenning, W. D.
1639 Kissling, N. Rønsted, et al. 2017. Fundamental species traits explain provisioning
1640 services of tropical American palms. *Nature Plants* 3. Nature Publishing Group: 1–7.
1641 doi:10.1038/nplants.2016.220.

1642 Charrad, M., N. Ghazzali, V. Boiteau, and A. Niknafs. 2014. Nbclust: An R package for
1643 determining the relevant number of clusters in a data set. *Journal of Statistical Software*
1644 61: 1–36. doi:10.18637/jss.v061.i06.

1645 Chettri, N., and E. Sharma. 2009. A scientific assessment of traditional knowledge on
1646 firewood and fodder values in Sikkim, India. *Forest Ecology and Management* 257.
1647 Elsevier: 2073–2078. doi:10.1016/J.FORECO.2009.02.002.

1648 Crawley, M. J. 2007. *The R book*. *The R Book*. doi:10.1002/9780470515075.

1649 Cunha, S. S. da, M. B. Ramos, H. A. de Almeida, M. G. R. Maciel, S. M. de Souza, K. M.
1650 Pedrosa, and S. de Faria Lopes. 2022. Vegetation cover and seasonality as indicators for
1651 selection of forage resources by local agro-pastoralists in the Brazilian semiarid region.
1652 *Scientific reports* 12. Nature Publishing Group UK: 15174. doi:10.1038/s41598-022-

1653 18282-w.

1654 Díaz, S., J. Fargione, F. S. Chapin, and D. Tilman. 2006. Biodiversity Loss Threatens Human
1655 Well-Being. *PLOS Biology* 4. Public Library of Science: e277.
1656 doi:10.1371/JOURNAL.PBIO.0040277.

1657 Díaz, S., S. Lavorel, S. McIntyre, V. Falczuk, F. Casanoves, D. G. Milchunas, C. Skarpe, G.
1658 Rusch, et al. 2007. Plant trait responses to grazing – a global synthesis. *Global Change*
1659 *Biology* 13. John Wiley & Sons, Ltd: 313–341. doi:10.1111/J.1365-2486.2006.01288.X.

1660 Díaz, S., F. Quétier, D. M. Cáceres, S. F. Trainor, N. Pérez-Harguindeguy, M. S. Bret-Harte,
1661 B. Finegan, M. Peña-Claros, et al. 2011. Linking functional diversity and social actor
1662 strategies in a framework for interdisciplinary analysis of nature’s benefits to society.
1663 *Proceedings of the National Academy of Sciences of the United States of America* 108.
1664 National Academy of Sciences: 895–902. doi:10.1073/pnas.1017993108.

1665 Ellis, E. C. 2015. Ecology in an anthropogenic biosphere. *Ecological Monographs* 85: 287–
1666 331. doi:10.1890/14-2274.1.

1667 Ellis, E. C., and N. Ramankutty. 2008. Putting people in the map: Anthropogenic biomes of
1668 the world. *Frontiers in Ecology and the Environment*. Ecological Society of America.
1669 doi:10.1890/070062.

1670 Farias, R., P. De Lucena, E. D. L. Araujo, and P. De Albuquerque. 2007. Does the Local
1671 Availability of Woody Caatinga Plants (Northeastern Brazil) Explain Their Use Value ?
1672 Author (s): Reinaldo Farias Paiva de Lucena , Elcida de Lima Araújo and Ulysses
1673 Paulino de Albuquerque Published by : Springer on behalf of New York B 61: 347–361.

1674 Fernandes, M. F., D. Cardoso, and L. P. de Queiroz. 2020. An updated plant checklist of the
1675 Brazilian Caatinga seasonally dry forests and woodlands reveals high species richness
1676 and endemism. *Journal of Arid Environments* 174. Elsevier: 104079.
1677 doi:10.1016/j.jaridenv.2019.104079.

1678 Gaoue, O., J. Jiang, W. Ding, and F. B. Augusto. 2016. Optimal harvesting strategies for
1679 timber and non-timber forest products in tropical ecosystems: 286–297.
1680 doi:10.1007/s12080-015-0286-4.

1681 Gaoue, O. G., M. A. Coe, M. Bond, G. Hart, B. C. Seyler, and H. McMillen. 2017. Theories
1682 and Major Hypotheses in Ethnobotany. *Economic Botany*. Springer New York LLC.
1683 doi:10.1007/s12231-017-9389-8.

1684 Gonçalves, P. H. S., U. P. Albuquerque, and P. M. De Medeiros. 2016. The most commonly
1685 available woody plant species are the most useful for human populations: A meta-
1686 analysis. *Ecological Applications* 26: 2238–2253. doi:10.1002/eap.1364.

1687 Gonçalves, P. H. S., P. M. de Medeiros, and U. P. Albuquerque. 2021a. Effects of domestic
1688 wood collection on tree community structure in a human-dominated seasonally dry
1689 tropical forest. *Journal of Arid Environments* 193. Academic Press: 104554.
1690 doi:10.1016/j.jaridenv.2021.104554.

1691 Gonçalves, P. H. S., C. V. S. da Cunha Melo, C. de Assis Andrade, D. V. B. de Oliveira, V.
1692 de Moura Brito Junior, K. F. Rito, P. M. de Medeiros, and U. P. Albuquerque. 2021b.
1693 Livelihood strategies and use of forest resources in a protected area in the Brazilian
1694 semiarid. *Environment, Development and Sustainability* 2021. Springer: 1–21.
1695 doi:10.1007/S10668-021-01529-3.

1696 Hassen, A., K. K. Zander, S. Manes, and M. Meragiaw. 2023. Local People’s perception of
1697 forest ecosystem services, traditional conservation, and management approaches in
1698 North Wollo, Ethiopia. *Journal of Environmental Management* 330. Academic Press:
1699 117118. doi:10.1016/J.JENVMAN.2022.117118.

1700 Hijmans, R. J. 2021. Geographic Data Analysis and Modeling [R package raster version 3.5-
1701 11]. Comprehensive R Archive Network (CRAN).

1702 Hijmans, R. J., S. E. Cameron, J. L. Parra, P. G. Jones, and A. Jarvis. 2005. Very high

1703 resolution interpolated climate surfaces for global land areas. *International Journal of*
1704 *Climatology* 25. John Wiley & Sons, Ltd: 1965–1978. doi:10.1002/JOC.1276.

1705 Hora, J. S. L., I. S. Feitosa, U. P. Albuquerque, M. A. Ramos, and P. M. Medeiros. 2021.
1706 Drivers of species' use for fuelwood purposes: A case study in the Brazilian semiarid
1707 region. *Journal of Arid Environments* 185. Academic Press: 104324.
1708 doi:10.1016/J.JARIDENV.2020.104324.

1709 Linstädter, A., B. Kemmerling, G. Baumann, and H. Kirscht. 2013. The importance of being
1710 reliable – Local ecological knowledge and management of forage plants in a dryland
1711 pastoral system (Morocco). *Journal of Arid Environments* 95. Academic Press: 30–40.
1712 doi:10.1016/J.JARIDENV.2013.03.008.

1713 Lopes, S. . 2017. The other side of Ecology: thinking about the human bias in our ecological
1714 analyses for biodiversity conservation. *Ethnobiology and Conservation* 6: 1–8.
1715 doi:10.15451/ec2017.

1716 Lucena, R. F. P. de, P. M. de Medeiros, E. de L. Araújo, A. G. C. Alves, and U. P. de
1717 Albuquerque. 2012. The ecological apparency hypothesis and the importance of useful
1718 plants in rural communities from Northeastern Brazil: An assessment based on use value.
1719 *Journal of Environmental Management* 96. Academic Press: 106–115.
1720 doi:10.1016/J.JENVMAN.2011.09.001.

1721 Luo, Y., J. Liu, S. Tan, M. W. Cadotte, K. Xu, L. Gao, and D. Li. 2016. Trait variation and
1722 functional diversity maintenance of understory herbaceous species coexisting along an
1723 elevational gradient in Yulong Mountain, Southwest China. *Plant Diversity* 38. KeAi
1724 Publishing Communications Ltd.: 303–311. doi:10.1016/j.pld.2016.11.002.

1725 Marquez-Reynoso, M. I., N. Ramírez-Marcial, S. Cortina-Villar, and S. Ochoa-Gaona. 2017.
1726 Purpose, preferences and fuel value index of trees used for firewood in El Ocote
1727 Biosphere Reserve, Chiapas, Mexico. *Biomass and Bioenergy* 100. Pergamon: 1–9.

1728 doi:10.1016/J.BIOMBIOE.2017.03.006.

1729 Medeiros, P. M., A. H. Ladio, and U. P. Albuquerque. 2014. Sampling problems in Brazilian
1730 research: A critical evaluation of studies on medicinal plants. *Revista Brasileira de*
1731 *Farmacognosia* 24. Sociedade Brasileira de Farmacognosia: 103–109.
1732 doi:10.1016/J.BJP.2014.01.010.

1733 Melo, F. P. L. 2017. The Socio-Ecology of the Caatinga: Understanding How Natural
1734 Resource Use Shapes an Ecosystem. *Caatinga: The Largest Tropical Dry Forest Region*
1735 *in South America*. Springer, Cham: 369–382. doi:10.1007/978-3-319-68339-3_14.

1736 Meuret, M., and F. D. Provenza. 2015. When art and science meet: Integrating knowledge of
1737 French herders with science of foraging behavior. *Rangeland Ecology and Management*
1738 68. Elsevier: 1–17. doi:10.1016/j.rama.2014.12.007.

1739 Millennium Ecosystem Assessment. 2005. Ecosystem and human well-being : synthesis. *World*
1740 *Resources Institute*: 86.

1741 Molnár, Z., A. Kelemen, R. Kun, J. Máté, L. Sáfián, F. Provenza, S. Díaz, H. Barani, et al.
1742 2020. Knowledge co-production with traditional herders on cattle grazing behaviour for
1743 better management of species-rich grasslands. *Journal of Applied Ecology* 57. John
1744 Wiley & Sons, Ltd: 1677–1687. doi:10.1111/1365-2664.13664.

1745 Moro, M. F., I. A. Silva, F. S. De Araújo, E. N. Lughadha, T. R. Meagher, and F. R. Martins.
1746 2015. The role of edaphic environment and climate in structuring phylogenetic pattern in
1747 seasonally dry tropical plant communities. *PLoS ONE* 10: 1–18.
1748 doi:10.1371/journal.pone.0119166.

1749 Moro, M. F., E. Nic Lughadha, F. S. de Araújo, and F. R. Martins. 2016. A
1750 Phytogeographical Metaanalysis of the Semiarid Caatinga Domain in Brazil. *Botanical*
1751 *Review* 82. The Botanical Review: 91–148. doi:10.1007/s12229-016-9164-z.

1752 Naah, J. B. S. N., and R. T. Guuroh. 2017. Factors influencing local ecological knowledge of

1753 forage resources: Ethnobotanical evidence from West Africa's savannas. *Journal of*
1754 *Environmental Management* 188. J Environ Manage: 297–307.
1755 doi:10.1016/j.jenvman.2016.11.064.

1756 Nunes, A. T., R. F. P. Lucena, M. V. F. dos Santos, and U. P. Albuquerque. 2015. Local
1757 knowledge about fodder plants in the semi-arid region of Northeastern Brazil. *Journal of*
1758 *Ethnobiology and Ethnomedicine* 2015 11:1 11. BioMed Central: 1–12.
1759 doi:10.1186/1746-4269-11-12.

1760 Nunes, A. T., D. L. V. Cabral, E. L. C. Amorim, M. V. F. dos Santos, and U. P. Albuquerque.
1761 2016. Plants used to feed ruminants in semi-arid Brazil: A study of nutritional
1762 composition guided by local ecological knowledge. *Journal of Arid Environments* 135:
1763 96–103. doi:10.1016/j.jaridenv.2016.08.015.

1764 Pedrosa, K. M., M. B. Ramos, S. S. da Cunha, M. G. R. Maciel, S. M. de Souza, H. K. de L.
1765 Soares, M. de los Á. La Torre-Cuadros, and S. de F. Lopes. 2022. Local ecological
1766 knowledge dynamics of farmers in areas which have been chronically disturbed by
1767 human actions in the Brazilian Caatinga. *Ethnobotany Research and Applications* 24: 1–
1768 21. doi:10.32859/era.24.23.1-21.

1769 Poorter, L., D. M. A. Rozendaal, F. Bongers, J. S. de Almeida-Cortez, A. M. Almeyda
1770 Zambrano, F. S. Álvarez, J. L. Andrade, L. F. A. Villa, et al. 2019. Wet and dry tropical
1771 forests show opposite successional pathways in wood density but converge over time.
1772 *Nature Ecology and Evolution* 3. Nature Publishing Group: 928–934.
1773 doi:10.1038/s41559-019-0882-6.

1774 de Queiroz, L. P., D. Cardoso, M. F. Fernandes, and M. F. Moro. 2018. *Diversity and*
1775 *evolution of flowering plants of the Caatinga Domain. Caatinga: The Largest Tropical*
1776 *Dry Forest Region in South America.* doi:10.1007/978-3-319-68339-3_2.

1777 R Core Team. 2019. R: The R Project for Statistical Computing.

1778 Ramos, M. A., P. M. de Medeiros, A. L. S. de Almeida, A. L. P. Feliciano, and U. P. de
1779 Albuquerque. 2008. Can wood quality justify local preferences for firewood in an area of
1780 caatinga (dryland) vegetation? *Biomass and Bioenergy* 32. Pergamon: 503–509.
1781 doi:10.1016/J.BIOMBIOE.2007.11.010.

1782 Ramos, M. B., F. C. Diniz, H. A. Almeida, G. R. Almeida, A. S. Pinto, J. A. Meave, and S. F.
1783 Lopes. 2020. The role of edaphic factors on plant species richness and diversity along
1784 altitudinal gradients in the Brazilian semi-arid region. *Journal of Tropical Ecology* 36:
1785 199–212. doi:10.1017/S0266467420000115.

1786 Ramos, M. B., M. G. R. Maciel, S. S. da Cunha, S. M. de Souza, K. M. Pedrosa, J. J. L. L. de
1787 Souza, E. J. González, J. A. Meave, et al. 2023. The role of chronic anthropogenic
1788 disturbances in plant community assembly along a water availability gradient in Brazil’s
1789 semiarid Caatinga region. *Forest Ecology and Management* 538. Elsevier: 120980.
1790 doi:10.1016/J.FORECO.2023.120980.

1791 Romero, E., R. Dávalos-Sotelo, J. A. Meave, and T. Terrazas. 2020. Wood density, deposits
1792 and mineral inclusions of successional tropical dry forest species. *European Journal of*
1793 *Forest Research* 139. Springer: 369–381. doi:10.1007/s10342-019-01236-9.

1794 Romero, E., T. Terrazas, E. J. González, and J. A. Meave. 2022. Wood anatomy of 13 species
1795 from a successional tropical dry forest: description and ecological implications. *IAWA*
1796 *Journal* 43. Brill: 372–402. doi:10.1163/22941932-BJA10104.

1797 Rouse, J. W. . J., R. H. Haas, J. A. Schell, D. W. Deering, R. H. Haas, J. A. Schell, and D. W.
1798 Deering. 1974. Monitoring vegetation systems in the Great Plains with ERTS. *NASA.*
1799 *Goddard Space Flight Center 3d ERTS-1 Symp., Vol. 1, Sect. A.*

1800 Santos, L. L. dos, M. A. Ramos, S. I. da Silva, M. F. de Sales, and U. P. de Albuquerque.
1801 2009. Caatinga Ethnobotany: Anthropogenic Landscape Modification and Useful
1802 Species in Brazil’s Semi-Arid Northeast. *Economic Botany* 2009 63:4 63. Springer: 363–

1803 374. doi:10.1007/S12231-009-9094-3.

1804 Schulz, K., M. Guschal, I. Kowarik, J. S. de Almeida-Cortez, E. V. de S. B. Sampaio, and A.
1805 Cierjacks. 2019. Grazing reduces plant species diversity of Caatinga dry forests in
1806 northeastern Brazil. *Applied Vegetation Science* 22: 348–359. doi:10.1111/avsc.12434.

1807 Silva, A. C., and A. F. Souza. 2018. Aridity drives plant biogeographical sub regions in the
1808 Caatinga, the largest tropical dry forest and woodland block in South America. *PLoS*
1809 *ONE* 13. Public Library of Science. doi:10.1371/journal.pone.0196130.

1810 Silva, J. M. C. da, I. R. Leal, and M. Tabarelli. 2018. Caatinga: The largest tropical dry forest
1811 region in South America. *Caatinga: The Largest Tropical Dry Forest Region in South*
1812 *America*. Springer International Publishing: 1–482. doi:10.1007/978-3-319-68339-3.

1813 Singh, S. P. 1998. Chronic disturbance, a principal cause of environmental degradation in
1814 developing countries. *Environmental Conservation* 25: 1–2.
1815 doi:10.1017/S0376892998000010.

1816 Smith, B. D. 2011. A Cultural Niche Construction Theory of Initial Domestication. *Biological*
1817 *Theory* 6: 260–271. doi:10.1007/s13752-012-0028-4.

1818 Souza, J. J. L. L., B. I. de Souza, R. A. Xavier, A. A. Pacheco, L. C. R. Pessenda, and E. dos
1819 Santos Brito. 2020. Archaeoanthrosol formation in the Brazilian semiarid. *Catena* 193.
1820 Elsevier: 104603. doi:10.1016/j.catena.2020.104603.

1821 Sugawara, E., and H. Nikaido. 2014. Properties of AdeABC and AdeIJK efflux systems of
1822 *Acinetobacter baumannii* compared with those of the AcrAB-TolC system of *Escherichia*
1823 *coli*. *Antimicrobial Agents and Chemotherapy* 58: 7250–7257. doi:10.1128/AAC.03728-
1824 14.

1825 The Angiosperm Phylogeny Group. 2016. An update of the Angiosperm Phylogeny Group
1826 classification for the orders and families of flowering plants: APG IV. *Botanical Journal*
1827 *of the Linnean Society* 181. John Wiley & Sons, Ltd: 1–20. doi:10.1111/boj.12385.

1828 Travers, S. K., D. J. Eldridge, J. Dorrough, J. Val, and I. Oliver. 2018. Introduced and native
1829 herbivores have different effects on plant composition in low productivity ecosystems.
1830 *Applied Vegetation Science* 21: 45–54. doi:10.1111/avsc.12334.

1831 Violle, C., Navas, M. L., Vile, D., Kazakou, E., Fortunel, C., Hummel, I., & Garnier, E.
1832 (2007). Let the concept of trait be functional! *Oikos* 116: 882–892. doi: 10.1111/j.0030-
1833 1299.2007.15559.x

1834

Figure Captions

Fig. 1. Location of the study area and vegetation cover map of the Cariri region, Paraíba State, NE Brazil, in the Caatinga realm (shaded). The icons indicate the location of the rural communities included in the study. The map was created using ArcGIS 10.5.1 (ESRI, <https://www.esri.com>). Images on the right side of the figure exemplify the livelihoods of the pastoralists living in this region.

Fig. 2. Different plant uses made by agropastoralists of the Cariri region, Paraíba State, NE Brazil. (A) Agropastoralist with branches of *Cereus jamacaru* used as fodder for the herds during the dry season. (B) Herds confined in pens made with stems of *Croton blanchetianus* and *Aspidosperma pyriforme*. (C) Wood stove. (D) Agropastoralist cutting branches of *Schinopsis brasiliensis* to feed the herds. (E) Herds confined in a pen being fed by agropastoralists with native plant species. (E) *Prosopis juliflora* timber used for rural construction.

Fig. 3. Multivariate analyses of the species used by agropastoralists in the Cariri region and their functional traits. (A) UPGMA dendrogram for 82 species with functional traits mentioned by the interviewed agropastoralists. (B) PCA ordination of the trait functional space. The vectors represent functional traits: LDMC, leaf dry matter content; Height, maximum height; BD, basal diameter; WD, wood density; LT, Leaf thickness; LA, Leaf area; SLA, Specific leaf area; N, Leaf nitrogen.

Fig. 4. Agropastoralists knowledge of plant services. Daily life agropastoralists in where plants products play a central role: (1) Agropastoralists using *Croton blanchetianus*, *Aspidosperma pyriformium*, *Astronium urundeuva* for fence construction. (2) Agropastoralists using *Cenostigma pyramidale*, *Croton blanchetianus*, *Mimosa ophthalmocentra*, for fuel. (3, 4) Agropastoralists supplying *Mimosa tenuiflora*, *Schinopsis brasiliensis*, and *Ziziphus joazeiro* branches to the herds. (5) Herds grazing freely through the Caatinga vegetation.

4. CAPÍTULO IV: The role of chronic anthropogenic disturbances in plant community assembly along a water availability gradient in Brazil's semiarid Caatinga region

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The role of chronic anthropogenic disturbances in plant community assembly along a water availability gradient in Brazil's semiarid Caatinga region

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ABSTRACT

The effects of anthropogenic disturbances on the structure and function of seasonally dry tropical forests (SDTFs) are mediated both by disturbance type and intensity, and the environmental conditions, including climate. In a climate change scenario, the interaction between low water availability and chronic anthropogenic disturbances may potentially have drastic consequences on SDTF dynamics and the ecosystem services it provides. We hypothesized that in these naturally stressed systems, the effects of human disturbance are strongest in the most water-stressed sites, implying a worrisome synergism between harsh abiotic conditions (lower precipitation, higher temperature) and human effects, ultimately resulting in a less developed community structure and reduced diversity. By integrating information for the adult and regenerating vegetation strata in 21 sites distributed across climatic, plant cover and disturbance gradients in the Caatinga (seasonally dry tropical) forest of Brazil, we analyzed the individual and combined effects of chronic anthropogenic disturbances, as measured through the Cattle Pressure Index and the Human Activity Index, along climate gradients, on community structure, diversity, and composition. As expected, the Cattle Pressure Index was negatively related to diameter at ground level, basal area, and diversity in the adult stratum, whereas for the regenerating stratum, litter stock had the strongest effect on vegetation structure. Climatic factors were poor predictors of the variation in the three

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ABSTRACT

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true diversity measures (0D , 1D , and 2D) of the regenerating community. Our results revealed important cross-site differences in both vegetation strata over a small range of climatic variation. We conclude that the most stressful conditions associated with a very low precipitation regime enhance the effects on chronic disturbances, gradually driving SDTF communities into new alternative states and rendering these environments more prone to desertification.

Keywords: Adult community • Cattle Pressure Index • Human Activity Index • Rainfall gradient • Regenerating community • Seasonally dry tropical forest

1. Introduction

Plant community assembly is the integrated result of the complex interplay among multiple abiotic and biotic factors operating at different spatio-temporal scales (HilleRisLambers et al., 2012; Vellend and Agrawal, 2010). Although this topic has long been central in ecological research, the implications of anthropogenic activities, including chronic disturbances, have only recently gained recognition as drivers of plant community assembly, both locally and globally (Ellis, 2015). Chronic disturbances consist in the continuous removal of relatively small amounts of biomass from ecosystems through different human activities, including the extraction of firewood and timber, and medicinal, ornamental, and fodder plants (Martorell and Peters, 2005; Singh, 1998). Collectively, these activities cause subtle, albeit constant, alterations with long-term effects in the plant community (Arnan et al., 2018; Jara-Guerrero et al., 2021). Depending on their intensity, anthropogenic chronic disturbances can strongly affect community composition and diversity (Eldridge et al., 2018; Jara-Guerrero et al., 2021; Marinho et al., 2016; Ribeiro et al., 2015; Schulz et al., 2019; Travers et al., 2018), the structure of both the adult and regenerating vegetation components (Derroire et al., 2016; Marinho et al., 2016; Schulz et al., 2016), and the functional and phylogenetic diversity of plant communities (Martínez-Blancas et al., 2018; Ribeiro et al., 2016; Zorger et al., 2019), ultimately putting the provision of ecosystem services at risk (Antongiovanni et al., 2020; Díaz et al., 2011).

The effects of chronic anthropogenic disturbances (CAD) are particularly worrisome in seasonally dry tropical forests (SDTF), mainly because of their potential to intensify climate

change effects on plant communities (Arroyo-Rodríguez et al., 2017; Chaturvedi et al., 2017). Northeastern Brazil hosts the most extensive SDTF of the New World (Pennington et al., 2018). Locally known as Caatinga, this vegetation type covers over 912,500 km²; it hosts a flora with a high level of endemism and a large variety of plant communities, whose geographic distribution is mainly determined by climatic, relief, geology, and edaphic heterogeneity (Apgaua et al., 2015; Moro et al., 2015; Silva et al., 2017). Despite representing a substantial proportion of Brazil's territory (ca. 11 %), for a long time the Caatinga was largely overlooked in the academic and political spheres (Miles 2006), which has delayed our understanding of this ecosystem's dynamics when compared to others such as the Amazon rain forest, for example (Santos et al., 2011).

The Caatinga is also characterized by complex and long-standing relationships between humans and biodiversity (Melo, 2017). Nearly 27 million people inhabit the region, which results in strong anthropogenic disturbance gradients, with a variable distribution in time and space (Antongiovanni et al., 2020). Cattle is the main anthropogenic disturbance type in the region (Silva et al., 2017). Domesticated herbivores like goats, oxen, and sheep graze freely on the native vegetation at rates apparently beyond viable limits (Silva et al., 2017). These herbivores have a broad generalist diet as they feed upon the leaves, fruits, barks, and cladodes of adult plants, as well as on seedlings and saplings (Cunha et al., 2022). Consequently, cattle reduce the accumulation of soil litter, affect plant community composition, structure, and diversity, inhibit plant regeneration, and promote desertification due to increased soil exposure (Almeida et al., 2020; Marinho et al., 2016; Schulz et al., 2016). The complexity of chronic anthropogenic disturbances remains insufficiently explored (Antongiovanni et al., 2020; Arnan et al., 2018); notwithstanding important efforts made in this direction, little is known about the isolated effects of different disturbance metrics in dry ecosystems (Jara-Guerrero et al., 2021; Marinho et al., 2016). In the Caatinga, such complexity of disturbances interacts in unclear ways with the variety and heterogeneity of socioecological systems (Pedrosa et al., 2022; Silva et al., 2018). For instance, human populations in different environmental contexts may differ in their species selection strategies (Cunha et al., 2020; Naah and Guuroh, 2017).

Precipitation and soil water retention play relevant roles in this ecosystem, acting as strong environmental filters and driving plant community distribution within the Caatinga realm (Moro et al., 2015; Silva and Souza, 2018). Habitats with higher water availability are associated with higher plant biomass, a variable exerting a strong influence on recruitment

success in SDTF, as seedlings rarely become established on exposed soil (Chazdon, 2003; Marinho et al., 2016). Hence, the semi-arid plant systems are highly vulnerable to climate change (Werneck et al., 2011) and its potential synergy with human activities (Arroyo-Rodríguez et al., 2020; Gibb et al. 2015; Rito et al. 2017a). Hence, the interplay of the high environmental heterogeneity of the Caatinga with the different forms of anthropic disturbances requires sustained efforts from scientists to advance our understanding of the dynamics and functioning of this highly threatened ecosystem.

In this study, we addressed the following questions: (1) How do chronic anthropogenic disturbances affect the structure, composition, and diversity of the adult and the regenerating vegetation components (strata) along climatic gradients in the Caatinga?, and (2) How do climatic variables interact with anthropogenic disturbances? We hypothesized that in these naturally stressed systems the effects of human disturbance are stronger in drier places than in wetter places, implying a worrying synergism between harsh abiotic conditions (lower precipitation, higher temperature) and human effects. Water availability in this forest does not only depend on precipitation but also on temperature, as this variable is involved in water loss from the system through evapotranspiration. Accordingly, we expected the effect of CAD to be stronger in areas of lower precipitation, higher temperature, or higher evapotranspiration. We predicted that under drier conditions the adult community would be less developed, i.e. that it would comprise smaller trees, and host lower diversity; similarly, the regenerating community should also have a low diversity and have a less developed structure under these conditions. Conversely, we predicted that forests occurring in areas with higher water availability (either higher precipitation or lower temperature and lower evapotranspiration) should display a better regenerating capacity and maintain higher species diversity, even when subjected to human disturbance. We also predicted that in drier areas, anthropic disturbances would enhance tree death in the forest canopy.

2. Material and methods

2.1. Study area

The study was conducted in the Cariri Velhas region, Paraíba state, NE Brazil (Fig. 1). Landscapes in the Cariri Velhas (ca. 11,225 km²) are typical of the semi-arid Brazilian region and stand out for having the lowest precipitation regime in the country, so that it is recognized

as a desertification core (Tomasella et al., 2018). Regional climate is hot, semi-arid (Bsw h type; Álvares et al., 2013), and markedly seasonal, with a short rainy period of ca. four months (March to July) and a longer dry season (August to February); evaporation rates are high (< 0.65). Superimposed to the seasonal variation, precipitation also shows large interannual variability, with total annual precipitation ranging from 300 to 700 mm. Elevation ranges from 350 to 1,100 m a.s.l. at the summits of mountains and granitoid *inselbergs* (Moro et al., 2016; Souza et al., 2020). Soils are mainly hyper-eutrophic, gravelly, shallow, and incipient, with nine soil classes recorded for the region (Santos et al., 2011). The dominant vegetation type is the seasonally dry tropical forest, known in Brazil as Caatinga. The flora of these communities embodies numerous species having morpho-physiological adaptations in response to water availability (Queiroz et al., 2017; Fernandes et al., 2020). Relatively short trees dominate the Caatinga, mostly attaining heights of ca. 10 m, along with numerous species of shrubs, cacti, and bromeliads (Fernandes et al., 2020).

Chronic anthropogenic disturbance has been widespread in the Cariri Velhas since the 16th century when the European colonizers conquered the Brazilian North-East and learned to use Caatinga as a source of fodder for the newly introduced cattle. At present, agriculture and extensive cattle ranching are the main economic activities, both having strong effects on vegetation attributes, apparently overrunning the system's carrying capacity (Marinho et al., 2016).

2.2. Selection of study sites and environmental gradients

To assess vegetation attributes along environmental and disturbance gradients, we selected 21 sites based on satellite imagery for the 2017 dry season (accessed through the SA Explorer search tool of the United States Geological Survey), precipitation maps, and field trips. The selected sites represent a range of vegetation cover development (as determined by the Normalized Difference Vegetation Index or NDVI), precipitation, and anthropic disturbances (Fig. 1), but were restricted to a 354–751 m asl elevational range. Sites difficult to access or where the vegetation had been recently removed were discarded and new ones were selected. NDVI is a proxy of photosynthetic activity intensity and thus provides a rough assessment of vegetation condition (Rouse et al., 1974); in the study region, NDVI ranged from 0.13 to 0.73 (Appendix A).

At each site, one 1000-m² (50 × 20 m) plot was established to sample the adult stratum of the vegetation. These plots were subdivided into 100 m² (10 × 10 m) subplots to sample the regenerating stratum. A minimum distance of 1 km between circular buffers (2-km radius; 12.56 km² in area) around the plots was set to maximize between-site independence.

2.3. Evaluation of vegetation attributes

Vegetation sampling was conducted in the rainy season of 2019. To examine the effects of environmental gradients on current vegetation structure while gaining insight into future vegetation development, we distinguished two vegetation components, namely the adult and the regenerating strata. The adult stratum comprised all individual trees and shrubs occurring in the 50 × 20 m plot that were alive or dead (but still standing) with heights ≥ 1 m and stem diameter at ground level (DGL) ≥ 3 cm (Rodal et al., 2008). In turn, the regenerating stratum included those individuals with DGL between 0.5 and 2.9 cm and heights < 1 m. This latter stratum was sampled in four randomly selected 10 × 10 m (100-m²) subplots at each site (Rodal et al., 1992). Tree height was measured with a 12-m tall, graduated pole; the heights of a few trees taller than this pole were estimated visually.

Whenever possible, species identities were recorded in the field; otherwise, we collected samples for inspection by trained botanists or comparison with vouchers deposited at the Manuel de Arruda Câmara Herbarium (HACAM; Universidade Estadual da Paraíba). Species nomenclature and synonymy were verified in the Flora do Brasil data bank. Family classification of species followed APG IV (The Angiosperm Phylogeny Group et al., 2016).

2.4. Predictor variables

We assessed two groups of predictor variables. The first group represented environmental variables and included mean annual precipitation (MAP) and mean annual temperature (MAT), two fundamental variables considered broad indicators of climatic conditions (Hijmans et al., 2005), along with mean annual evapotranspiration (MAE) and Köppen's Aridity Index (AI), calculated as (Quan et al., 2013):

$$AI = MAP / (MAT + 33).$$

We used the *raster* and *sp* packages (Hijmans, 2021; Pebesma and Bivand, 2005) to manipulate raster data within R. Litter stock (LS) was an additional environmental variable

measured in each plot by collecting all necromass accumulated on the ground in five 1-m² (1 × 1 m) subplots randomly distributed in the plot. Litter samples were dried in a forced-air oven at 65 °C for 24 h and weighed in a precision balance to obtain litter dry matter (Appendix A). The mean annual precipitation of each plot ranged from 379 a 690 mm year.

The second group of predictive variables represented human-related factors and included nine metrics considered good indicators of chronic human disturbance in dry environments (Martorell and Peters, 2005), including SDTF (Arnan et al., 2018; Appendix A). Four of these were assessed directly in the plots, whereas the remaining five were derived from the geographical context or extracted from satellite images in the circular buffer around each plot. Based on these metrics, we assessed two main disturbance types to calculate two chronic anthropogenic disturbance indices, one related to cattle pressure and the other to human activity.

Cattle pressure is related to grazing intensity by domesticated herbivores (cattle, horses, donkeys, goats, and sheep) at the sites. To assess this disturbance, at each site we tallied all droppings of cattle (cows), equines (horses and donkeys), and goats/sheep in the 10 × 10 m plots (Appendix A). We then collected, dried, and weighed three dropping samples for each of these groups to calculate mean weight by plot and to estimate total biomass of bovine, equine, and goats/sheep feces by plot. In addition, soil compaction was assessed by measuring bulk density on undeformed soil samples collected from the surface soil horizon and the horizon immediately underneath, through the volumetric ring and balloon methods (Teixeira et al., 2017).

The second disturbance type is related to human activity and is expressed as signs of land use left behind by people. These were recorded as indirect measures from the geographical context based on satellite imagery. First, to assess proximity to human habitation we recorded the distance from each plot's center to the closest house and the closest urban center (city or town). In addition, housing density (number of houses per unit area) and land use were evaluated in buffers around the plots. Land use was defined as the percent cover of the land devoted to agriculture, pasture, residential areas, and human infrastructure; land use and trail density were visually estimated in a smaller 100-m radius buffer.

We integrated the different variables used to measure each disturbance type into a single, comprehensive metric by calculating two indices: the Cattle Pressure Index (CPI) and the Human Activity Index (HAI). Both CPI and HAI were calculated using the following general formula (Legendre and Legendre, 1998; Blüthgen et al., 2012):

$$I_j = \frac{1}{n} \sum_{i=1}^n \frac{y_{i,j} - y_{i,\min}}{y_{i,\max} - y_{i,\min}} \times 100$$

where I_j represents either CPI or HAI in plot j , $y_{i,j}$ is the observed value for the disturbance metric i in plot j , $y_{i,\min}$ is the minimum observed value for the disturbance metric i across all plots, $y_{i,\max}$ is the maximum observed value for the disturbance metric i across all plots, and n is the number of individual disturbance metrics included in the respective index (CPI or HAI).

2.5. Response variables

At each site, we estimated the following response variables for both the adult and regenerating strata: mean diameter at ground level (DGL; cm), mean tree height (H; m), mean tree height of the tallest trees in the canopy, i.e., those included in the uppermost decile at each site (HD10; m or cm, depending on the stratum), basal area (BA; m²), abundance of individuals (Abu), and true diversity measures (Hill numbers, see below). In addition, for the adult stratum we also calculated the fraction of standing dead individuals (FDI).

Community diversity measures were based on cover-based estimators (Chao and Jost, 2012). Sample coverage was very high (> 95%) in all sites, both for the adult and regenerating strata, indicating an adequate sampling effort and confirming that diversity estimations were not biased due to between-plot differences in sample coverage (Chao and Jost, 2012; Chao et al., 2014). To assess diversity in the two above-defined vegetation strata we calculated true diversity values (Hill numbers) by integrating rarefaction and extrapolation curves. We used Hill numbers of three orders (q): $q = 0$ diversity (0D , species richness), which is insensitive to species abundances and thus gives disproportionate weight to rare species; $q = 1$ diversity (1D , exponential Shannon entropy), which is roughly interpreted as the number of common species, as it weights species according to the relative abundances in the community; and $q = 2$ diversity (2D , inverse Simpson concentration), which favors the most abundant species (Jost, 2006; 2007). True diversity values were calculated with the iNEXT package (Hsieh et al., 2016). Community homogeneity was assessed through the Evenness Factor (EF), which represents the proportion of dominant species in the community (Jost, 2010):

$$EF = {}^2D/{}^0D$$

2.6 Assessing the effects of predictor variables on response variables

To assess the effects of predictor variables on response variables from the adult and regenerating strata, we constructed generalized linear models (GLM) with the *glm* function. In addition to the four climatic variables described above, models for the regenerating community also included litter stock (LS) as predictor variable. We used Gaussian and Gamma distributions for the positive, continuous variables (DGL, HD10, BA, ¹D, ²D), with the identity and log link functions for the former distribution, and the log link function for the latter. For count variables (Abundance and ⁰D) a negative binomial distribution was used (Crawley, 2007).

For each response variable, we fitted models that included one, two or three predictor variables (factors), in addition to a null model (i.e., excluding all predictor variables), with the following general structure:

$$y = f(\beta_0)$$

$$y = f(\beta_0 + \beta_1 \cdot x_1)$$

$$y = f(\beta_0 + \beta_1 \cdot x_1 + \beta_2 \cdot x_2)$$

$$y = f(\beta_0 + \beta_1 \cdot x_1 + \beta_2 \cdot x_2 + \beta_3 \cdot x_1 \cdot x_2)$$

$$y = f(\beta_0 + \beta_1 \cdot x_1 + \beta_2 \cdot x_2 + \beta_3 \cdot x_3)$$

where y is any of the 10 response variables of the adult stratum and nine of the regenerating stratum, x_i is any of the six predictor variables of the adult stratum or seven of the regenerating stratum, and f is the link function associated with each response variable. By considering all possible combinations of predictor variables, we fitted 51 models for each response variable for the adult stratum, and 78 models for each response variable for the regenerating stratum (the number of fitted models differed between strata due to the inclusion of the extra predictor [LS] for the regenerating stratum). We assessed multicollinearity among predictor variables with the variance inflation factor (VIF) with the *car* package (Fox and Weisberg, 2019); all variables included in our models had $VIF < 4$, suggesting independence between them (Bondell and Reich, 2008). Next, we performed a model selection procedure based on the sample corrected Akaike Information Criterion (AICc); models with $\Delta AICc < 2$ were considered equally supported and capable of explaining the variation of the response variable. To assess the importance of factors in describing each response variable, we performed a relative importance variable analysis using the AICc values, following Burnham & Anderson (2002). One plot was

considered an outlier because of its abnormally high temperature record and thus was removed from the modeling.

To identify the influence of predictor variables (both environmental and anthropic) on species composition of the adult and regenerating strata, we performed a Non-Metric Multidimensional Scaling (NMDS) ordination with the metaMDS function in the *vegan* package (Oksanen et al., 2015). We first tested for significant differences in vegetation composition based on a Bray-Curtis dissimilarity matrix (Anderson, 2001; McCune et al., 2002). Species abundances were square root-transformed to minimize the influence of the most abundant ones (Kindt and Coe, 2005). Then, we fitted the environmental variables on the NMDS ordination axes with the *envfit* function. This function estimated the strength and direction of the maximum relation between NMDS configuration and the environmental variables, and thus allowed us to test the effects of predictor variables on species composition. The linear fit of the variables on the NMDS axes was calculated, and its significance ($\alpha = 0.05$) assessed with a randomization test based on 999 permutations.

All analyses were performed in R version 3.6.0 (R Core Team, 2019).

3. Results

3.1. Community structure and diversity

For the adult stratum, we recorded a total of 4,501 individuals in the 21 sites (mean \pm SD by site, 214.3 ± 68.7 individuals) representing 71 species (19 with unknown specific identities and thus treated as morphospecies) and 21 families. This figure for total abundance includes 1,285 dead individuals (Table 1, Appendix C). The most speciose families were Fabaceae (16 species) and Euphorbiaceae (10 species) (Appendix C). Observed per plot species richness (0D) ranged between 5 and 35 species (Table 1). This implies an almost 7-fold difference between the least diverse site (Site 7), located centrally in the study area, and the most diverse one (Site 13), located in its western part. Other vegetation attributes also showed large cross-site variation: stem density ranged from 132 to 451 plants per 1000 m², and maximum and mean diameters ranged from 27.4 to 87.9 cm, and between 6.8 to 15.4 cm, respectively. In turn, mean height ranged from 1.8 (site 7) to 7 m (site 17) (Table 1); likewise, canopy height (HD10) was very heterogeneous, ranging from 3 m at Site 6 to the largest

recorded value of 12.7 m at Site 5.

The total number of individuals recorded in the regenerating stratum was 1,648, among which 40 species and 18 families were represented (Appendix C). Again, Fabaceae (9 species) and Euphorbiaceae (7) were the most speciose families. There were enormous cross-site differences in the abundance of individuals in this stratum, with values ranging from 10 (Site 5) to 264 (Site 20) individuals per site. Finally, ranges for mean height and diameter at ground level were 37.9 to 72.3 cm and 0.7 to 1.9 cm, respectively (Table 1).

3.1. Environmental factors affecting adult community structure and diversity

A total of 51 models were fitted to explain the variation of the different attributes of the adult community. After identifying those models with the best fit ($\Delta\text{AIC} < 2$), we selected one model for each response variable ($\Delta\text{AIC} = 0$; Appendix B). The Cattle Pressure Index (CPI), mean annual precipitation (MAP), mean annual temperature (MAT), mean annual evapotranspiration (MAE), aridity index (AI) and their interactions were important predictors of all response variables along the environmental gradients examined (Appendix B). Overall, the selected models pointed to a key role of impacts caused by livestock on the adult community structure and diversity, as indicated by the inclusion of CPI in most selected models (Figs. 2 and 3).

The variation of the adult stratum structure was most strongly influenced by the individual effects of the predictor variables, except for the fraction of standing dead individuals which was also best described by a single model containing MAT, AI, and MAT: AI (Fig. 3; Appendix B). The abundance of live individuals and basal area were best described by a single model that only included CPI as predictor (Fig. 2).

Community height (measured both through HD10 and mean height) was not only best described by MAP, MAE, and AI, but also influenced by CPI and HAI. By contrast, mean diameter at ground level was not affected by any predictor variable. As expected, the three true diversity measures (0D , 1D , and 2D) were strongly affected by CPI (Fig. 2; Appendix B). Between-predictor interactions were only included in the best models for 0D and 1D . In turn, MAP, MAT, MAE, and AI were included in the best models for 0D and 1D (Fig. 2) with significant effects, suggesting an important role of climate on diversity variation. In turn, 2D was influenced by HAI along with MAT, according to the best model for this variable.

3.2. Environmental factors affecting regenerating community structure and diversity

We compared the 78 models that were fitted to explain the variability of the regenerating stratum. Overall, litter stock was the factor having the strongest effect (Fig. 2), particularly for response variables related to vegetation structure, as it was the most frequently included in the best supported models (Appendix B). The importance of CPI seems to be shared between the adult and the regenerating communities, since this factor was present in most selected models (Fig. 4 and Appendix B).

Interestingly, climatic predictors had both positive and negative effects, and therefore they had a weak power to explain the variation in the three true diversity measures (0D , 1D , and 2D) of the regenerating community (Appendix B). Regarding community structure, climatic variables were infrequently included in the best models, suggesting for them a weak influence on the regenerating community (Fig. 2).

3.3. Species composition

The NMDS ordination revealed a gradual change in community composition across the study region. Site distribution on the ordination space revealed the segregation of areas with low (left) and high (right) precipitation (Fig. 5). Among the predictor variables analyzed for the adult stratum, CPI was significantly related with species composition ($R^2 = 0.29$; $P = 0.048$) (Table 2), while mean annual precipitation presented a trend, albeit non-significant ($R^2 = 0.25$; $P = 0.077$) (Table 2).

Croton blanchetianus, *Cenostigma pyramidale*, *Aspidosperma pyriformium* and *Mimosa ophthalmocentra* were the most abundant species, mainly in the most impacted sites, and they were also widely distributed along the environmental gradients. By contrast, *Cnidocolus obtusifolius*, *Erythroxylum* sp., *Maytenus rigida*, *Sebastiania macrocarpa*, *Senegalia bahiensis*, *Senegalia* sp., *Schinopsis brasiliensis*, and *Ziziphus joazeiro* were only recorded in sites with the highest precipitation, and all of them had a low abundance.

Community composition of the regenerating stratum was significantly related with CPI ($R^2 = 0.55$; $P = 0.003$) and litter stock ($R^2 = 0.54$; $P = 0.001$) (Fig. 5, Table 2). *Tacinga palmadora*, *Aspidosperma pyriformium*, *Cenostigma pyramidale*, *Jatropha mollissima* and *M. ophthalmocentra* were the most abundant species in the regenerating stratum (Appendix C). *T.*

palmadora and *A. pyrifolium* also stood out because of their very large mean diameters.

4. Discussion

Here, we report an integrated analysis of the adult and regenerating components of Caatinga plant communities subjected to natural and anthropic influences. The floristic, structural, and environmental information examined provides answers that will help fill the gaps in our understanding of the effects of chronic anthropic disturbance (CAD) on plant community dynamics in semi-arid environments. Our results attest to the magnitude of vegetation variability displayed within a small range of climatic variation and CAD intensities. Along with other factors, CAD idiosyncrasies determine the heterogeneity of present scenarios in the Caatinga. Most importantly, our results provide evidence for the capacity of low precipitation to intensify the effects of CAD on the SDTF vegetation of the Cariri Velhas region, which may potentially drive plant communities into alternative stable states. By focusing on the regenerating stratum, we were able to make predictions about how human impacts, in synergy with low precipitation, may be reflected in the structuring of future plant communities in this region.

4.1. Responses to anthropic disturbances in community structuring

Our results demonstrated the large variation in vegetation attributes along the examined gradients (Table 1), particularly of abundance of individuals and 0D (species richness), both of which had higher values in areas with the lowest CPI (Fig. 3a, f). GLMs showed the existence of direct and indirect effects of CPI and HAI, the two human disturbance indices, on vegetation attributes (Appendix B). Such relationships provide evidence for the role of CAD in the Caatinga which, like other SDTFs, shows a dynamic response to different disturbance types (Álvarez-Yépiz et al., 2008; Arroyo-Rodríguez et al., 2017; Romero-Duque et al., 2007), derived from the heterogeneity of ecosystem properties, particularly the climatic regime.

SDTFs have been historically transformed into human-modified landscapes (Castillo et al., 2005; Ellis, 2015), with chronic anthropogenic disturbance being the fundamental driver of these changes, particularly due to their intensity and frequency, effectively retarding ecosystem recovery (Ellis et al., 2010; Singh, 1998). Feces biomass, a component of CPI, is related to the

direct effects of grazing on plant communities. Large, domesticated herbivores may exert a top-down control that may potentially drive plant communities along two pathways. On the one hand, the action of these herbivores may result in higher community diversity due to the consumption of dominant species and the resulting release of space that becomes available to colonization by other species (Belsky, 1987; Olff and Ritchie, 1998; Osborne et al., 2018). However, herbivores may also reduce the size of the flora in a given environment if they affect mainly rare species or the most sensitive taxa to disturbance (Ford et al., 2018; Hobbs and Huenneke, 1992; Rutherford et al., 2014). This latter situation is common in low productivity environments (Marinho et al., 2016; Travers et al., 2018), as is the case of the Cariri Velhas.

Domestic herbivore pressure in the Caatinga began in the 16th century with the European colonization. The recognition of the foraging potential of native vegetation set the basis for the rapid expansion of cattle ranching in Brazil's semi-arid region (Hemming, 1978). At present, goats (*Capra hircus*) are the most abundant domesticated herbivores in the Caatinga (IBGE, 2017), and their generalist diet virtually spares no plant organ or individual of any age, size, and growth form. Goats browsing is considered the main agent of the alteration of plant community dynamics in the Caatinga, with effects ranging from plant cover reduction to the simplification of community taxonomic and functional diversity (Marinho et al., 2016; Sfair et al., 2018). Ultimately, these factors may potentially drive the socio-ecological systems of the Caatinga to their collapse.

The consequences of grazing are not limited to the damage to individual plants; domesticated herbivores can also cause community-level changes, including litter stock and species composition (Almeida et al., 2020). Litter stock plays a key role in micro-habitat formation by reducing evapotranspiration while increasing organic matter and nutrient contents, and soil water retention, all of which are fundamental factors for seed germination and seedling establishment (Mejía-Domínguez et al., 2012), particularly in semi-arid ecosystems (Bradford et al., 2015; Sobrinho et al., 2016). In addition, soil bulk density, also included in CPI, increases because of the constant trampling of herds (van der Sande et al., 2023). Overgrazing decreases macropore volume and connectivity, reducing soil water infiltration and storage (Minasny and McBratney, 2018). The edaphic stress due to lower water availability negatively affects the establishment of new individuals (Marinho et al., 2016), leading to a low-diversity community, and contributing to the number of dead adults, a variable that was positively related to CPI in this study.

The analysis of the Human Activity Index (HAI), whose metrics comprise various activities of the local human populations (Martorell and Peters, 2005), showed that the areas closest to the largest human settlements and undergoing a more intensive land use are related to an increasing number of dead individuals and a decreasing community height (HD10; Appendix B). These results suggest that the historical pressures to which plant communities have been subjected have caused a reduction in plant stature and a species composition characterized by taxa resistant to the joint action of CPI and HAI.

Throughout their evolution, humans have established complex interactions with biodiversity, which are fundamental for understanding biodiversity patterns worldwide (Ellis, 2015). In our study area, the extraction of plants from nature for fodder, fuel, construction materials, and even for veterinary purposes is common (Cunha et al., 2022; Pedrosa et al., 2022). This activity has different consequences depending on plant life history and the plant part extracted (Ticktin, 2004). For example, the removal of trees for timber generally selects bigger individuals and implies their total or partial cutting, interfering with ecological processes at different levels. This suggests the occurrence of short-term changes in the vertical and horizontal structural patterns in the community, with more individuals concentrating in the smallest size classes, a reduction in species richness, and an increase in plant mortality in the most impacted areas. These effects potentially impact both the adult and regenerating strata, since moving through the forest to access adult plants damages young ones, affecting their growth and survival (Schulz et al., 2016). A number of other negative processes may take place later, including the reduction in plant reproductive potential and the concomitant limitation in seed availability (Chaturvedi et al., 2017; Driscoll and Freudenberger, 2000; Luna-Nieves et al., 2019). In addition to alterations in soil cover, changes in carbon stocks and the physical and chemical composition of soil may occur due to the burning of plant remains (leaves and small branches left after wood removal), a common practice in human-modified landscapes (Fraga and Salcedo, 2004; Schulz et al., 2016; Specht et al., 2015). All these processes directly affect both adult vegetation dynamics and the regeneration process (Burgos and Maass, 2004).

4.2. The synergism between climate and anthropic disturbances on community assembly

Precipitation is an effective environmental filter with positive effects on species diversity in both the adult and the regenerating strata (Fig. 5), and on vegetation development

(mean community height and diameter). Our results show that sites with low precipitation and higher temperature represent more stressful scenarios, with particularly strong effects on tree death (as indicated by the proportion of dead standing individuals). Admittedly, the high number of dead trees recorded in our study may also be the outcome of the prolonged drought that affected the region between 2011 and 2017. Extreme droughts have been shown to have profound effects on the structure and composition of semi-arid ecosystems (Allen et al., 2010; Chaturvedi et al., 2017; Powers et al., 2020; Pérez-Navarro et al., 2021). Moreover, the combination of a low water regime, high temperatures, and strong chronic anthropogenic disturbances induces physiological stress in plants through changes in the soil, vegetation cover and microclimate, especially in drought-sensitive species. This is the case of plants with shallow root systems (Anderegg et al., 2019; Suresh et al., 2010), exemplified in the Cariri Velhas by *Croton blanchetianus* and *Mimosa ophthalmocentra*, the two species that concentrated the largest numbers of dead individuals and possess this morphological characteristic.

The hydrological regime is largely responsible for regulating the growth and the spatial distribution of plant species, as well as for determining species richness in different ecosystems (Pineiro et al., 2017; Silva and Souza, 2018). Most climatic projections forecast for the Caatinga increased temperatures and intense droughts in the coming decades (IPCC, 2014); these climatic scenarios will likely bring about community structure modifications and altered species distributions, mainly through increased tree mortality due to a higher frequency and duration of extreme droughts (Suresh et al., 2010; Powers et al., 2020). Our results point to the overarching role of MAP and MAT in driving the different responses of plant communities to CAD (Appendix B).

We provide evidence of a broad separation between two distinctly different environments according to floristic composition, associated with differences in the precipitation regime, both for the adult and the regenerating strata (Fig. 5). The most humid areas sustained higher species richness, as well as many species typical of more humid environments within the Caatinga (Ramos et al., 2020). In tropical regions, the distribution of plant species is associated with climatic gradients (Boucher-Lalonde et al., 2013; Currie et al., 2004); while water stress-intolerant species occur preferably in wetter areas and hence display more restricted ranges, species more tolerant to extreme conditions are more widely distributed along these climatic gradients (Esquivel-Muelbert et al., 2017).

Despite the preeminent role of climatic variables as drivers of species distributions in

arid and semiarid ecosystems (Currie et al., 2004; Moro et al., 2015), CAD can also act as an environmental filter by selecting species with certain functional characteristics in response to disturbances. For example, increased livestock pressure affects community diversity because it favors species with physical defenses and secondary compounds that deter herbivory (Olf and Ritchie, 1998) and higher resistance to hydric stress. Similarly, human pressure (selective logging, for example) favors species of low wood density and small stature, as these are not highly valued wood sources (Pedrosa et al., 2022). In our study, *Cenostigma pyramidale*, *Aspidosperma pyriformium*, *Croton blanchetianus*, *Jatropha mollissima*, *Mimosa ophthalmocentra*, *Tacinga palmadora* and *Croton heliotropiifolius* accounted for 74.5% of total abundance in the adult stratum, and for 83.2% in the regenerating stratum. Interestingly, all these species are successful resprouters, an important ecological strategy in the recovery of disturbed SDTF (Kennard et al., 2002; Vieira and Scariot, 2006).

Our results show that areas with low rainfall and severe anthropogenic disturbance are associated with lower species richness and greater dominance, a likely result of uneven competitive and colonizing abilities of species under such unfavorable conditions, confirming the negative synergy between disturbance, low rainfall, and low productivity (Gibb et al., 2015; Rito et al., 2017a). Among those species exerting a strong dominance in these stressing environments are *Croton heliotropiifolius* and *C. blanchetianus*, both of which are amply distributed along the studied gradients. These rather shrubby species seldom attain tree stature and thus have small diameters; their high resprouting ability after selective logging or herbivore pressure contributes to cross-site diameter homogeneity.

In the most anthropized areas, *Aspidosperma pyriformium* dominates the regenerating stratum. Several characteristics seem to confer adaptive advantages to this species, making it very drought-resistant and ensuring its persistence in these stressful environments: high lignin and phenolic compound leaf contents (Almeida et al., 2005), which makes it unpalatable to domestic herbivores, a robust root system, and thick epicuticular wax on its leaves (Medeiros et al., 2017). Also, this species exhibits some degree of dwarfism when impacted, a phenomenon commonly observed in plants from arid ecosystems (Díaz et al., 2007; Ren et al., 2017).

The low environmental heterogeneity resulting from disturbances leads to biotic homogenization, affecting diversity by filtering the most sensitive species to environmental changes (Rito et al., 2017b; Ramos et al., 2020). Our results for the adult stratum showed that

Anadenanthera colubrina, *Bauhinia cheilantha*, *Cochlospermum vitifolium*, *Commiphora leptophloeos*, *Myracrodruon urundeuva*, *Pseudobombax marginatum* and *Schinopsis brasiliensis* are restricted to the most humid and less impacted environments. By contrast, populations of disturbance-tolerant species, such as *Aspidosperma pyriformium*, *Cenostigma pyramidale*, *Croton blanchetianus*, *C. heliotropiifolius*, and *Jatropha mollissima* are more successful. In addition to being generalists in disturbed environments of the Caatinga, these species benefit from the edaphic conditions prevailing in them, such as high base saturation (Tabarelli et al., 2012; Ramos et al., 2020).

Our results also provide evidence for novel changes in community composition patterns. For example, *Croton blanchetianus* is a common Caatinga species with high densities reported in numerous studies, especially in impacted areas. To our surprise, we found that this species is also noticeable for a large number of dead individuals in the study sites, suggesting a potential deviation from the compositional pattern repeatedly reported for the Caatinga (Andrade et al., 2005; Araujo et al., 2010; Barbosa et al., 2007; Cabral et al., 2013; Ribeiro et al., 2015; Rodal et al., 2008), and opens the floor to speculations as to which species will occupy this prominent place in the future. Based on our results, we point to *Cenostigma pyramidale* and *Aspidosperma pyriformium* as potential future dominants in these Caatinga communities, provided the persistence of current trends in future decades.

Finally, drier and more impacted areas frequently host communities characterized by lower densities of individuals. For the Cariri Velhas, it is worth noting that three species alone accounted for over 50% of all individuals in the regenerating stratum, which delineates a worrisome scenario regarding this ecosystem's resilience (Poorter et al., 2021). These areas have lower primary productivity, lower plant species diversity, and a larger proportion of exposed soil (Tomasella et al., 2018). The increased soil exposure in areas with reduced vegetation cover renders these environments more susceptible to severe erosion and may be indicative of incipient desertification (Marinho et al., 2016).

5. Conclusions

This study offers new insights into key aspects of community assembly along climatic gradients in plant communities of a tropical semi-arid region affected by a diverse array of chronic anthropogenic disturbances. Communities in areas receiving little rain and subjected to

high chronic anthropogenic disturbance emerge as the most vulnerable ones. This situation is mainly due to grazing by domesticated herbivores, whose effects range from reduced vegetation cover and litter stock, to increased soil compaction, ultimately altering the course of natural regeneration and affecting ecosystem resilience. Only species tolerant to large herbivore grazing and human pressure are likely to persist in this environment in the future, potentially leading to the creation of communities with homogeneous species composition in both their adult and regenerating strata, with a concomitant reduction in regional diversity.

Considering that climate change projections forecast reduced rainfall regimes in semiarid regions, future research should focus on areas with low precipitation and high levels of anthropogenic disturbance as research models of climate change scenarios. The combination of low vegetation cover, low rainfall, and high anthropogenic disturbance makes these environments prone to desertification, and therefore restoration and rehabilitation efforts in these regions are urgently needed.

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References

Allen, C.D., Macalady, A.K., Chenchouni, H., Bachelet, D., McDowell, N., Vennetier, M., Kitzberger, T., Rigling, A., Breshears, D.D., Hogg, E.H.T., Gonzalez, P., Fensham, R., Zhang, Z., Castro, J., Demidova, N., Lim, J.H., Allard, G., Running, S.W., Semerci, A., Cobb, N., 2010. A global overview of drought and heat-induced tree mortality reveals emerging climate change risks for forests. *For. Ecol. Manag.* 259, 660–684. <https://doi.org/10.1016/j.foreco.2009.09.001>

- Almeida, C.F.C.B.R., De Lima E Silva, T.C., De Amorim, E.L.C., Maia, M.B.D.S., De Albuquerque, U.P., 2005. Life strategy and chemical composition as predictors of the selection of medicinal plants from the caatinga (Northeast Brazil). *J. Arid Environ.* 62, 127–142. <https://doi.org/10.1016/j.jaridenv.2004.09.020>
- Almeida, H.A., Ramos, M.B., Diniz, F.C., Lopes, S. de F., 2020. What role does elevational variation play in determining the stock and composition of litter? *Floresta e Ambient.* 27, e20180196. <https://doi.org/10.1590/2179-8087.019618>
- Álvares, C.A., Stape, J.L., Sentelhas, P.C., de Moraes Gonçalves, J.L., Sparovek, G., 2013. Köppen's climate classification map for Brazil. *Meteorol. Zeitschrift* 711–728. <https://doi.org/10.1127/0941-2948/2013/0507>
- Álvarez-Yépez, J.C., Martínez-Yrizar, A., Búrquez, A., Lindquist, C., 2008. Variation in vegetation structure and soil properties related to land use history of old-growth and secondary tropical dry forests in northwestern Mexico. *For. Ecol. Manag.* 256, 355–366. <https://doi.org/10.1016/J.FORECO.2008.04.049>
- Anderson, M.J., 2001. Permutation tests for univariate or multivariate analysis of variance and regression. *Can. J. Fish. Aquat. Sci.* 58, 626–639. <https://doi.org/10.1139/cjfas-58-3-626>
- Anderegg, W.R.L., Anderegg, L.D.L., Kerr, K.L., Trugman, A.T., 2019. Widespread drought-induced tree mortality at dry range edges indicates that climate stress exceeds species' compensating mechanisms. *Glob. Change Biol.* 25, 3793–3802. <https://doi.org/10.1111/gcb.14771>
- Andrade, L.A., Pereira, I.M., Leite, U.T., Barbosa, M.R.V, 2005. Análise da cobertura de duas fitofisionomias de caatinga, com diferentes históricos de uso, no município de São João do Cariri, estado da Paraíba. *Cerne* 11, 253–262.
- Antongiovanni, M., Venticinque, E.M., Matsumoto, M., Fonseca, C.R., 2020. Chronic anthropogenic disturbance on Caatinga dry forest fragments. *J. Appl. Ecol.* 57, 2064–2074. <https://doi.org/10.1111/1365-2664.13686>
- Apgaua, D.M.G., Pereira, D.G.S., Santos, R.M., Menino, G.C.O., Pires, G.G., Fontes, M.A.L., Tng, D.Y.P., 2015. Floristic variation within seasonally dry tropical forests of the Caatinga Biogeographic Domain, Brazil, and its conservation implications. *Int. For. Rev.* 17, 33–44. <https://doi.org/10.1505/146554815815834840>
- Araujo, K.D., Parente, H.N., Éder-Silva, É., Ramalho, C.I., Dantas, R.T., de Andrade, A.P., da Silva, D.S., 2010. Levantamento florístico do estrato arbustivo-arbóreo em áreas

contíguas de Caatinga no Cariri Paraibano. *Rev. Caatinga* 23, 63–70.

- Arnan, X., Arcoverde, G.B., Pie, M.R., Ribeiro-Neto, J.D., Leal, I.R., 2018. Increased anthropogenic disturbance and aridity reduce phylogenetic and functional diversity of ant communities in Caatinga dry forest. *Sci. Total Environ.* 631–632, 429–438. <https://doi.org/10.1016/j.scitotenv.2018.03.037>
- Arroyo-Rodríguez, V., Fahrig, L., Tabarelli, M., Watling, J.I., Tischendorf, L., Benchimol, M., Cazetta, E., Faria, D., Leal, I.R., Melo, F.P.L., Morante-Filho, J.C., Santos, B.A., Arasa-Gisbert, R., Arce-Peña, N., Cervantes-López, M.J., Cudney-Valenzuela, S., Galán-Acedo, C., San-José, M., Vieira, I.C.G., Slik, J.W.F., Nowakowski, A.J., Tschardtke, T., 2020. Designing optimal human-modified landscapes for forest biodiversity conservation. *Ecol. Lett.* 23, 1404–1420. <https://doi.org/10.1111/ele.13535>
- Arroyo-Rodríguez, V., Melo, F.P.L., Martínez-Ramos, M., Bongers, F., Chazdon, R.L., Meave, J.A., Norden, N., Santos, B.A., Leal, I.R., Tabarelli, M., 2017. Multiple successional pathways in human-modified tropical landscapes: new insights from forest succession, forest fragmentation and landscape ecology research. *Biol. Rev.* 92, 326–340. <https://doi.org/10.1111/brv.12231>
- Barbosa, M.R.V., Lima, I.B., Lima, J.R., Cunha, J.P., Agra, M.F., Thomas, W.W., 2007. Vegetação e flora no Cariri Paraibano. *Oecol. Bras.* 11, 313–322. <https://doi.org/10.4257/oeco.2007.1103.01>
- Belsky, A.J., 1987. Revegetation of natural and human-caused disturbances in the Serengeti National Park, Tanzania. *Vegetatio* 70, 51–60. <https://doi.org/10.1007/BF00040758>
- Blüthgen, N., Dormann, C.F., Prati, D., Klaus, V.H., Kleinebecker, T., Hölzel, N., Alt, F., Boch, S., Gockel, S., Hemp, A., Müller, J., Nieschulze, J., Renner, S., Schöning, I., Schumacher, U., Socher, S.A., Wells, K., Birkhofer, K., Buscot, F., Oelmann, Y., Rothenwöhler, C., Scherber, C., Tschardtke, T., Weiner, C.N., Fischer, M., Kalko, E.K.V., Linsenmair, K.E., Schulze, E.-D., Weisser, W.W., 2012. A quantitative index of land-use intensity in grasslands: Integrating mowing, grazing and fertilization. *Basic Appl. Ecol.* 13, 207–220. <https://doi.org/10.1016/j.baae.2012.04.001>
- Bondell, H.D., Reich, B.J., 2008. Simultaneous regression shrinkage, variable selection, and supervised clustering of predictors with OSCAR. *Biometrics* 64, 115–123. <https://doi.org/10.1111/j.1541-0420.2007.00843.x>
- Boucher-Lalonde, V., Kerr, J.T., Currie, D.J., 2013. Does climate limit species richness by

- limiting individual species' ranges? *Proc. Royal Soc. B Biol. Sci.* 281, 20132695. <https://doi.org/10.1098/rspb.2013.2695>
- Bradford, M.A., Berg, B., Maynard, D.S., Wieder, W.R., Wood, S.A., 2015. Understanding the dominant controls on litter decomposition. *J. Ecol.* 104, 229–238. <https://doi.org/10.1111/1365-2745.12507>
- Burgos, A., Maass, J.M., 2004. Vegetation change associated with land-use in tropical dry forest areas of Western Mexico. *Agric. Ecosyst. Environ.* 104, 475–481. <https://doi.org/10.1016/j.agee.2004.01.038>
- Burnham, K. P., and Anderson, D. R. (2002) *Model Selection and Multimodel Inference: a practical information-theoretic approach*. Second edition. Springer: New York.
- Cabral, G.A.L., Sampaio, E.V.S.B., Almeida-Cortez, J.S., 2013. Estrutura espacial e biomassa da parte aérea em diferentes estádios sucessionais de caatinga, em Santa Terezinha, Paraíba. *Rev. Bras. Geog. Fís.* 6, 566–574. <https://doi.org/10.5935/1984-2295.20130057>
- Cámara-Leret, R., Faurby, S., Macía, M.J., Balslev, H., Gödel, B., Svenning, J.C., Kissling, W.D., Rønsted, N., Saslis-Lagoudakis, C.H., 2017. Fundamental species traits explain provisioning services of tropical American palms. *Nat. Plants* 3, 1–7. <https://doi.org/10.1038/nplants.2016.220>
- Castillo, A., Magaña, A., Pujadas, A., Martínez, L., Godínez, C., 2005. Understanding the interaction of rural people with ecosystems: A case study in a tropical dry forest of Mexico. *Ecosystems* 8, 630–643. <https://doi.org/10.1007/S10021-005-0127-1>
- Chao, A., Jost, L., 2012. Coverage-based rarefaction and extrapolation: standardizing samples by completeness rather than size, *Ecology* 93, 2533–2547. <https://doi.org/10.1890/11-1952.1>
- Chao, A., Gotelli, N.J., Hsieh, T.C., Sander, E.L., Ma, K.H., Colwell, R.K., Ellison, A.M., 2014. Rarefaction and extrapolation with Hill numbers: A framework for sampling and estimation in species diversity studies. *Ecol. Monogr.* 84, 45–67. <https://doi.org/10.1890/13-0133.1>
- Chaturvedi, R.K., Raghubanshi, A.S., Tomlinson, K.W., Singh, J.S., 2017. Impacts of human disturbance in tropical dry forests increase with soil moisture stress. *J. Veg. Sci.* 28, 997–1007. <https://doi.org/10.1111/jvs.12547>
- Chazdon, R.L., 2003. Tropical forest recovery: legacies of human impact and natural disturbances. *Perspect. Plant Ecol. Evol. Syst.* 6, 51–71 <https://doi.org/10.1078/1433->

8319-00042

- Crawley, M.J., 2007. *The R Book*. John Wiley and Sons.
<https://doi.org/10.1002/9780470515075>
- Cunha, S.S., Ramos, M.B., de Almeida, H.A., Maciel, M.G.R., de Souza, S.M., Pedrosa, K.M., de Faria Lopes, S., 2022. Vegetation cover and seasonality as indicators for selection of forage resources by local agro-pastoralists in the Brazilian semiarid region. *Sci. Rep.* 12, 15174. <https://doi.org/10.1038/s41598-022-18282-w>
- Currie, D.J., Mittelbach, G.G., Cornell, H.V., Field, R., Guégan, J.F., Hawkins, B.A., Kaufman, D.M., Kerr, J.T., Oberdorff, T., O'Brien, E., Turner, J.R.G., 2004. Predictions and tests of climate-based hypotheses of broad-scale variation in taxonomic richness. *Ecol. Lett.* 7, 1121-1134 <https://doi.org/10.1111/j.1461-0248.2004.00671.x>
- Derroire, G., Balvanera, P., Castellanos-Castro, C., Decocq, G., Kennard, D.K., Lebrija-Trejos, E., Leiva, J.A., Odén, P.C., Powers, J.S., Rico-Gray, V., Tigabu, M., Healey, J.R., 2016. Resilience of tropical dry forests – a meta-analysis of changes in species diversity and composition during secondary succession. *Oikos* 125, 1386–1397.
<https://doi.org/10.1111/oik.03229>
- Díaz, S., Lavorel, S., McIntyre, S., Falczuk, V., Casanoves, F., Milchunas, D.G., Skarpe, C., Rusch, G., Sternberg, M., Noy-Meir, I., Landsberg, J., Zhang, W., Clark, H., Campbell, B.D., 2007. Plant trait responses to grazing - A global synthesis. *Glob. Chang. Biol.* 13, 313–341. <https://doi.org/10.1111/j.1365-2486.2006.01288.x>
- Díaz, S., Quétier, F., Cáceres, D.M., Trainor, S.F., Pérez-Harguindeguy, N., Bret-Harte, M.S., Finegan, B., Peña-Claros, M., Poorter, L., 2011. Linking functional diversity and social actor strategies in a framework for interdisciplinary analysis of nature's benefits to society. *Proc. Natl. Acad. Sci. U.S.A.* 108, 895–902.
<https://doi.org/10.1073/pnas.1017993108>
- Driscoll, D., Milkovits, G., Freudenberger, D., 2000. *Impact and Use of Firewood in Australia*. CSIRO Sustainable Ecosystems, Canberra.
<http://hdl.handle.net/102.100.100/206681?index=1>
- Eldridge, D.J., Delgado-Baquerizo, M., Travers, S.K., Val, J., Oliver, I., 2018. Livestock grazing and forest structure regulate the assembly of ecological clusters within plant networks in eastern Australia. *J. Veg. Sci.* 29, 788–797.
<https://doi.org/10.1111/jvs.12665>

- Ellis, E.C., 2015. Ecology in an anthropogenic biosphere. *Ecol. Monogr.* 85, 287–331. <https://doi.org/10.1890/14-2274.1>
- Ellis, E.C., Goldewijk, K.K., Siebert, S., Lightman, D., Ramankutty, N., 2010. Anthropogenic transformation of the biomes, 1700 to 2000. *Glob. Ecol. Biogeog.* 19, 589–606. <https://doi.org/10.1111/j.1466-8238.2010.00540.x>
- Esquivel-Muelbert, A., Baker, T.R., Dexter, K.G., Lewis, S.L., ter Steege, H., Lopez-Gonzalez, G., Monteagudo Mendoza, A., Brienen, R., Feldpausch, T.R., Pitman, N., Alonso, A., van der Heijden, G., Peña-Claros, M., Ahuite, M., Alexiades, M., Álvarez Dávila, E., Murakami, A.A., Arroyo, L., Aulestia, M., Balslev, H., Barroso, J., Boot, R., Cano, A., Chama Moscoso, V., Comiskey, J.A., Cornejo, F., Dallmeier, F., Daly, D.C., Dávila, N., Duivenvoorden, J.F., Duque Montoya, A.J., Erwin, T., Di Fiore, A., Fredericksen, T., Fuentes, A., García-Villacorta, R., Gonzales, T., Guevara Andino, J.E., Honorio Coronado, E.N., Huamantupa-Chuquimaco, I., Jiménez, R.E.M., Killeen, T.J., Malhi, Y., Mendoza, C., Mogollón, H., Jørgensen, P.M., Montero, J.C., Mostacedo, B., Nauray, W., Neill, D., Vargas, P.N., Palacios, S., Palacios Cuenca, W., Pallqui Camacho, N.C., Peacock, J., Phillips, J.F., Pickavance, G., Quesada, C.A., Ramírez-Angulo, H., Restrepo, Z., Reynel Rodriguez, C., Paredes, M.R., Peñuela-Mora, M.C., Sierra, R., Silveira, M., Stevenson, P., Stropp, J., Terborgh, J., Tirado, M., Toledo, M., Torres-Lezama, A., Umaña, M.N., Urrego, L.E., Vasquez Martinez, R., Gamarra, L.V., Vela, C.I.A., Vilanova Torre, E., Vos, V., von Hildebrand, P., Vriesendorp, C., Wang, O., Young, K.R., Zartman, C.E., Phillips, O.L., 2017. Seasonal drought limits tree species across the Neotropics. *Ecography* 40, 618–629. <https://doi.org/10.1111/ecog.01904>
- Fernandes, M.F., Cardoso, D., de Queiroz, L.P., 2020. An updated plant checklist of the Brazilian Caatinga seasonally dry forests and woodlands reveals high species richness and endemism. *J. Arid Environ.* 174, 104079. <https://doi.org/10.1016/j.jaridenv.2019.104079>
- Ford, H., Healey, J.R., Markesteijn, L., Smith, A.R., 2018. How does grazing management influence the functional diversity of oak woodland ecosystems? A plant trait approach. *Agric. Ecosys. Environ.* 258, 154–161. <https://doi.org/10.1016/j.agee.2018.02.025>
- Fox, J., Weisberg, S., 2019. *An R Companion to Applied Regression* (3rd ed.), Sage, Los Angeles.
- Fraga, V.S., Salcedo, I.H., 2004. Declines of organic nutrient pools in tropical semi-arid soils

under subsistence farming. *Soil Sci. Soc. Am. J.* 68, 215–224.
<https://doi.org/10.2136/sssaj2004.2150>

- Gibb, H., Sanders, N.J., Dunn, R.R., Watson, S., Photakis, M., Abril, S., Andersen, A.N., Angulo, E., Armbrecht, I., Arnan, X., Baccaro, F.B., Bishop, T.R., Boulay, R., Castracani, C., Del Toro, I., Delsinne, T., Diaz, M., Donoso, D.A., Enríquez, M.L., Fayle, T.M., Feener, D.H., Fitzpatrick, M.C., Gómez, C., Grasso, D.A., Groc, S., Heterick, B., Hoffmann, B.D., Lach, L., Lattke, J., Leponce, M., Lessard, J.-P., Longino, J., Lucky, A., Majer, J., Menke, S.B., Mezger, D., Mori, A., Munyai, T.C., Paknia, O., Pearce-Duvet, J., Pfeiffer, M., Philpott, S.M., de Souza, J.L.P., Tista, M., Vasconcelos, H.L., Vonshak, M., Parr, C.L., 2015. Climate mediates the effects of disturbance on ant assemblage structure. *Proc. R. Soc. B Biol. Sci.* 282, 20150418
<https://doi.org/10.1098/rspb.2015.0418>
- Hemming, J., 1978. *Red Gold: The Conquest of the Brazilian Indians*. Harvard University Press Cambridge, Mass.
- Hijmans, R.J., 2021. *Geographic Data Analysis and Modeling [R package raster version 3.5-11]*
- Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G., Jarvis, A., 2005. Very high resolution interpolated climate surfaces for global land areas. *Int. J. Climatol.*, 25, 1965–1978.
<https://doi.org/10.1002/joc.1276>
- HilleRisLambers, J., Adler, P.B., Harpole, W.S., Levine, J.M., Mayfield, M.M., 2012. Rethinking community assembly through the lens of coexistence theory. *Annu. Rev. Ecol. Evol. Syst.* 43, 227–248. <https://doi.org/10.1146/annurev-ecolsys-110411-160411>
- Hobbs, R.J., Huenneke, L.F., 1992. Disturbance, diversity, and invasion: implications for conservation. *Conserv. Biol.* 6,324–337. <http://www.jstor.org/stable/2386033>.
- Hsieh, T.C., Ma, K.H., Chao, A., 2016. Introduction to iNEXT (R package): Excerpt from iNEXT User’ s Guide 1–8.
- IBGE [Instituto Brasileiro de Geografia e Estatística]. 2017. Censo Agropecuário. IBGE, Rio de Janeiro. <https://censos.ibge.gov.br/agro/2017/> (accessed 5.28.21).
- IPCC, 2014. *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, IPCC, Geneva.
- Jara-Guerrero, A., González-Sánchez, D., Escudero, A., Espinosa, C.I., 2021. Chronic

- Disturbance in a Tropical Dry Forest: Disentangling direct and indirect pathways behind the loss of plant richness. *Front. For. Glob. Chang.* 4, 146. <https://doi.org/10.3389/FFGC.2021.723985/BIBTEX>
- Jost, L., 2006. Entropy and diversity. *Oikos*, 113, 363–375. <https://doi.org/10.1111/j.2006.0030-1299.14714.x>
- Jost, L., 2007. Partitioning diversity into independent alpha and beta components. *Ecology* 88, 2427–2439. <https://doi.org/10.1890/06-1736.1>
- Jost, L., 2010. The relation between evenness and diversity. *Diversity* 2, 207–232. <https://doi.org/10.3390/d2020207>
- Kennard, D.K., Gould, K., Putz, F.E., Fredericksen, T.S., Morales, F., 2002. Effect of disturbance intensity on regeneration mechanisms in a tropical dry forest. *For. Ecol. Manag.* 162, 197–208. [https://doi.org/10.1016/S0378-1127\(01\)00506-0](https://doi.org/10.1016/S0378-1127(01)00506-0)
- Kindt, R., Coe, R., 2005. *Tree Diversity Analysis: A Manual and Software for Common Statistical Methods for Ecological and Biodiversity Studies*. World Agroforestry Centre, Nairobi.
- Legendre, P., Legendre, L., 1998. *Numerical Ecology*. Elsevier, Amsterdam.
- Luna-Nieves, A.L., Meave, J.A., González, E.J., Cortés-Flores, J., Ibarra-Manríquez, G., 2019. Guiding seed source selection for the production of Tropical Dry Forest trees: *Couleria platyloba* as study model. *For. Ecol. Manag.* 446, 105–114. <https://doi.org/10.1016/j.foreco.2019.05.020>
- Marinho, F.P., Mazzochini, G.G., Manhães, A.P., Weisser, W.W., Ganade, G., 2016. Effects of past and present land use on vegetation cover and regeneration in a tropical dryland forest. *J. Arid Environ.* 132, 26–33. <https://doi.org/10.1016/j.jaridenv.2016.04.006>
- Martínez-Blancas, A., Paz, H., Salazar, G.A., Martorell, C., 2018. Related plant species respond similarly to chronic anthropogenic disturbance: implications for conservation decision-making. *J. Appl. Ecol.* 55, 1860–1870. <https://doi.org/10.1111/1365-2664.13151>
- Martorell, C., Peters, E.M., 2005. The measurement of chronic disturbance and its effects on the threatened cactus *Mammillaria pectinifera*. *Biol. Conserv.* 124, 199–207. <https://doi.org/10.1016/j.biocon.2005.01.025>
- Mejía-Domínguez, N.R., Meave, J.A., Díaz-Ávalos, C., 2012. Spatial structure of the abiotic environment and its association with sapling community structure and dynamics in a cloud forest. *Int. J. Biometeorol.* 56, 305–318. <https://doi.org/10.1007/s00484-011-0434->

- Miles, L., Newton, A.C., DeFries, R.S., Ravilious, C., May, I., Blyth, S., Kapos, V., Gordon, J.E., 2006. A global overview of the conservation status of tropical dry forests. *J. Biogeogr.* 33, 491–505. <https://doi.org/10.1111/J.1365-2699.2005.01424.X>
- Medeiros, C.D., Falcão, H.M., Almeida-Cortez, J., Santos, D.Y.A.C., Oliveira, A.F.M., Santos, M.G., 2017. Leaf epicuticular wax content changes under different rainfall regimes, and its removal affects the leaf chlorophyll content and gas exchanges of *Aspidosperma pyriformium* in a seasonally dry tropical forest. *S. Afr. J. Bot.* 111, 267–274. <https://doi.org/10.1016/j.sajb.2017.03.033>
- Melo, F.P.L., 2017. The socio-ecology of the Caatinga: Understanding how natural resource use shapes an ecosystem. In: Silva J.M.C., Leal I.R., Tabarelli M. (Eds.), *Caatinga: The Largest Tropical Dry Forest Region in South America*. Springer, Cham, pp. 369–382. https://doi.org/10.1007/978-3-319-68339-3_14
- Minasny, B., McBratney, A.B., 2018. Limited effect of organic matter on soil available water capacity. *Eur. J. Soil Sci.* 69, 39–47. <https://doi.org/10.1111/ejss.12475>
- Moro, M.F., Nic Lughadha, E., de Araújo, F.S., Martins, F.R., 2016. A phytogeographical metaanalysis of the semiarid Caatinga domain in Brazil. *Bot. Rev.* 82, 91–148. <https://doi.org/10.1007/s12229-016-9164-z>
- Moro, M.F., Silva, I.A., de Araújo, F.S., Nic Lughadha, E., Meagher, T.R., Martins, F.R., 2015. The role of edaphic environment and climate in structuring phylogenetic pattern in seasonally dry tropical plant communities. *PLoS One* 10, e0119166. <https://doi.org/10.1371/journal.pone.0119166>
- McCune, B., Grace, J.B., Urban, D.L., 2002. *Analysis of Ecological Communities*. MjM Software Design, Glenden Beach, Oregon.
- Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O’Hara, R.B., Simpson, G.L., Solymos, P., Stevens, H., Wagner, H.H., 2015. *Vegan: community ecology package*. R package *vegan*, ver. 2.2-1.
- Olf, H., Ritchie, M.E., 1998. Effects of herbivores on grassland plant diversity. *Trends Ecol. Evol.* 13, 261–265 [https://doi.org/10.1016/S0169-5347\(98\)01364-0](https://doi.org/10.1016/S0169-5347(98)01364-0)
- Osborne, C.P., Charles-Dominique, T., Stevens, N., Bond, W.J., Midgley, G., Lehmann, C.E.R., 2018. Human impacts in African savannas are mediated by plant functional traits. *New Phytol.* 220, 10–24. <https://doi.org/10.1111/nph.15236>

- Pebesma, E., Bivand, R.S., 2005. Classes and Methods for Spatial Data: the sp Package.
- Pedrosa, K.M., Ramos, M.B., Cunha, S.S. da, Maciel, M.G.R., Souza, S.M. de, Soares, H.K. de L., Torre-Cuadros, M. de los Á. La, Lopes, S. de F., 2022. Local ecological knowledge dynamics of farmers in areas which have been chronically disturbed by human actions in the Brazilian Caatinga. *Ethnobot. Res. Appl.* 24, 1–21. <https://doi.org/10.32859/era.24.23.1-21>
- Pennington, R.T., Lehmann, C.E.R., Rowland, L.M., 2018. Tropical savannas and dry forests. *Curr. Biol.* 28, R541–R545. <https://doi.org/10.1016/j.cub.2018.03.014>
- Pérez-Navarro, M., Serra-Diaz, J.M., Svenning, J.C., Esteve-Selma, M.Á., Hernández-Bastida, J., Lloret, F., 2021. Extreme drought reduces climatic disequilibrium in dryland plant communities. *Oikos* 130, 680–690. <https://doi.org/10.1111/OIK.07882>
- Pinheiro, E.A.R., de van Lier, Q.J., Bezerra, A.H.F., 2017. Hydrology of a water-limited forest under climate change scenarios: the case of the Caatinga biome, Brazil. *Forests* 8. <https://doi.org/10.3390/f8030062>
- Poorter, L., Craven, D., Jakovac, C.C., Sande, M.T. van der, Amisshah, L., Bongers, F., Chazdon, R.L., Farrior, C.E., Kambach, S., Meave, J.A., Muñoz, R., Norden, N., Rüger, N., Breugel, M. van, Zambrano, A.M.A., Amani, B., Andrade, J.L., Brancalion, P.H.S., Broadbent, E.N., Foresta, H. de, Dent, D.H., Derroire, G., DeWalt, S.J., Dupuy, J.M., Durán, S.M., Fantini, A.C., Finegan, B., Hernández-Jaramillo, A., Hernández-Stefanoni, J.L., Hietz, P., Junqueira, A.B., N'dja, J.K., Letcher, S.G., Lohbeck, M., López-Camacho, R., Martínez-Ramos, M., Melo, F.P.L., Mora, F., Müller, S.C., N'Guessan, A.E., Oberleitner, F., Ortiz-Malavassi, E., Pérez-García, E.A., Pinho, B.X., Piotto, D., Powers, J.S., Rodríguez-Buriticá, S., Rozendaal, D.M.A., Ruíz, J., Tabarelli, M., Teixeira, H.M., Sampaio, E.V. de S.B., Wal, H. van der, Villa, P.M., Fernandes, G.W., Santos, B.A., Aguilar-Cano, J., Almeida-Cortez, J.S. de, Alvarez-Davila, E., Arreola-Villa, F., Balvanera, P., Becknell, J.M., Cabral, G.A.L., Castellanos-Castro, C., Jong, B.H.J. de, Nieto, J.E., Espírito-Santo, M.M., Fandino, M.C., García, H., García-Villalobos, D., Hall, J.S., Idárraga, A., Jiménez-Montoya, J., Kennard, D., Marín-Spiotta, E., Mesquita, R., Nunes, Y.R.F., Ochoa-Gaona, S., Peña-Claros, M., Pérez-Cárdenas, N., Rodríguez-Velázquez, J., Villanueva, L.S., Schwartz, N.B., Steininger, M.K., Veloso, M.D.M., Vester, H.F.M., Vieira, I.C.G., Williamson, G.B., Zanini, K., Hérault, B., 2021. Multidimensional tropical forest recovery. *Science* 374, 1370–1376.

<https://doi.org/10.1126/science.abh3629>

- Powers, J.S., Vargas, G.G., Brodribb, T.J., Schwartz, N.B., Pérez-Aviles, D., Smith-Martin, C.M., Becknell, J.M., Aureli, F., Blanco, R., Calderón-Morales, E., Calvo-Alvarado, J.C., Calvo-Obando, A.J., Chavarría, M.M., Carvajal-Vanegas, D., Jiménez-Rodríguez, C.D., Murillo Chacon, E., Schaffner, C.M., Werden, L.K., Xu, X., Medvigy, D., 2020. A catastrophic tropical drought kills hydraulically vulnerable tree species. *Glob. Chang. Biol.* 26, 3122–3133. <https://doi.org/10.1111/gcb.15037>
- Quan, C., Han, S., Utescher, T., Zhang, C., Liu, Y.-S.C., 2013. Validation of temperature–precipitation based aridity index: Paleoclimatic implications. *Palaeogeogr. Palaeoclimatol. Palaeoecol.* 386, 86–95. <https://doi.org/10.1016/j.palaeo.2013.05.008>
- Queiroz, L.P., Cardoso, D., Fernandes, M.F., Moro, M.F., 2017. Diversity and evolution of flowering plants of the Caatinga Domain, In: Silva J.M.C., Leal I.R., Tabarelli M. (Eds.), *Caatinga: The Largest Tropical Dry Forest Region in South America*. Springer, Cham, pp. 23–63. https://doi.org/10.1007/978-3-319-68339-3_2
- R Core Team, 2019. *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing, Vienna.
- Ramos, M.B., Diniz, F.C., de Almeida, H.A., de Almeida, G.R., Pinto, A.S., Meave, J.A., Lopes, S. de F., 2020. The role of edaphic factors on plant species richness and diversity along altitudinal gradients in the Brazilian semi-arid region. *J. Trop. Ecol.* 36, 199–212. <https://doi.org/10.1017/S0266467420000115>.
- Ren, W., Hu, N., Hou, X., Zhang, J., Guo, H., Liu, Z., Kong, L., Wu, Z., Wang, H., Li, X., 2017. Long-term overgrazing-induced memory decreases photosynthesis of clonal offspring in a perennial grassland plant. *Front. Plant Sci.* 8, 419. <https://doi.org/10.3389/FPLS.2017.00419/BIBTEX>
- Ribeiro, E.M.S., Arroyo-Rodríguez, V., Santos, B.A., Tabarelli, M., Leal, I.R., 2015. Chronic anthropogenic disturbance drives the biological impoverishment of the Brazilian Caatinga vegetation. *J. Appl. Ecol.* 52, 611–620. <https://doi.org/10.1111/1365-2664.12420>
- Ribeiro, E.M.S., Santos, B.A., Arroyo-Rodríguez, V., Tabarelli, M., Souza, G., Leal, I.R., 2016. Phylogenetic impoverishment of plant communities following chronic human disturbances in the Brazilian Caatinga. *Ecology* 97, 1583–1592. <https://doi.org/10.1890/15-1122.1>

- Rito, K.F., Arroyo-Rodríguez, V., Queiroz, R.T., Leal, I.R., Tabarelli, M., 2017a. Precipitation mediates the effect of human disturbance on the Brazilian Caatinga vegetation. *J. Ecol.* 105, 828–838. <https://doi.org/10.1111/1365-2745.12712>
- Rito, K.F., Tabarelli, M., Leal, I.R., 2017b. Euphorbiaceae responses to chronic anthropogenic disturbances in Caatinga vegetation: from species proliferation to biotic homogenization. *Plant Ecol.* 218, 749–759. <https://doi.org/10.1007/s11258-017-0726-x>
- Rodal, M.J.N., Costa, K.C.C., Lins e Silva, A.C.B., 2008. Estrutura da vegetação caducifolia espinhosa (caatinga) de uma área do sertão central de Pernambuco. *Hoehnea* 35, 209–217. <https://doi.org/10.1590/s2236-89062008000200004>
- Rodal, M.J.N., Sampaio, E.V.S.B., Figueiredo, M.A., 1992. Manual sobre métodos de estudo florístico e fitossociológico: Ecossistema Caatinga. Sociedade Botânica do Brasil, Brasília.
- Romero-Duque, L.P., Jaramillo, V.J., Pérez-Jiménez, A., 2007. Structure and diversity of secondary tropical dry forests in Mexico, differing in their prior land-use history. *For. Ecol. Manag.* 253, 38–47. <https://doi.org/10.1016/J.FORECO.2007.07.002>
- Rouse, J.W. Jr., Haas, R.H., Schell, J.A., Deering, D.W., 1974. Monitoring vegetation systems in the Great Plains with ERTS. NASA. Goddard Space Flight Center 3d ERTS-1 Symp., Vol. 1, Sect. A.
- Rutherford, M.C., Powrie, L.W., Husted, L.B., 2014. Herbivore-driven land degradation: Consequences for plant diversity and soil in arid subtropical thicket in South-Eastern Africa. *Land Degrad. Dev.* 25, 541–553. <https://doi.org/10.1002/ldr.2181>
- Santos, J.C., Leal, I.R., Almeida-Cortez, J.S., Fernandes, G.W., Tabarelli, M., 2011. Caatinga: the scientific negligence experienced by a dry tropical forest. *Tropical Conserv. Sci.* 4, 276–286.
- Santos, H.G. dos, Carvalho Júnior, W., Dart, R.O., Áglio, M.L.D., Souza, J.S., Pares, J.G., Fontana, A., Martins, A.L.S., Oliveira, A.P., 2011. O Novo Mapa de Solos do Brasil. Legenda Atualizada. Documentos/Embrapa Solos, 130. Rio de Janeiro.
- Schulz, K., Guschal, M., Kowarik, I., Almeida-Cortez, J.S. de, Sampaio, E.V. de S.B., Cierjacks, A., 2019. Grazing reduces plant species diversity of Caatinga dry forests in northeastern Brazil. *Appl. Veg. Sci.* 22, 348–359. <https://doi.org/10.1111/avsc.12434>
- Schulz, K., Voigt, K., Beusch, C., Almeida-Cortez, J.S., Kowarik, I., Walz, A., Cierjacks, A., 2016. Grazing deteriorates the soil carbon stocks of Caatinga forest ecosystems in Brazil.

- For. Ecol. Manag. 367, 62–70. <https://doi.org/10.1016/j.foreco.2016.02.011>
- Sfair, J.C., de Bello, F., de França, T.Q., Baldauf, C., Tabarelli, M., 2018. Chronic human disturbance affects plant trait distribution in a seasonally dry tropical forest. *Environ. Res. Lett.* 13, 025005 <https://doi.org/10.1088/1748-9326/aa9f5e>
- Silva, A.C., Souza, A.F., 2018. Aridity drives plant biogeographical sub regions in the Caatinga, the largest tropical dry forest and woodland block in South America. *PLoS One* 13, e0196130. <https://doi.org/10.1371/journal.pone.0196130>
- Silva, J.M.C., Barbosa, L.C.F., Leal, I.R., Tabarelli, M., 2017. The Caatinga: understanding the challenges. In: Silva J.M.C., Leal I.R., Tabarelli M. (Eds.), *Caatinga: The Largest Tropical Dry Forest Region in South America*. Springer, Cham, pp. 3–19. https://doi.org/10.1007/978-3-319-68339-3_1
- Singh, S.P., 1998. Chronic disturbance, a principal cause of environmental degradation in developing countries. *Environ. Conserv.* 25, 1–2. <https://doi.org/10.1017/S0376892998000010>
- Sobrinho, M.S., Tabarelli, M., Machado, I.C., Sfair, J.C., Bruna, E.M., Lopes, A.V., 2016. Land use, fallow period and the recovery of a Caatinga forest. *Biotropica* 48, 586–597. <https://doi.org/10.1111/btp.12334>
- Souza, J.J.L.L., de Souza, B.I., Xavier, R.A., Pacheco, A.A., Pessenda, L.C.R., dos Santos Brito, E., 2020. Archaeoanthrosol formation in the Brazilian semiarid. *Catena* 193, 104603. <https://doi.org/10.1016/j.catena.2020.104603>
- Specht, M.J., Pinto, S.R.R., Albuquerque, U.P., Tabarelli, M., Melo, F.P.L., 2015. Burning biodiversity: fuelwood harvesting causes forest degradation in human-dominated tropical landscapes. *Global Ecol. Conserv.* 3, 200–209. <https://doi.org/10.1016/J.GECCO.2014.12.002>
- Suresh, H.S., Dattaraja, H.S., Sukumar, R., 2010. Relationship between annual rainfall and tree mortality in a tropical dry forest: Results of a 19-year study at Mudumalai, southern India. *For. Ecol. Manag.* 259, 762–769. <https://doi.org/10.1016/j.foreco.2009.09.025>
- Tabarelli, M., Peres, C.A., Melo, F.P.L., 2012. The ‘few winners and many losers’ paradigm revisited: emerging prospects for tropical forest biodiversity. *Biol. Conserv.* 155, 136–140. <https://doi.org/10.1016/j.biocon.2012.06.020>
- Teixeira, P.C., Donagemma, G.K., Fontana, A., Teixeira, W.G., (Eds.), 2017. *Manual de métodos de análise de solo*, 3rd ed. Rio de Janeiro: Embrapa Solos.

- The Angiosperm Phylogeny Group, Chase, M.W., Christenhusz, M.J.M., Fay, M.F., Byng, J.W., Judd, W.S. Soltis, D.E., Mabberley, D.J., Sennikov, A.N. Soltis, P.S. Stevens, P.F., 2016. An update of the Angiosperm Phylogeny Group classification for the orders and families of flowering plants: APG IV. *Bot. J. Linn. Soc.* 181, 1–20, <https://doi.org/10.1111/boj.12385>
- Ticktin, T., 2004. The ecological implications of harvesting non-timber forest products. *J. Appl. Ecol.* 41, 11–21. <https://doi.org/10.1111/j.1365-2664.2004.00859.x>
- Tomasella, J., Silva Pinto Vieira, R.M., Barbosa, A.A., Rodriguez, D.A., de Oliveira Santana, M., Sestini, M.F., 2018. Desertification trends in the Northeast of Brazil over the period 2000–2016. *Int. J. Appl. Earth Obs. Geoinf.* 73, 197–206. <https://doi.org/10.1016/j.jag.2018.06.012>
- Travers, S.K., Eldridge, D.J., Dorrrough, J., Val, J., Oliver, I., 2018. Introduced and native herbivores have different effects on plant composition in low productivity ecosystems. *Appl. Veg. Sci.* 21, 45–54. <https://doi.org/10.1111/avsc.12334>
- van der Sande, M.T., Powers, J.S., Kuyper, T.W., Norden, N., Salgado-Negret, B., de Almeida, J.S., Bongers, F., Delgado, D., Dent, D.H., Derroire, G., Espirito Santo, M.M., Dupuy, J.M., Fernandes, G.W., Finegan, B. Gavito, M.E., Hernández-Stefanoni, J.L. Jakovac, C.C., Jones, I.L., Magalhães Veloso, M.D., Meave, J.A., Mora, F., Muñoz, R., Pérez-Cárdenas, N., Piotto, D., Álvarez-Dávila, E., Caceres-Siani, Y., Dalban-Pilon, C., Dourdain, A., Du, D.V., Villalobos, D.G., Nunes, Y.R.F., Sanchez-Azofeifa, A., Poorter, L., 2023. Soil resistance and recovery during Neotropical forest succession. *Philos. Trans. R. Soc. Lond., B, Biol. Sci.* 378, 20210074. [<https://doi.org/10.1098/rstb.2021.0074>]
- Vellend, M., Agrawal, A. 2010. Conceptual synthesis in community ecology. *Q. Rev. Biol.* 85, 183–206. <https://doi.org/10.1086/652373>
- Vieira, D.L.M., Scariot, A., 2006. Principles of natural regeneration of tropical dry forests for restoration. *Restor. Ecol.* 14, 11–20. <https://doi.org/10.1111/J.1526-100X.2006.00100.X>
- Werneck, F.P., Costa, G.C., Colli, G.R., Prado, D.E., Sites, J.W., 2011. Revisiting the historical distribution of Seasonally Dry Tropical Forests: new insights based on palaeodistribution modelling and palynological evidence. *Glob. Ecol. Biogeogr.* 20, 272–288. <https://doi.org/10.1111/j.1466-8238.2010.00596.x>
- Zorger, B.B., Tabarelli, M., de Queiroz, R.T., Rosado, B.H.P., Pinho, B.X., 2019. Functional

organization of woody plant assemblages along precipitation and human disturbance gradients in a seasonally dry tropical forest. *Biotropica* 51, 838–850.
<https://doi.org/10.1111/btp.12721>

Table 1. Statistical summary of per-plot attributes of the adult and regenerating communities in 21 Caatinga sites distributed along environmental and human disturbance gradients in the semi-arid region of Brazil. 0D , 1D and 2D are Hill numbers of order $q = 0, 1,$ and $2,$ respectively, representing diversity measures (effective number of species). NM, not measured

Community variables	Adult stratum		Regenerating stratum	
	Mean (\pm SD)	Range	Mean (\pm SD)	Range
Abundance (ind./plot)	214.3 \pm 68.7	89–332	78.5 \pm 56.1	10–264
Dead individuals (ind./plot)	61.2 \pm 51.9	3–226	NM	NM
Mean height (m ^a or cm ^b)	3.5 \pm 1.3	1.8–7.0	51.6 \pm 8.6	37.9–72.3
TOP 10% height (m ^a or cm ^b)	6.7 \pm 2.6	3.0–12.7	89.2 \pm 7.2	75–98.7
Diameter at ground level (cm)	9.0 \pm 1.9	6.8–15.4	1.3 \pm 0.4	0.7–1.9
Basal area (m ²)	105.0 \pm 54.7	0.42–3.2	1.7 \pm 0.8	54.0–309.4
0D	14.64 \pm 6.6	5–35	9.0 \pm 4.2	3.0–18.6
1D	6.9 \pm 2.4	3.1–12.0	5.2 \pm 2.3	1.9–11.4
2D	5.0 \pm 1.5	2.3–7.6	3.9 \pm 1.6	1.37–8.4
Family richness	7.2 \pm 2.5	4–13	5.2 \pm 2.6	1–33
Evenness factor	0.4 \pm 0.1	0.20–0.6	0.5 \pm 0.2	0.1–0.8

^a for the adult stratum; ^b for the regenerating stratum.

Table 2. Relationships between the scores in the first two NMDS axes (NMDS1 and NMDS2) and predictor variables. R^2 = linear regression fit. Environmental factors: MAP, mean annual precipitation; MAT, mean annual temperature; MAE, mean annual evapotranspiration; AI, aridity index; LS, litter stock. Human factors: HAI, human activity index; CPI, Cattle Pressure Index. Significant P -values are highlighted in bold typeface.

Predictor variables	Adult stratum				Regenerating stratum			
	NMDS1	NMDS2	R^2	P	NMDS1	NMDS2	R^2	P
Environmental predictor variables								
MAP	0.507	0.862	0.018	0.841	0.500	0.866	0.131	0.282
MAT	0.972	0.237	0.263	0.069	0.936	0.353	0.130	0.276
MAE	-0.980	0.199	0.111	0.364	0.554	0.833	0.084	0.454
AI	0.976	0.216	0.254	0.077	0.466	0.885	0.117	0.336
LS	—	—	—	—	0.579	0.815	0.550	0.003
Chronic anthropogenic disturbance variables								
CPI	-0.686	-0.727	0.303	0.049	-0.912	0.409	0.545	0.001
HAI	-0.834	-0.550	0.110	0.349	-0.518	0.855	0.114	0.325

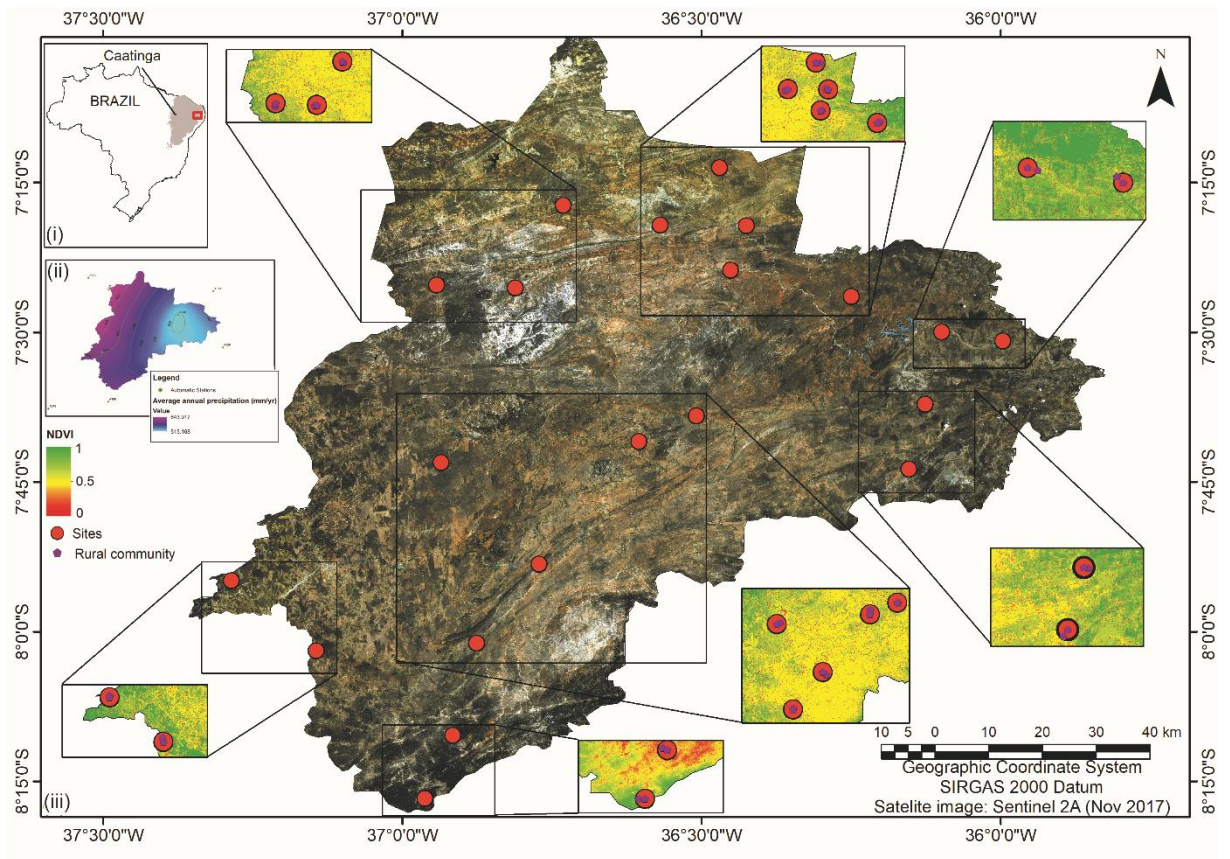


Fig. 1. Map of the Cariri Velhas region, Paraíba state, northeastern Brazil. i. location of the study region in the Caatinga realm. ii. regional distribution of the average annual precipitation. iii: detailed map of the study region; the red circles show the locations of the 21 study sites; insets next to them show the distribution of NDVI used to assess vegetation development in their vicinity, and the presence of rural communities within their buffers.

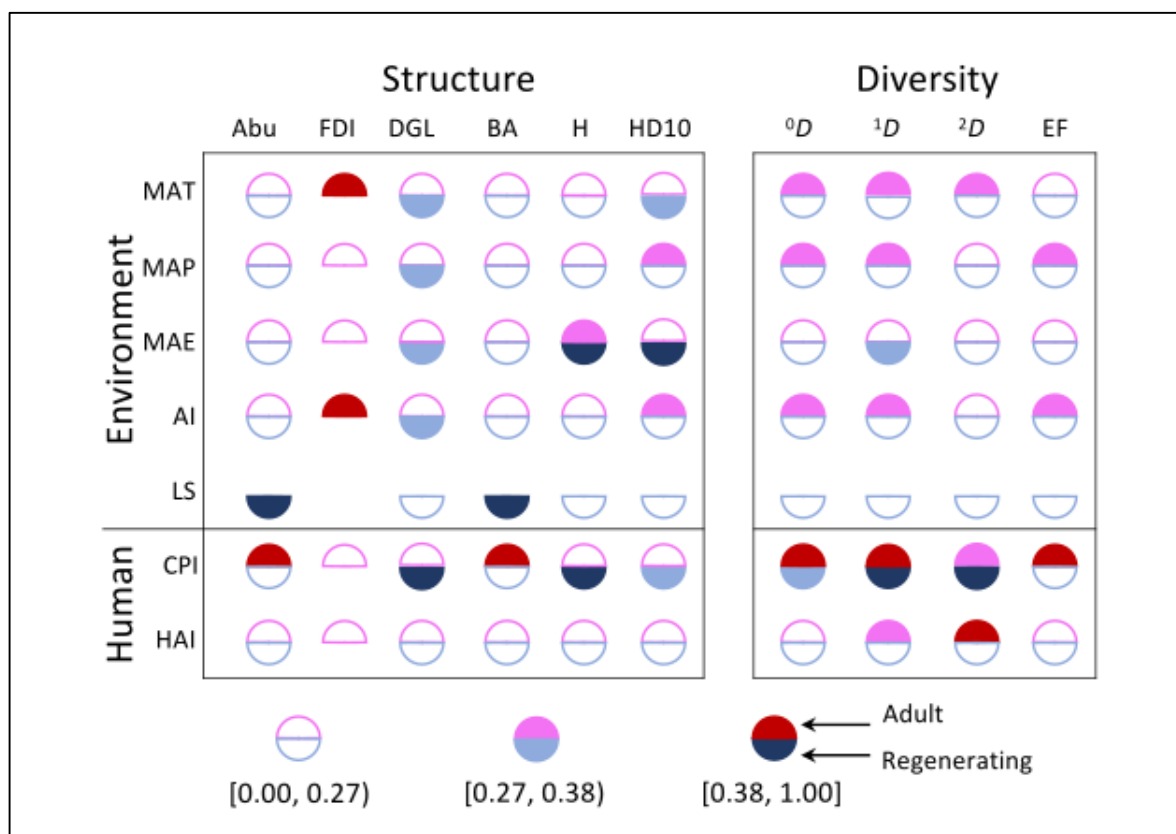


Fig. 2. Graphic representation of the assessment of factor (predictive variable) importance in the generalized linear models constructed to analyze their role in explaining the variation of response variables for the adult and the regenerating strata in Caatinga forest of NE Brazil along climatic and disturbance gradients. Environmental predictive variables: MAT, mean annual temperature; MAP, mean annual precipitation; MAE, mean annual evapotranspiration; AI, aridity index; LS, litter stock; human disturbance variables: CPI, Cattle Pressure Index; HAI, human activity index. Community structure response variables: Abu, abundance; FDI, fraction of death individuals; DGL, diameter at ground level; BA, basal area; H, mean tree height; HD10, mean height of the tallest (uppermost decile) trees; community diversity response variables: ⁰D, ¹D, ²D, true diversity (Hill numbers) of orders q = 0, 1, and 2; EF, evenness factor). The upper halves of the circles correspond to the adult stratum (shades of red) and the lower halves to the regenerating stratum (shades of blue). Increasing color intensity reflects increasing importance, according to thresholds defined in Burnham and Anderson (2002). Empty circles: [0.00, 0.27); light colored circles: [0.27, 0.38); dark-colored circles: [0.38, 1.00]. Only dark colors represent truly important variables.

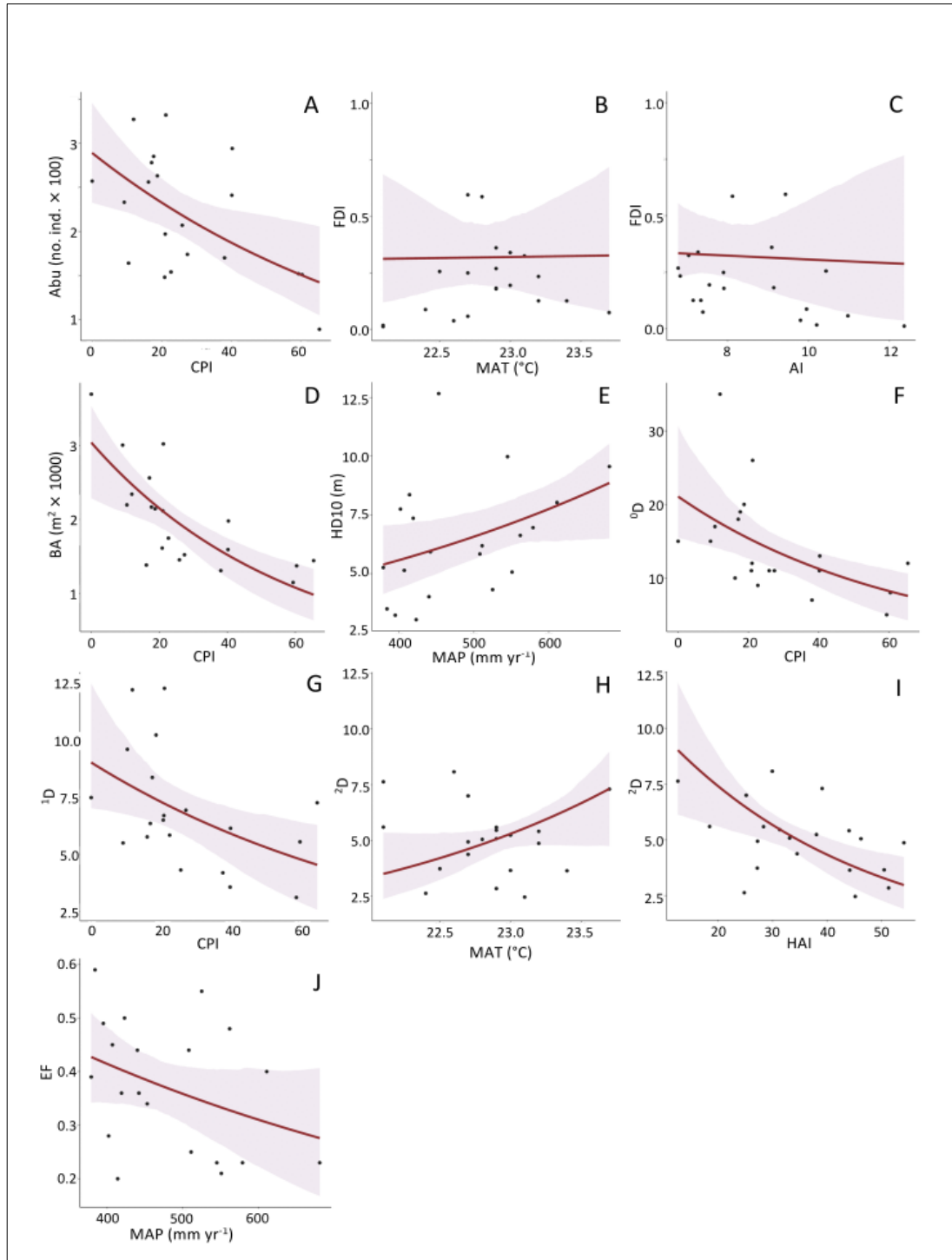


Fig. 3. Model predictions from the best-supported generalized linear models explaining the variation of community structural and diversity response variables as a function of environmental and human disturbance variables for the adult stratum in the Caatinga forest, NE Brazil. Solid red lines represent model predictions and red shading represents the 95% confidence envelope. Black dots represent the data used in fitting the models. Response variables: Abu, stem abundance in the plot (A); FDI, fraction of standing dead trees (B, C); BA, stem abundance in the plot (D); HD10, diameter at 10 m (E); HD, diameter at 10 m (F, G); HD, diameter at 10 m (H); HD, diameter at 10 m (I); EF, edge effect (J).

basal area (D); HD10, mean height of the tallest (uppermost decile) trees in the community (E); 0D, 1D and 2D, true diversity measure through Hill numbers of orders $q = 0, 1$ and 2 , respectively (F-I); EF, evenness factor (J). Environmental predictive variables: AI, aridity index; MAP, mean annual precipitation; MAT, mean annual temperature. Human disturbance predictor variables: HAI, human activity index; CPI Cattle Pressure Index. Response variables examined in the study for which no model is shown were best fit by the null model (no factor effect). Note that for some response variables the best-supported model included more than one predictive variable, in which case one graph is shown for each variable separately.

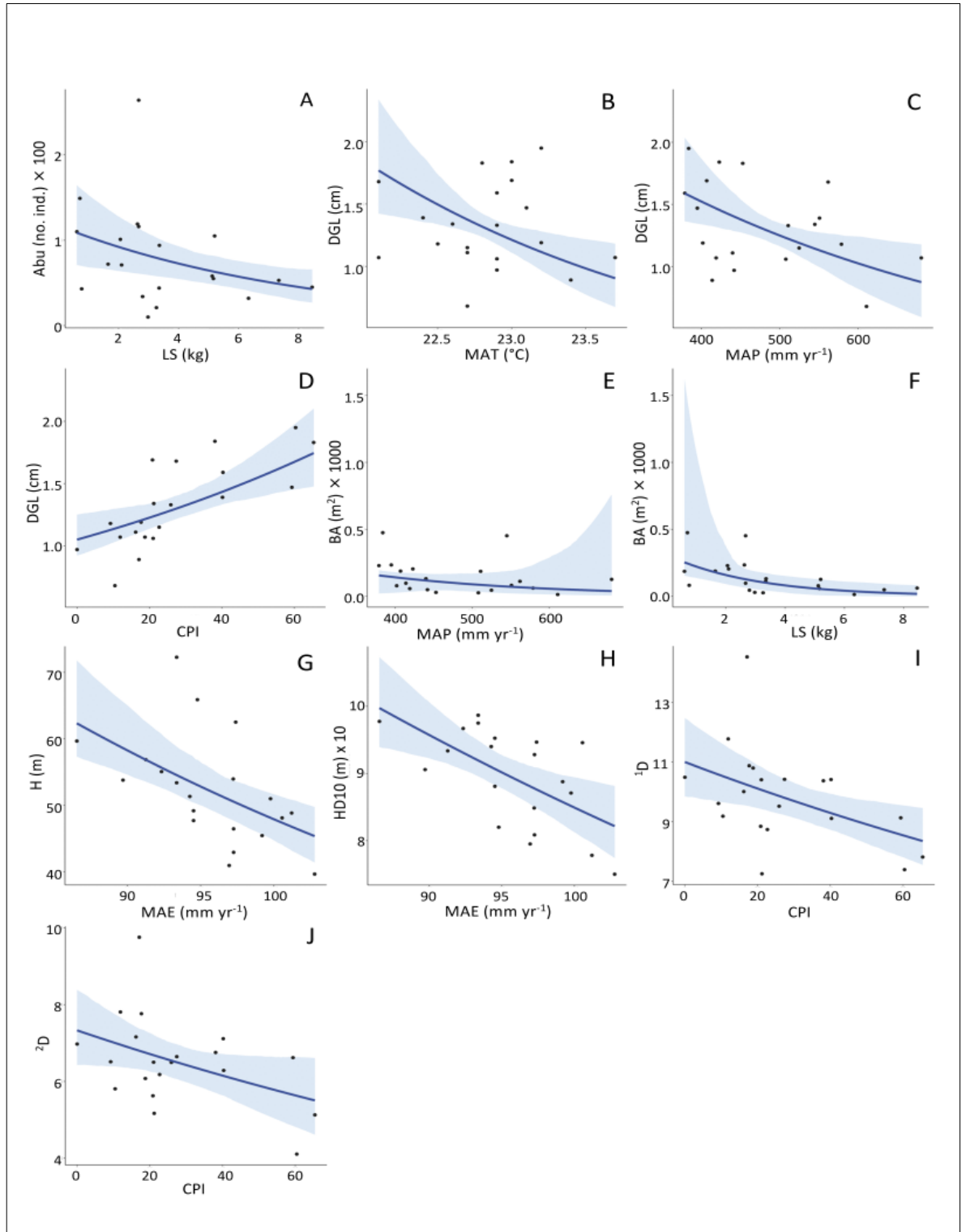


Fig. 4. Model predictions from the best-supported generalized linear models explaining the variation of community structural and diversity response variables as a function of environmental and human disturbance variables for the regenerating stratum in the Caatinga forest, NE Brazil. Solid blue lines represent model predictions for each response variable and

blue shading represents the 95% confidence envelope. Black dots represent the data used for fitting the models. Response variables: Abu, stem abundance in the plot (A); DGL, mean diameter at ground level (B-D); BA, basal area (E, F); H, mean plant height in this stratum (G); HD10, mean height of the tallest (uppermost decile) trees in this stratum (H); 1D and 2D , true diversity measure through Hill numbers of orders $q = 1$ and 2 , respectively (I, J). Environmental predictive variables: LS, litter stock; MAE, mean annual evapotranspiration; MAP, mean annual precipitation; MAT, mean annual temperature. Human disturbance predictor variable: CPI Cattle Pressure Index. Response variables examined in the study for which no model is shown were best fit by the null model (no factor effect). Note that for some response variables the best-supported model included more than one predictive variable, in which case one graph is shown for each variable separately.

Guappira sp.; Jamo, *Jatropha mollissima*; Jari, *Jatropha ribifolia*; Lavi, *Lachesiodendron viridiflorum*; Laca, *Lantana camara*; Life, *Libidibia ferrea*; Luau, *Lutzemburgia auriculata*; Maca, *Manihot carthagenensis*; Miop, *Mimosa ophthalmocentra*; Misp, *Mimosa* sp. 2; Mite, *Mimosa tenuiflora*; Mori, *Monteverdia rigida*; Myur, *Myracrodruon urundeuva*; Pigo, *Pilosocereus gounellei*; Pipa, *Pilosocereus pachycladus*; Pist, *Piptadenia stipulacea*; Pimo, *Pityrocarpa moniliformis*; Psma, *Pseudobombax marginatum*; Pssp, *Psidium* sp.; Ptneu, *Ptilochaeta nudipes*; Sagl, *Sapium glandulosum*; Scbr, *Schinopsis brasiliensis*; Sema, *Sebastiania macrocarpa*, Seba, *Senegalia bahiensis*; Sesp, *Senegalia* sp.; Sesp, *Senna spectabilis*; Siob, *Sideroxylon obtusifolium*; Sptu, *Spondias tuberosa*; Taau, *Tabebuia aurea*; Tapa, *Tacinga palmadora*; Vale, *Varronia leucocephala*; Zijo, *Ziziphus joazeiro*; Mo1 to M19, unidentified morphospecies (see Appendix C for full species names and authorities).

Appendix A. Statistical descriptors of predictor variables analyzed in this study. MAP, mean total annual precipitation; MAT, mean annual temperature; LS, litter stock; CATTLE, cattle droppings biomass; EQUINE, horse and donkey droppings biomass; GOAT, goat and sheep droppings biomass; COMP, soil compaction; PH, proximity to rural property; PUC, proximity to urban center; HD housing density; LUSE, land use; TRAN, trail density. See text for a detailed description of these variables.

Predictor variable	Mean (SD)	Min	Max
<i>Environmental indicators</i>			
MAP	491.48 (95.29)	379.0	690.0
MAT	22.94 (0.61)	22.1	25.0
LS	3.54 (2.15)	0.61	8.45
<i>Cattle pressure indicators</i>			
CATTLE	2514.33 (5287.76)	0.00	22896.00
EQUINE	643.24 (1348.63)	0.00	4873.00
GOAT	752.33 (1173.18)	0.00	3343.00
COMP	1.56 (0.14)	1.16	1.81
<i>Human activity indicators</i>			
PH	519.90 (385.04)	1.00	1552.00
PUC	51.22 (197.95)	3.10	915.00
HD	2.93 (2.83)	0.40	13.30
LUSE	4.38 (2.16)	1.00	9.00
TRAN	1.07 (1.14)	0.00	3.52

Appendix B. Best supported models (i.e., $\Delta\text{AICc} \leq 2$) constructed to assess the effects of environmental and human factors on community attributes of the adult and regenerating strata in the Caatinga forest, NE Brazil. Environmental factors: MAP, mean annual precipitation; MAT, mean annual temperature; MAE, mean annual evapotranspiration; AI, aridity index; LS, litter stock. Human factors: HAI, human activity index; CPI, Cattle Pressure Index. Response variables: Abu, abundance (no. of individuals) per site; FDI, fraction of standing death individuals (adult stratum only); DGL, diameter at ground level; BA, basal area; H, mean height for all trees recorded at the site; HD10, mean height of the tallest trees (uppermost decile) at the site; 0D , 1D , and 2D , community true diversity measures (Hill numbers) for $q = 0$, $q = 1$, and $q = 2$, respectively; EF, evenness factor, calculated as ${}^2D/{}^0D$. The total number of models fitted for each response variable is shown on the second column. The label ‘Null’ indicates that the best supported model did not include any explanatory variable, thus it was the null model.

Response variable	Total no. of models	Model	ΔAICc
<i>Adult stratum</i>			
Abu	48	$5.37 - 0.19 \cdot \text{CPI}$	0
FDI	49	$-0.76 + 0.02 \cdot \text{MAT} - 0.06 \cdot \text{AI} + 0.98 \cdot \text{MAT} \cdot \text{AI}$	0
DGL	48	2.17 (Null)	0
BA	46	$9.85 - 0.31 \cdot \text{CPI}$	0
H	47	1.20 (Null)	0
		$1.19 + 0.09 \cdot \text{MAP}$	0.962
		$1.19 - 0.09 \cdot \text{MAE}$	0.974
		$1.19 + 0.09 \cdot \text{AI}$	1.027
		$1.19 - 0.09 \cdot \text{CPI}$	1.352
		$1.20 - 0.08 \cdot \text{HAI}$	1.410
HD10	44	$1.85 + 0.14 \cdot \text{MAP}$	0
		$1.85 + 0.14 \cdot \text{AI}$	0.072
		1.86 (Null)	0.382
		$1.86 - 0.12 \cdot \text{MAE}$	0.825

⁰ D	48	$2.62 - 0.28 \cdot \text{CPI}$	0
		$2.61 + 0.15 \cdot \text{MAP} - 0.22 \cdot \text{CPI}$	0.180
		$2.61 - 0.22 \cdot \text{CPI} + 0.14 \cdot \text{AI}$	0.231
		$2.48 + 0.21 \cdot \text{MAT} + 0.29 \cdot \text{MAP} - 0.17 \cdot \text{MAT} \cdot \text{MAP}$	0.845
		$2.48 + 0.22 \cdot \text{MAT} + 0.30 \cdot \text{AI} - 0.17 \cdot \text{MAT} \cdot \text{AI}$	0.903
		$2.62 - 0.21 \cdot \text{CPI} - 0.13 \cdot \text{HAI}$	1.288
		$2.62 - 0.27 \cdot \text{CPI} - 0.10 \cdot \text{MAE}$	1.576
		$2.62 + 0.26 \cdot \text{MAT} + 0.43 \cdot \text{AI}$	1.774
		$2.62 + 0.25 \cdot \text{MAT} + 0.42 \cdot \text{MAP}$	1.905
¹ D	45	$1.91 - 0.19 \cdot \text{CPI}$	0
		$1.92 - 0.17 \cdot \text{HAI}$	0.486
		$1.90 + 0.19 \cdot \text{MAT} - 0.34 \cdot \text{HAI}$	0.702
		$1.92 + 0.16 \cdot \text{MAP}$	0.703
		$1.92 + 0.16 \cdot \text{AI}$	0.815
		$1.91 + 0.21 \cdot \text{MAT} + 0.33 \cdot \text{AI}$	1.065
		$1.91 + 0.11 \cdot \text{MAP} - 0.15 \cdot \text{CPI}$	1.142
		$1.91 + 0.20 \cdot \text{MAT} + 0.31 \cdot \text{MAP}$	1.172
		$1.91 - 0.14 \cdot \text{CPI} - 0.12 \cdot \text{HAI}$	1.185
		$1.91 - 0.15 \cdot \text{CPI} + 0.11 \cdot \text{AI}$	1.188
		$1.91 - 0.19 \cdot \text{CPI} - 0.09 \cdot \text{MAE}$	1.660
		$1.82 + 0.20 \cdot \text{MAT} + 0.25 \cdot \text{AI} - 0.10 \cdot \text{MAT} \cdot \text{AI}$	1.963
		$1.83 + 0.19 \cdot \text{MAT} + 0.24 \cdot \text{MAP} - 0.10 \cdot \text{MAT} \cdot \text{MAP}$	1.982
² D	47	$1.59 + 0.18 \cdot \text{MAT} - 0.30 \cdot \text{HAI}$	0
		$1.61 - 0.14 \cdot \text{HAI}$	0.592
		$1.61 - 0.14 \cdot \text{CPI}$	1.750
EF	48	$-1.00 - 0.12 \cdot \text{MAP}$	0
		$-1.00 + 0.11 \cdot \text{CPI}$	0.018
		$-1.00 - 0.12 \cdot \text{AI}$	0.089
		-0.99 (Null)	0.092

		$-1.00 - 0.09 \cdot \text{MAP} + 0.08 \cdot \text{CPI}$	1.631
		$-1.00 + 0.08 \cdot \text{CPI} - 0.09 \cdot \text{AI}$	1.673
		$-1.00 - 0.13 \cdot \text{MAT} - 0.23 \cdot \text{AI}$	1.874
		$-1.00 - 0.13 \cdot \text{MAT} - 0.22 \cdot \text{MAP}$	1.881
		$-0.99 + 0.07 \cdot \text{HAI}$	1.966
		$-0.99 + 0.07 \cdot \text{MAE}$	1.971
<i>Regenerating stratum</i>			
Abu	75	$4.35 - 0.26 \cdot \text{LS}$	0
		4.38 (Null)	0.069
DGL	74	$0.26 - 0.17 \cdot \text{MAT} - 0.17 \cdot \text{MAP} + 0.14 \cdot \text{CPI}$	0
		$0.26 - 0.17 \cdot \text{MAT} + 0.14 \cdot \text{CPI} - 0.17 \cdot \text{AI}$	0.050
		$0.27 + 0.18 \cdot \text{CPI}$	0.164
		$0.26 - 0.11 \cdot \text{MAP} + 0.16 \cdot \text{CPI} - 0.11 \cdot \text{MAE}$	0.372
		$0.26 + 0.16 \cdot \text{CPI} - 0.11 \cdot \text{AI} - 0.11 \cdot \text{MAE}$	0.531
		$0.26 + 0.15 \cdot \text{CPI} - 0.08 \cdot \text{LS}$	1.190
		$0.26 + 0.19 \cdot \text{CPI} - 0.06 \cdot \text{MAE}$	1.209
		$0.26 + 0.15 \cdot \text{CPI} + 0.10 \cdot \text{HAI} - 0.11 \cdot \text{MAE}$	1.436
BA	75	$4.58 - 0.39 \cdot \text{MAP} - 0.73 \cdot \text{LS}$	0
		$4.58 - 0.73 \cdot \text{LS} - 0.40 \cdot \text{AI}$	0.033
		$4.83 - 0.58 \cdot \text{LS}$	0.061
		$4.63 - 0.67 \cdot \text{LS} - 0.04 \cdot \text{AI} + 0.41 \cdot \text{LS} \cdot \text{AI}$	0.147
		$4.63 - 0.04 \cdot \text{MAP} - 0.68 \cdot \text{LS} + 0.40 \cdot \text{MAP} \cdot \text{LS}$	0.182
		$4.69 - 0.04 \cdot \text{MAT} - 0.65 \cdot \text{LS} - 0.52 \cdot \text{MAT} \cdot \text{LS}$	0.640
		$4.71 - 0.61 \cdot \text{LS} - 0.23 \cdot \text{MAE} - 0.55 \cdot \text{LS} \cdot \text{MAE}$	1.045
		$4.72 + 0.16 \cdot \text{CPI} - 0.23 \cdot \text{MAE} + 0.54 \cdot \text{CPI} \cdot \text{MAE}$	1.412
		$4.79 + 0.20 \cdot \text{HAI} - 0.50 \cdot \text{LS}$	1.647
		$4.72 + 0.22 \cdot \text{MAT} - 0.68 \cdot \text{LS}$	1.962
H	74	$3.95 - 0.08 \cdot \text{MAE}$	0

		$3.95 + 0.05 \cdot \text{CPI} - 0.09 \cdot \text{MAE}$	0.089
		$3.95 + 0.05 \cdot \text{HAI} - 0.11 \cdot \text{MAE}$	1.356
HD10	75	$4.49 - 0.05 \cdot \text{MAE}$	0
		$4.49 - 0.02 \cdot \text{CPI} - 0.05 \cdot \text{MAE}$	0.508
		$4.49 + 0.02 \cdot \text{LS} - 0.05 \cdot \text{MAE}$	1.197
		$4.49 - 0.05 \cdot \text{MAT}$	1.573
⁰ D	75	3.49 (Null)	0
		$3.49 - 0.06 \cdot \text{CPI}$	0.872
¹ D	75	$2.29 - 0.08 \cdot \text{CPI}$	0
		$2.28 - 0.08 \cdot \text{CPI} + 0.05 \cdot \text{MAE}$	0.946
		2.29 (Null)	1.758
		$2.28 + 0.04 \cdot \text{MAT} - 0.08 \cdot \text{CPI}$	1.971
² D	75	$1.87 - 0.08 \cdot \text{CPI}$	0
		1.88 (Null)	1.332
EF	75	-0.73 (Null)	0
		$-0.73 + 0.08 \cdot \text{CPI}$	1.804

Appendix C. List of species and families, along with number of individuals, recorded in the adult and the regenerating strata in 21 Caatinga forest sites, semiarid Cariri Velhas region, Paraíba state, Brazil.

Families/Species	Adult stratum	Regenerating stratum
Anacardiaceae		
<i>Myracrodruon urundeuva</i> Allemão	38	4
<i>Schinopsis brasiliensis</i> Engl.	25	6
<i>Spondias tuberosa</i> Arruda	3	0
Apocynaceae		
<i>Aspidosperma pyrifolium</i> Mart.	856	362
Bignoniaceae		
<i>Tabebuia aurea</i> (Silva Manso) Benth. & Hook.f. ex S.Moore	6	1
Bixaceae		
<i>Cochlospermum vitifolium</i> (Willd.) Spreng.	18	0
Boraginaceae		
<i>Varronia leucocephala</i> (Moric.) J.S.Mill.	6	7
Burseraceae		
<i>Commiphora leptophloeos</i> (Mart.) J.B.Gillett	92	34
Cactaceae		
<i>Cereus jamacaru</i> DC.	8	0
<i>Pilosocereus gounellei</i> (F.A.C.Weber) Byles & Rowley	83	11
<i>Pilosocereus pachycladus</i> F.Ritter	32	8
<i>Tacinga palmadora</i> (Britton & Rose) N.P.Taylor & Stuppy	263	384
Capparaceae		
<i>Cynophalla flexuosa</i> (L.) J.Presl	15	10
Celastraceae		
<i>Maytenus rigida</i> Mart.	1	1
Combretaceae		
<i>Combretum leprosum</i> Mart.	2	7

Families/Species	Adult stratum	Regenerating stratum
<i>Combretum</i> sp.	16	0
Erythroxylaceae		
<i>Erythroxylum caatingae</i> Plowman	5	2
Euphorbiaceae		
<i>Cnidoscolus quercifolius</i> Pohl	2	0
<i>Cnidoscolus obtusifolius</i> Pohl ex Baill	26	1
<i>Croton blanchetianus</i> Baill.	536	107
<i>Croton heliotropiifolius</i> Kunth	114	5
<i>Croton</i> sp.	1	0
<i>Jatropha mollissima</i> (Pohl) Baill.	373	144
<i>Jatropha ribifolia</i> (Pohl) Baill.	9	40
<i>Manihot carthagenensis</i> (Jacq.) Müll.Arg.	57	6
<i>Sapium glandulosum</i> (L.) Morong	23	3
<i>Sebastiania macrocarpa</i> Müll.Arg.	6	0
Fabaceae		
<i>Albizia polycephala</i> (Benth.) Killip ex Record	55	3
<i>Anadenanthera peregrina</i> (L.) Speg.	10	0
<i>Anadenanthera colubrina</i> (Vell.) Brenan	50	27
<i>Bauhinia cheilantha</i> (Bong.) Steud	85	22
<i>Cenostigma pyramidale</i> (Tul.) E.Gagnon & G.P.Lewis	1031	250
<i>Chloroleucon foliolosum</i> (Benth.) G.P.Lewis	18	16
<i>Erythrina velutina</i> Willd.	4	0
<i>Lachesiodendron viridiflorum</i> (Kunth) P.G.Ribeiro, L.P.Queiroz	2	0
<i>Libidibia ferrea</i> (Mart. ex Tul.) L.P.Queiroz	8	0
<i>Luetzelburgia auriculata</i> (Allemão) Ducke	2	0
<i>Mimosa ophthalmocentra</i> Mart. ex Benth.	293	124
<i>Mimosa</i> sp.	1	4

Families/Species	Adult stratum	Regenerating stratum
<i>Mimosa tenuiflora</i> (Willd.) Poir.	7	1
<i>Piptadenia retusa</i> P.G.Ribeiro, Seigler & Ebinger	76	1
<i>Pityrocarpa moniliformis</i> (Benth.) Luckow & R.W.Jobson	63	10
<i>Senegalia bahiensis</i> Benth.	8	0
<i>Senegalia</i> sp.	3	0
<i>Senna spectabilis</i> (DC.) H.S.Irwin & Barneby	2	0
Malpighiaceae		
<i>Ptilochaeta nudipes</i> Griseb	1	0
Malvaceae		
<i>Ceiba glaziovii</i> (Kuntze) K.Schum.	1	0
<i>Pseudobombax marginatum</i> (A.St.-Hil., Juss. & Cambess.)	23	6
Myrtaceae		
<i>Psidium</i> sp.	2	0
Nyctaginaceae		
<i>Guapira laxa</i> (Netto) Furlan	2	0
<i>Guapira</i> sp.	3	2
Rhamnaceae		
<i>Ziziphus joazeiro</i> Mart.	4	5
Sapotaceae		
<i>Sideroxylon obtusifolium</i> (Roem. & Schult.) T.D.Penn.	3	8
Solanaceae		
<i>Capiscum caatingae</i> Barboza & Agra	20	3
Verbenaceae		
<i>Lantana camara</i> L.	25	7
Not identified		
Morphospecies I	25	1
Morphospecies II	1	1

Families/Species	Adult stratum	Regenerating stratum
Morphospecies III	7	1
Morphospecies IV	1	0
Morphospecies V	1	0
Morphospecies VI	12	0
Morphospecies VII	1	0
Morphospecies VIII	1	0
Morphospecies VX	14	0
Morphospecies X	2	0
Morphospecies XI	1	0
Morphospecies XII	2	0
Morphospecies XIII	1	0
Morphospecies XIV	4	0
Morphospecies XV	1	0
Morphospecies XVI	3	0
Morphospecies XVII	3	0
Morphospecies XVIII	2	0
Morphospecies XIX	1	13

6. CAPÍTULO V: CONSIDERAÇÕES FINAIS

6.1 Principais conclusões

A Caatinga compreende a maior região semiárida do mundo, possui uma elevada heterogeneidade ambiental, com diversas formações geológicas, uma variedade de tipos de solo, de relevo, alta variabilidade espacial e temporal pluviométrica. Além disso, o histórico de uso por populações humanas que ocorre a milhares de anos, culminou em sistemas socioecológicos complexos e dinâmicos. Esse histórico de uso resultou em variados tipos e intensidades de distúrbios antrópicos, tornando o ambiente ainda mais complexo. Embora algumas informações já tenham sido identificadas, pouco se sabe sobre como todos esses fatores interagem, juntamente com os sistemas sociais e os danos ao ecossistema. Dessa forma, acreditamos que muitas lacunas ainda existem, e são necessários ainda muitos esforços de pesquisa. É necessário compreender os diferentes efeitos de forma combinada e isolada, bem como avaliar as particularidades de cada sub-região da Caatinga e os cenários representados em cada projeto de pesquisa. Nosso estudo foi realizado em uma área submetida a múltiplos distúrbios antrópicos crônicos e totalmente desprovida de qualquer tipo de proteção ambiental. Outro ponto importante é que nossa região de estudo possui a menor precipitação pluviométrica do semiárido brasileiro e compreende um núcleo de desertificação, fazendo com que essa porção específica da Caatinga necessite urgentemente de mais estudos.

A integração de conhecimento ecológico de populações humanas sobre a utilização de espécies vegetais, aos aspectos da ecologia de comunidades vegetais, empregada nessa tese foi fundamental para traçar novas ideias sobre os fatores que afetam o uso e a seleção de espécies de plantas por populações humanas da Caatinga. Assim, a abordagem a partir das características funcionais nos permitiu traçar um perfil funcional das espécies de planta para cada serviço ecossistêmico de interesse dos humanos que vivem nessa região.

Em geral, observamos que os agropastores apontam uma série de importantes características, positivas e negativas, que distingue as necessidades desejadas para cada serviço ecossistêmico. Por exemplo, para o uso forrageiro, os agropastores descrevem características associadas a textura, sabor, cheiro como fundamentais para uso e seleção das espécies pelos herbívoros domésticos. Nossas análises confirmaram que as características funcionais das plantas correspondem a fatores fundamentais na tomada de decisão para seleção de espécies vegetais. A densidade de madeira por exemplo, é uma característica fundamental para seleção de espécies nas categorias combustível e construção, conferindo maior resistência e

durabilidade a madeira. Destacamos ainda que as relações de manejo das plantas podem ter vários impactos econômicos, sociais e ambientais. Por exemplo, a pressão de uso sobre alguns grupos de plantas pode levar a perda de espécies e funções específicas alterando a prestação de serviços ecossistêmicos. Portanto, compreender a complexidade do uso e seleção de espécies pelos agropastores têm um papel significativo na conservação da biodiversidade.

Em relação as comunidades vegetais, entendemos que as condições climáticas da Caatinga podem trazer informações substanciais e previsões para cenários futuros perante as mudanças climática. Utilizamos esses gradientes ambientais para analisar de forma isolada e interativa os efeitos das variáveis bióticas e abióticas sobre a estruturação das comunidades de plantas, e quais efeitos podem reverberar na dinâmica do sistema ecológico e social. A interdependência com a disponibilidade de água, torna os sistemas vegetais de ambientes semiáridos bastante vulneráveis as mudanças climáticas. Há grande preocupação com a resiliência destes sistemas, sobretudo aos efeitos sinérgicos com o crescimento exponencial das atividades antrópicas, visto que os estudos sobre as interações entre essas variáveis são bastante incipientes e ainda obscuros.

Nossos resultados revelam que nas áreas com menor regime hídrico, os impactos antrópicos têm um forte efeito sob a diversidade das comunidades, levando a uma baixa riqueza de espécies, em função da homogeneização biótica, e dificuldade no restabelecimento e recrutamento de indivíduos. Esse cenário pode ocasionar uma desestabilização nos sistemas socioecológicos, uma vez que as populações locais dependem de maneira direta de alguns serviços ecossistêmicos que estão sendo perdidos. Por fim, nos apêndices da presente tese apresentamos o registro de ocorrência de uma espécie endêmica do semiárido brasileiro, cuja ocorrência era desconhecida para o estado da Paraíba, esse achado reforça a ideia de que muitas lacunas ainda existem, e são necessários grandes esforços de pesquisa para conhecermos e compreendermos a biodiversidade da Caatinga.

6.2 Contribuições teóricas e /ou metodológicas da tese

As evidências encontradas neste estudo apresentam contribuições metodológicas para estudos em Ecologia vegetal e Etnobiologia, sobretudo para ambientes áridos e semiáridos. Do ponto de vista teórico, fornecemos novas informações sobre as interações entre seres humanos e os vegetais, e seus respectivos efeitos para a montagem de comunidades vegetais. Acreditamos que o conhecimento ecológico local é uma ferramenta essencial para tomada de

decisões que buscam alcançar a sustentabilidade de um sistema. A integração entre o conhecimento ecológico local e a ecologia de comunidades vegetais nos permitiu traçar um perfil funcional das espécies de plantas utilizadas e dessa forma, identificar as características funcionais mais desejadas, e aquelas que são indesejadas pelos agropastores da região. Demonstramos ainda, os estratos adultos e regenerantes de comunidades vegetais da Caatinga, elucidando os aspectos da estrutura, composição e diversidade de espécies, frente a distintos distúrbios antrópicos crônicos e condições climáticas em uma área de Caatinga desprovida de qualquer tipo de proteção ambiental. Esse contexto revelou resultados críticos para os processos de resiliência e manutenção dos serviços ecossistêmicos desse ambiente.

Com relação as contribuições metodológicas, apresentamos aqui uma estrutura interdisciplinar que nos permitiu analisar as relações entre atividades humanas, serviços ecossistêmicos, características funcionais e as prioridades dos agropastores envolvidos. Do ponto de vista do entendimento da montagem de comunidades vegetais, mensuramos variáveis climáticas, um conjunto variado de métricas de distúrbios antrópicos, e associamos ao acúmulo de serrapilheira para compreender a estrutura, e diversidade de espécies das comunidades vegetais.

6.3 Principais limitações do estudo

Com a pandemia do COVID-19 e conseqüentemente interrupções das expedições a campo, houve atraso no desenvolvimento de algumas partes da pesquisa, como o levantamento dos dados etnobotânicos. Dentro desse contexto, não foi possível a inclusão do número de comunidades locais propostas inicialmente, bem como o retorno a campo para coleta de material vegetal para mensuração de traços funcionais e informações etnobotânicas relacionadas as características funcionais das espécies. Tais problemáticas interferiram diretamente na finalização dos trabalhos elaborados. Além disso, enfrentamos algumas dificuldades para coletar e depositar material botânico no herbário. Embora com a chegada das vacinas houvesse uma diminuição das taxas de transmissão do vírus, existia uma certa resistência das pessoas em participarem das pesquisas. Dentro desse contexto, a pesquisa teve que ser redirecionada, o número de comunidades rurais foi reduzido e a inserção de características funcionais foi compactada.

6.4 Propostas de investigações futuras

Os produtos apresentados na tese sugerem a necessidade de mais trabalhos que investiguem de maneira mais detalhada os efeitos dos distúrbios antrópicos crônicos, especialmente devido a heterogeneidade ambiental da Caatinga. Acreditamos que os efeitos dos distúrbios responderam de maneiras distintas de acordo com a interação com outras variáveis.

Através das investigações etnobotânicas apontamos a necessidade de mais estudos que integrem conhecimento ecológico local a outros campos da Ciência, especialmente a ecologia, conservação e gestão de recursos. Além disso, no decorrer do desenvolvimento desta tese, diversos questionamentos surgiram e podem ser possíveis propostas para investigações futuras. Destacamos a necessidade de compreender a influência dos gradientes climáticos na seleção de espécies, bem como, entender de que maneira os critérios de seleção de espécies foram moldados ao longo do tempo. Dessa forma, entenderemos quais as espécies e ou serviços são prioritários no cenário atual.

6.5 Orçamento

O presente trabalho foi realizado com apoio do Conselho Nacional de Desenvolvimento Científico e Tecnológico (CNPq) (nº 42 490/2018-6) e Pró-Reitoria de Pós-Graduação e Pesquisa -UEPB (2.05.03.00-8-375 / 2017-1). Além disso, a Agência de fomento Fundação de Apoio à Pesquisa do Estado da Paraíba (FAPESq) forneceu bolsa de estudos a Maiara Bezerra Ramos. A Universidade Estadual da Paraíba também ofereceu apoio através da liberação de transporte para as expedições à campo. As despesas para as coletas dos dados incluem: material de campo, incluindo material permanente e material de consumo, alimentação e hospedagens. Com esse investimento foram realizadas: reconhecimento e marcação das áreas de interesse, levantamento florístico e fitossociológico das comunidades vegetais, coleta de variáveis ambientais, mensuração de métricas de distúrbios antrópicos crônicos, coleta de material botânico, 120 entrevistas com agropastores, moradores da região do Cariri, no estado da Paraíba, Brasil.

6.6 Referencias bibliográficas

ALBUQUERQUE, U. P. et al. *Methods and Techniques in Ethnobiology and Ethnoecology*. [S.l.]: Springer New York, 2014.

ALBUQUERQUE, Ulysses Paulino et al. The influence of the environment on natural resource use: Evidence of apparency. *Evolutionary Ethnobiology*, 23 set. 2015. p. 131–147.

ALBUQUERQUE et al. Humans as niche constructors: Revisiting the concept of chronic anthropogenic disturbances in ecology. *Perspectives in Ecology and Conservation*, 2018. v. 16, n. 1, p. 1–11.

ALBUQUERQUE; FERREIRA JÚNIOR, W. S. What Do We Study in Evolutionary Ethnobiology? Defining the Theoretical Basis for a Research Program. *Evolutionary Biology*, 2017. v. 44, n. 2, p. 206–215.

ALMEIDA, C. F. C. B. R. DE et al. Life strategy and chemical composition as predictors of the selection of medicinal plants from the caatinga (Northeast Brazil). *Journal of Arid Environments*, 1 jul. 2005. v. 62, n. 1, p. 127–142.

ANDERSON., E. N. et al. *Ethnobiology*. [S.l.]: [s.n.], 2011.

ANTONGIOVANNI, M. et al. Chronic anthropogenic disturbance on Caatinga dry forest fragments. *Journal of Applied Ecology*, 5 jul. 2020. p. 1365- 2664.13686. Disponível em: <<https://onlinelibrary.wiley.com/doi/abs/10.1111/1365-2664.13686>>. Acesso em: 24 ago. 2020.

APGAUA, D. M. G. et al. Floristic variation within seasonally dry tropical forests of the Caatinga Biogeographic Domain, Brazil, and its conservation implications. *International Forestry Review*, 2015. v. 17, n. Special Issue 2, p. 33–44.

ARAÚJO FILHO, J. C. et al. *Pedologia: solos dos biomas brasileiros*. ESALQ, Sociedade Brasileira de Ciência do Solo, Viçosa-MG, 2017. p. 227–260. Disponível em: <https://www.sbcs.org.br/loja/index.php?route=product/product&product_id=131>. Acesso em: 4 maio 2023.

ARRUDA, D. M. et al. Landforms and soil attributes determine the vegetation structure in the Brazilian semiarid. *Folia Geobotanica*, 2015. v. 50, n. 3, p. 175–184.

BENNETT, E. M. et al. Linking biodiversity, ecosystem services, and human well-being: three challenges for designing research for sustainability. *Current Opinion in Environmental Sustainability*, 2015. v. 14, n. June, p. 76–85.

BERKES, F.; COLDING, J.; FOLKE, C. Rediscovery of Traditional Ecological Knowledge as Adaptive Management. *Ecological Applications*, out. 2000. v. 10, n. 5, p. 1251.

BOWLER, D. E. et al. Mapping human pressures on biodiversity across the planet uncovers anthropogenic threat complexes. *People and Nature*, 27 jun. 2020. v. 2, n. 2, p. 380–394. Disponível em: <<https://onlinelibrary.wiley.com/doi/abs/10.1002/pan3.10071>>. Acesso em: 4 set. 2020.

BROWN, K. A. et al. Assessing natural resource use by forest-reliant communities in madagascar using functional diversity and functional redundancy metrics. *PLoS ONE*, 2011. v. 6, n. 9.

C, T. et al. Traditional ecological knowledge underlying herding decisions of pastoralists. *Animal : an international journal of animal bioscience*, 1 abr. 2018. v. 12, n. 4, p. 831–843. Disponível em: <<https://pubmed.ncbi.nlm.nih.gov/28849752/>>. Acesso em: 11 jul. 2021.

CÁCERES, D. M. et al. The social value of biodiversity and ecosystem services from the perspectives of different social actors. *Ecology and Society*, 2015. v. 20, n. 1.

CALZADA, Leonardo et al. Lands at risk: Land use/land cover change in two contrasting tropical dry regions of Mexico. *Applied Geography*, 1 out. 2018. v. 99, p. 22–30.

CÁMARA-LERET, R. et al. Fundamental species traits explain provisioning services of tropical American palms. *Nature Plants*, 2017. v. 3, n. January, p. 1–7.

CUNHA, S. S. Da et al. Vegetation cover and seasonality as indicators for selection of forage resources by local agro-pastoralists in the Brazilian semiarid region. *Scientific reports*, 2022. v. 12, n. 1, p. 15174. Disponível em: <<https://doi.org/10.1038/s41598-022-18282-w>>.

DEXTER, K. G. et al. Inserting tropical dry forests into the discussion on biome transitions in the tropics Phylogeny of 497 Amazonian tree genera from Evolutionary heritage influences Amazon tree ecology View project Climate Change Effects on Distribution and Preservation of Atlantic Forest View project. 2018. Disponível em: <www.frontiersin.org>.

DÍAZ, S.; LAVOREL, S.; BELLO, F. DE; et al. Incorporating plant functional diversity effects in ecosystem service assessment. *Proceedings of the National Academy of Sciences of the United States of America*, 2007. v. 104, n. 52, p. 20684–20689.

DÍAZ; LAVOREL; MCINTYRE, S.; et al. Plant trait responses to grazing – a global synthesis. *Global Change Biology*, 1 fev. 2007. v. 13, n. 2, p. 313–341. Disponível em: <<https://onlinelibrary.wiley.com/doi/full/10.1111/j.1365-2486.2006.01288.x>>. Acesso em: 10 fev. 2022.

DÍAZ et al. Linking functional diversity and social actor strategies in a framework for interdisciplinary analysis of nature’s benefits to society. *Proceedings of the National Academy of Sciences of the United States of America*, 1 jan. 2011. v. 108, n. 3, p. 895–902. Disponível em: <[pmc/articles/PMC3024663/](https://pubmed.ncbi.nlm.nih.gov/21111111/)>. Acesso em: 11 jan. 2023.

DÍAZ et al. Assessing nature’s contributions to people: Recognizing culture, and diverse sources of knowledge, can improve assessments. *Science*, 2018. v. 359, n. 6373, p. 270–272.

DORNELAS, M. Disturbance and change in biodiversity. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 27 nov. 2010. v. 365, n. 1558, p. 3719–3727. Disponível em: <<https://royalsocietypublishing.org/doi/10.1098/rstb.2010.0295>>. Acesso em: 26 out. 2020.

DUBUIS, A. et al. Improving the prediction of plant species distribution and community composition by adding edaphic to topo-climatic variables. *Journal of Vegetation Science*, 2013. v. 24, n. 4, p. 593–606.

ELLIS, E. C. Anthropogenic transformation of the terrestrial biosphere. *Philosophical Transactions of the Royal Society A: Mathematical, Physical and Engineering Sciences*, 13 mar. 2011. v. 369, n. 1938, p. 1010–1035. Disponível em: <<https://royalsocietypublishing.org/doi/abs/10.1098/rsta.2010.0331>>. Acesso em: 30 maio 2022.

_____. Ecology in an anthropogenic biosphere. *Ecological Monographs*, 2015. v. 85, n. 3, p. 287–331.

_____; RAMANKUTTY, N. Putting people in the map: Anthropogenic biomes of the world. *Frontiers in Ecology and the Environment*. Ecological Society of America.

ESTOMBA, D.; LADIO, A.; LOZADA, M. Medicinal wild plant knowledge and gathering patterns in a Mapuche community from North-western Patagonia. *Journal of ethnopharmacology*, 3 jan. 2006. v. 103, n. 1, p. 109–119. Disponível em: <<https://pubmed.ncbi.nlm.nih.gov/16157460/>>. Acesso em: 4 maio 2022.

FARIAS, R. et al. Does the Local Availability of Woody Caatinga Plants (Northeastern Brazil) Explain Their Use Value ? Author (s): Reinaldo Farias Paiva de Lucena , Elcida de Lima Araújo and Ulysses Paulino de Albuquerque Published by : Springer on behalf of New York B. 2007. v. 61, n. 4, p. 347–361.

FERNÁNDEZ-GIMÉNEZ, M. E. et al. Using an integrated social-ecological analysis to detect effects of household herding practices on indicators of rangeland resilience in Mongolia. *Environmental Research Letters*, 6 jul. 2018. v. 13, n. 7, p. 075010. Disponível em: <<https://doi.org/10.1088/1748-9326/aacf6f>>. Acesso em: 31 ago. 2020.

FIGUEROA, F; CALZADA, L; MEAVE, J A. Integrating pattern-based modelling and political ecology in land-use change research: the case of Mexican dry tropics. *Journal of Land Use Science*, 2020. v. 15, n. 2–3, p. 252–269. Disponível em: <<https://doi.org/10.1080/1747423X.2019.1681527>>.

FOLKE, C. Resilience: The emergence of a perspective for social–ecological systems analyses. *Global Environmental Change*, 1 ago. 2006. v. 16, n. 3, p. 253–267.

GAOUE, O. et al. Optimal harvesting strategies for timber and non-timber forest products in tropical ecosystems. 2016. n. July, p. 286–297.

GIBB, H. et al. Climate mediates the effects of disturbance on ant assemblage structure. *Proceedings of the Royal Society B: Biological Sciences*, 20 maio. 2015. v. 282, n. 1808.

GUREVITCH, J.; SCHEINER, S. M.; FOX, G. A. *Ecologia vegetal*. 2009.

HUDSON, A. et al. Natural Plant Resources for Sustainable Development: Insights from Community Use in the Chimanimani Trans-Frontier Conservation Area, Mozambique. *Human Ecology*, 2020. v. 48, n. 1, p. 55–67.

ISBELL, F. et al. Linking the influence and dependence of people on biodiversity across scales. *Nature* 2017 546:7656, 1 jun. 2017. v. 546, n. 7656, p. 65–72. Disponível em: <<https://www.nature.com/articles/nature22899>>. Acesso em: 1o fev. 2023.

- JENNY, H.; AMUNDSON, R. Factors of soil formation. System of Quantitative Pedology. Library of Congress Cataloging-in-Publication Data. [S.l.]: Dover Publications, 1941.
- KELLY, R. et al. Ten tips for developing interdisciplinary socio-ecological researchers. *Socio-Ecological Practice Research*, jun. 2019. v. 1, n. 2, p. 149–161.
- LA RIVA, E. G. DE et al. Disentangling the relative importance of species occurrence, abundance and intraspecific variability in community assembly: A trait-based approach at the whole-plant level in Mediterranean forests. *Oikos*, 2016. v. 125, n. 3, p. 354–363.
- LAUGHLIN, D. C. et al. Environmental Filtering and Positive Plant Litter Feedback Simultaneously Explain Correlations Between Leaf Traits and Soil Fertility. *Ecosystems*, 2015. v. 18, n. 7, p. 1269–1280.
- LEBRIJA-TREJOS, E. et al. Pathways, mechanisms and predictability of vegetation change during tropical dry forest succession. *Perspectives in Plant Ecology, Evolution and Systematics*, 2010. v. 12, n. 4, p. 267–275.
- LESO, L. K. et al. Ethnobotany at a local scale: Diversity of knowledge of medicinal plants and assessment of plant cultural importance in the Polokwane local municipality, South Africa. *Botany Letters*, 2017. v. 164, n. 1, p. 93–102.
- LETCHER, S. G. et al. Environmental gradients and the evolution of successional habitat specialization: A test case with 14 Neotropical forest sites. *Journal of Ecology*, 2015. v. 103, n. 5, p. 1276–1290.
- LOPES, S. . The other side of Ecology: thinking about the human bias in our ecological analyses for biodiversity conservation. *Ethnobiology and Conservation*, 2017. v. 6, n. 19, p. 1–8. Disponível em: <<http://ethnobiococonservation.com/index.php/ebc/article/view/103>>.
- MARTORELL, C.; PETERS, E. M. The measurement of chronic disturbance and its effects on the threatened cactus *Mammillaria pectinifera*. *Biological Conservation*, 2005. v. 124, n. 2, p. 199–207.
- MCGILL, B. J. et al. Rebuilding community ecology from functional traits. *Trends in Ecology and Evolution*, 2006. v. 21, n. 4, p. 178–185.

- MEDEIROS, P. M. DE et al. Why do people use exotic plants in their local medical systems? A systematic review based on Brazilian local communities. *PLoS ONE*, 2017. v. 12, n. 9.
- MÉNDEZ-TORIBIO, M. et al. Effects of slope aspect and topographic position on environmental variables, disturbance regime and tree community attributes in a seasonal tropical dry forest. *Journal of Vegetation Science*, 2016. v. 27, n. 6, p. 1094–1103.
- MEYNARD, C. N. et al. Disentangling the drivers of metacommunity structure across spatial scales. *Journal of Biogeography*, 8 ago. 2013. v. 40, n. 8, p. 1560–1571. Disponível em: <[pmc/articles/PMC4000944/](https://pubmed.ncbi.nlm.nih.gov/2400944/)>. Acesso em: 28 maio 2022.
- MIGUEL, M. F. et al. Context-dependency and anthropogenic effects on individual plant–frugivore networks. *Oikos*, 2018. v. 127, n. 7, p. 1045–1059.
- MILLENNIUM ECOSYSTEM ASSESSMENT. Ecosystem and human well-being : synthesis. World Resources Institute, 2005. p. 86. Disponível em: <<https://wedocs.unep.org/handle/20.500.11822/8755>>. Acesso em: 30 maio 2022.
- MORO, M. F. et al. The role of edaphic environment and climate in structuring phylogenetic pattern in seasonally dry tropical plant communities. *PLoS ONE*, 2015. v. 10, n. 3, p. 1–18.
- _____ et al. A Phylogeographical Metaanalysis of the Semiarid Caatinga Domain in Brazil. *Botanical Review*, 2016. v. 82, n. 2, p. 91–148. Disponível em: <<http://dx.doi.org/10.1007/s12229-016-9164-z>>.
- NIU, K. et al. Grazing increases functional richness but not functional divergence in Tibetan alpine meadow plant communities. *Biodiversity and Conservation*, 2016. v. 25, n. 12, p. 2441–2452.
- NUNES, A. T. et al. Plants used to feed ruminants in semi-arid Brazil: A study of nutritional composition guided by local ecological knowledge. *Journal of Arid Environments*, 2016. v. 135, p. 96–103.
- ODLING-SMEE, F. J.; LALAND, K. N.; FELDMAN, M. W. Niche Construction. *Niche Construction*, 31 dez. 2013.

PEDROSA, K. M. et al. Plants with similar characteristics drive their use by local populations in the semi-arid region of Brazil. *Environment, Development and Sustainability*, 1 nov. 2021. v. 23, n. 11, p. 16834–16847.

PERIAGO, M. E. et al. Combining ecological aspects and local knowledge for the conservation of two native mammals in the Gran Chaco. *Journal of Arid Environments*, 2017. v. 147, n. August, p. 54–62. Disponível em: <<https://doi.org/10.1016/j.jaridenv.2017.07.017>>.

PETCHEY, O. L.; GASTON, K. J. Functional diversity: Back to basics and looking forward. *Ecology Letters*, 2006. v. 9, n. 6, p. 741–758.

PUTTEN, W. H. VAN DER et al. Plant–soil feedbacks: the past, the present and future challenges. *Journal of Ecology*, 1 mar. 2013. v. 101, n. 2, p. 265–276. Disponível em: <<https://onlinelibrary.wiley.com/doi/full/10.1111/1365-2745.12054>>. Acesso em: 4 maio 2023.

RAMOS, M. A. et al. Can wood quality justify local preferences for firewood in an area of caatinga (dryland) vegetation? *Biomass and Bioenergy*, 1 jun. 2008. v. 32, n. 6, p. 503–509.

RAMOS, M. B. et al. The role of edaphic factors on plant species richness and diversity along altitudinal gradients in the Brazilian semi-arid region. *Journal of Tropical Ecology*, 25 set. 2020. v. 36, n. 5, p. 199–212. Disponível em: <https://www.cambridge.org/core/product/identifier/S0266467420000115/type/journal_article>.

RIBEIRO, E. M. . et al. Chronic anthropogenic disturbance drives the biological impoverishment of the Brazilian Caatinga vegetation. *Journal of Applied Ecology*, 2015. v. 52, n. 3, p. 611–620.

SANDE, M. T. VAN DER et al. Abiotic and biotic drivers of biomass change in a Neotropical forest. *Journal of Ecology*, 2017. v. 105, n. 5, p. 1223–1234.

SARMENTO, C. D. E; FRANCA, M. G. C. Neotropical Forests from their Emergence to the Future Scenario of Climatic Changes. *Vegetation*, 2018.

SCHULZ, K. et al. Grazing deteriorates the soil carbon stocks of Caatinga forest ecosystems in Brazil. *Forest Ecology and Management*, 2016. v. 367, p. 62–70. Disponível em: <<http://dx.doi.org/10.1016/j.foreco.2016.02.011>>.

_____ et al. Grazing reduces plant species diversity of Caatinga dry forests in northeastern Brazil. *Applied Vegetation Science*, 1 abr. 2019. v. 22, n. 2, p. 348–359. Disponível em: <<https://onlinelibrary.wiley.com/doi/full/10.1111/avsc.12434>>. Acesso em: 22 fev. 2023.

SILVA, A. C.; SOUZA, A. F. Aridity drives plant biogeographical sub regions in the Caatinga, the largest tropical dry forest and woodland block in South America. *PLoS ONE*, 1 abr. 2018. v. 13, n. 4.

SILVA OLIVEIRA, R. C. DA et al. Ethnobotany and Harvesting Impacts on Candombá (*Vellozia* aff. *sincorana*), A Multiple Use Shrub Species Endemic to Northeast Brazil1. *Economic Botany*, 2015. v. 69, n. 4, p. 318–329.

SILVERTOWN, J. Plant coexistence and the niche. *Trends in Ecology and Evolution*.

SMITH, B. D. Resource resilience, human niche construction, and the long-term sustainability of pre-Columbian subsistence economies in the Mississippi River Valley corridor. *Journal of Ethnobiology*, 2009. v. 29, n. 2, p. 167–183.

TABARELLI, M.; PERES, C. A.; MELO, F. P. L. The “few winners and many losers” paradigm revisited: Emerging prospects for tropical forest biodiversity. *Biological Conservation*, 2012. v. 155, p. 136–140. Disponível em: <<http://dx.doi.org/10.1016/j.biocon.2012.06.020>>.

VELLEND, M. Conceptual synthesis in community ecology. *Quarterly Review of Biology*, 2010. v. 85, n. 2, p. 183–206.

WOLVERTON, S. Ethnobiology 5: Interdisciplinarity in an era of rapid environmental change. *Ethnobiology Letters*, 2013. v. 4, n. 1, p. 21–25.

ANEXOS

ANEXO I



CENTRO DE ENSINO
SUPERIOR E
DESENVOLVIMENTO



COMPROVANTE DE ENVIO DO PROJETO

DADOS DO PROJETO DE PESQUISA

Título da Pesquisa: EFEITO DOS IMPACTOS ANTRÓPICOS NA MONTAGEM DE COMUNIDADES VEGETAIS NO ECOSISTEMA DA CAATINGA

Pesquisador: Maiara Bezerra Ramos

Versão: 2

CAAE: 07564918.2.0000.5175

Instituição Proponente: Universidade Estadual da Paraíba - UEPB

DADOS DO COMPROVANTE

Número do Comprovante: 010713/2019

Patrocinador Principal: Universidade Estadual da Paraíba - UEPB

Informamos que o projeto EFEITO DOS IMPACTOS ANTRÓPICOS NA MONTAGEM DE COMUNIDADES VEGETAIS NO ECOSISTEMA DA CAATINGA que tem como pesquisador responsável Maiara Bezerra Ramos, foi recebido para análise ética no CEP Centro de Ensino Superior e Desenvolvimento-CESED/PB em 11/02/2019 às 16:56.

Endereço:	SENADOR ARGEMIRO DE FIGUEIREDO 1901				
Bairro:	ITARARE	CEP:	58.411-020		
UF:	PB	Município:	CAMPINA GRANDE		
Telefone:	(83)2101-8857	Fax:	(83)2101-8857	E-mail:	cep@cesed.br

ANEXO II



Ministério do Meio Ambiente - MMA
 Instituto Chico Mendes de Conservação da Biodiversidade - ICMBio
 Sistema de Autorização e Informação em Biodiversidade - SISBIO

Comprovante de registro para coleta de material botânico, fúngico e microbiológico

Número: 68447-1	Data da Emissão: 13/07/2019 23:05:00
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Dados do titular

Nome: Malara Bezerra Ramos	CPF: 094.771.594-09
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SISBIO

Observações e ressalvas

1	Este documento não abrange a coleta de vegetais hidróbios, tendo em vista que o Decreto-Lei nº 221/1967 e o Art. 36 da Lei nº 9.605/1998 estabelecem a necessidade de obtenção de autorização para coleta de vegetais hidróbios para fins científicos.
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Taxons autorizados

#	Nível taxonômico	Taxon(s)
1	Reino	Plantae

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APÊNDICES

APÊNDICE A: Registro de nova ocorrência de espécie da Família Euphorbiaceae

(Manuscrito publicado no periódico Acta brasiliensis)

Link de acesso para o artigo:

<http://revistas.ufcg.edu.br/ActaBra/index.php/actabra/article/view/378>

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First record of *Cnidoscolus obtusifolius* Pohl (Euphorbiaceae) for Paraíba State, northeastern Brazil

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Abstract

Cnidoscolus obtusifolius Pohl (Euphorbiaceae), species so far known from Minas Gerais, Bahia, Alagoas and Pernambuco States in Brazil is reported for the first time for the State of Paraíba, in the northeastern region of the country. Specimens of this taxon were collected in a fragmented area considered a Caatinga vegetation relict, where total annual precipitation is 700 mm on average and elevation of 644 m a.s.l. The records were made in September and October 2019, when the species was in fertile stage as it bore flowers and fruits. Here we provide a description of its morphology along with taxonomic comments, data on the geographical range and detailed images of the species.

Keywords: Caatinga; diversity; floristics; malpighiales

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Abstract

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Keywords: Caatinga; diversity; floristics; malpighiales

Primeiro registro de *Cnidoscolus obtusifolius* Pohl (Euphorbiaceae) no estado da Paraíba, nordeste do Brasil

Resumo

Cnidoscolus obtusifolius Pohl (Euphorbiaceae) espécie até então conhecida para os Estados de Minas Gerais (Sudeste), Bahia, Alagoas e Pernambuco (Nordeste), Brasil, está sendo registrada pela primeira vez no Estado da Paraíba, nordeste do Brasil. A espécie foi coletada em uma área residual, que representa um relicto para a vegetação da Caatinga, com altitude de 644 m.s.n.m e regime pluviométrico de 700 mm anuais. Os registros foram obtidos nos meses de Setembro e Outubro de 2019, onde a espécie se encontrava florida e frutificada, a identificação foi baseada

na literatura especializada. Apresentamos descrição morfológica, comentários taxonômicos, dados de distribuição geográfica e imagens detalhadas para a espécie.

Palavras-chave: Caatinga, diversidade, florística, malpighiales.

Introduction

Euphorbiaceae, a belonging family in the order Malpighiales, comprises 334 genera (Webster, 1994b) and nearly 8,000 species (Radcliffe-Smith & Esser, 2001). Representatives of this family occur in different plant formations in the Neotropics, but their distribution is mainly associated with tropical and subtropical regions, particularly in Africa and the Americas, with only a few genera and species occurring in temperate regions. The family is recognized as being one the most complex and morphologically diverse taxonomic groups within Rosidae (Cronquist & Takhtadzhian, 1981), as it exhibits all growth habits, ranging from small herbs to large size trees, in which the presence of latex is a frequent trait (Cronquist & Takhtadzhian, 1981; APG IV, 2016).

In Brazil, Euphorbiaceae is represented by 950 species distributed in 64 genera, of which four genera and 633 species are endemic in the country. Despite having a high number of species and genera, the family's diversity is still little known (Secco *et al.*, 2012). This family has a wide geographical distribution, and is associated with all phytogeographic domains. Euphorbiaceae is considered an important group plants in the Brazilian semiarid (Alcofora do Filho *et al.* 2003, Andrade *et al.* 2004, Sátiro & Roque 2008, Lucena 2009). In the Caatinga domain, the family is represented by 228 species, distributed in 31 genera (Flora do Brasil 2020, under construction). However, it remains relatively little studied, since most works are centered only in the states of Pernambuco and Bahia (Secco *et al.*, 2012).

Studies of the flora of Pernambuco such as those of Lucena (2009) and Lucena and Alves (2010) have contributed to the family's knowledge and distribution in the Northeast and Caatinga. Lucena and Alves (2009) added 29 occurrences to the list of Euphorbiaceae species for the Brazilian Northeast. Despite the advances provided by these studies, other work needs to be done; in order to cover the entire Northeast of Brazil to increase knowledge Euphorbiaceae diversity in the Caatinga and other northeastern ecosystems, in addition to contributing to other areas of knowledge.

Cnidoscolus Pohl is a prominent genus of Euphorbiaceae. This genus comprises 97 species exclusive of the Americas (Maya-Lastra & Steinmann, 2018), where its main

diversification centers are located in Brazil, with 42 species among which 37 are endemic to the country (Flora do Brasil 2020, under construction), and Mexico, with 25 species (Maya-Lastra & Steinmann, 2019). To this date, Müller (1873) monograph remains the most comprehensive treatment on *Cnidoscolus* for Brazil, whereas Melo and Sales (2008) study has this position for the country's Northeast. With regard to the Brazilian species of this genus, 22 of them have been recorded in Caatinga vegetation, with 19 species being endemic to this floristic domain (Flora do Brasil 2020, under construction). It is worth mentioning that the genus stands out for presenting species with different therapeutic purposes and widespread in popular medicine (Moura *et al.*, 2019).

Representatives of *Cnidoscolus* are characterized by the presence of simple, alternate leaves, which are membranaceous to subcoriaceous, glabrous to velutinous-tomentose; their flowers are sessile or pedicellate, staminate ones are distal, with 10-15 free stamens organized in 2-4 whorls, while pistillate flowers are proximal; the ovary is ovoid to pyriform, glabrous to velutinous, with one ovule per locule and three styles, free or adnate at the base. The nectariferous disc is extrastaminal and annular; the presence of urticating trichomes in almost all vegetative and floral parts is particularly important as this is one of the diagnostic features of this genus (Pohl, 1827; Melo & Sales, 2008).

Among the species of this genus, *Cnidoscolus obtusifolius* Pohl was known from Minas Gerais, Bahia, Alagoas and Pernambuco States in Brazil. Now was reported for the first time for the State of Paraíba. While conducting fieldwork in the municipality of São João do Tigre, in the Brazilian semiarid, it was verified that collections of the species *Cnidoscolus obtusifolius* (Euphorbiaceae) became the first record for the state's flora, which is reported here.

Caatinga, as it is known for being the main plant formation in the Brazilian semiarid region, presents predominantly a vegetation classified as Seasonally Dry Tropical Forest, being considered the most diverse dry forest in the world (Moro *et al.*, 2015; Queiroz *et al.*, 2017). The Caatinga is marked by the influence of spatial and temporal variability of rain and landscapes, with different types of soils and a high number of endemisms (Moro *et al.* 2015; Queiroz *et al.* 2017; Silva *et al.* 2017). The results presented here were obtained in the municipality of São João do Tigre. (*município*) is located in the Cariri microrregion of Paraíba, Borborema Plateau (8°10' 27.5" S, 36° 55' 00.2" W), South extreme of Paraíba State, immersed in the Brazilian semiarid region (Mascarenhas *et al.*, 2005). This *município* is located in the portion of the Paraiban Cariri where the highest elevations are found with maxima of around

1,200 m a.s.l. (Souza *et al.*, 2009). The regional climate is of the Bsh type, that is hot, semi-arid (Alvares *et al.*, 2013), with mean annual precipitation ranging around 700 mm, and mean annual temperature around 26 °C (Mascarenhas *et al.*, 2005; Moro *et al.*, 2015). With regard to geology, the study area is inserted in the domain of the crystalline, characterized in general by shallow, rocky and highly fertile soils. Specifically the Cariri region of Paraíba state, has a total area of approximately 11,225.736 km², the region encompasses relictual reliefs that witness the pediplanation processes developed in the area; the landscape is dominated by narrow valleys and dry slopes, in addition to multiple mountain ranges, which combined with higher levels of precipitation are responsible for the occurrence of more benign temperatures in the region (Mascarenhas *et al.*, 2005).

Fertile material of *Cnidoscolus obtusifolius* was collected during two expeditions conducted between September and October 2019. The vouchers were deposited at the Manuel de Arruda Câmara Herbarium (HACAM) of the Universidade Estadual da Paraíba (Paraíba State University), Campina Grande, Paraíba, Brazil. For the taxonomic identification, we used specialized literature (Müller Argoviensis, 1873; Melo & Sales, 2008), in addition to consultation of the electronic data bases of Reflora, the Herbário Virtual da Flora e dos Fungos (Virtual Flora and Fungal Virtual Herbarium), Jstor (<https://plants.jstor.org/>), and Tropicos (<https://www.tropicos.org/>).

Here, we present morphological description for the taxon, taxonomic and distribution notes, and illustrations (Figure 2). In the field we collected material from several individuals, and for all characteristics we analyzed three to five samples.

Cnidoscolus obtusifolius Pohl, Pl. Bras. Icon. Descr. 1: 62. 1827.

Type: BRAZIL: In deserto, Bahiensis prov., *C.F.P. von Martius* 2230 (Holotype: M0233240!)

Tree, 6-8 m tall. Urticating trichomes acicular (stinging), 0.5–4 mm long, covering the branches; stinging trichomes of 2–6 mm long covering in the petioles, leaf blade, inflorescence, perianth, and fruits. Stems cylindrical, grayish to reddish, glabrous to pubescent at their distal branches. Stipules 1–3 × 1–2 cm irregularly triangular, generally deciduous, sometimes persistent after the leaves are shed, margins glandulous-papillose, pubescent to puberulent. Petiole absent or, when present, up to 1.8 cm long, cylindrical, pubescent to velutinous; petiolar glands absent. Leaf blade 3–9 × 2–5.2 cm, membranaceous to chartaceous, entire, elliptic, obovate, base obtuse to subcordate, apex acuminate to rounded, margin crenate, with urticating trichomes at each border, ciliate and undulate, adaxial face pubescent with urticating trichomes

on primary veins; abaxial face sparsely velutinous with urticating trichomes on primary and secondary veins in mature (senescent) leaves only; leaf venation craspedodromous and brochidodromous, main veins 3–6. Inflorescence an imperfect dichasium 2.0×2.5 –3 cm, or with 3 dichasia arranged in 2 levels of branching, puberulous, hispidulous-pubescent to sericeous, peduncle 0.3–0.9 (1.0 cm); bracts and bracteoles 3–4 mm, triangular, with glandulous-papillose margins at the base, pubescent. Staminate flowers sessile, located from the second level of ramification; perianth ca. 16 mm, tubular-campanulate or tubular hypocrateriform, greenish white, pubescent; tube $3\text{--}4 \times 10$ mm; lobes $8\text{--}9 \times 7$ mm, oblong-elliptic, apex rounded; stamens 18 arranged in 2-3 whorls, partially joined forming a column 2–10 mm, glabrous; filaments 3-6 mm; anthers $1 \times 0.25\text{--}0.5$ mm, dorsifixed, oblong; staminodes 3, filiform, ca. 5 mm; glandular disc 0.5×1 mm, annular, slightly lobed internally, glabrous. Pistillate flowers 3, located up to the second level of branching; pedicel ca. 2 mm; receptacle/base persistent of inconspicuous perianth; perianth ca. 13 mm, white, externally greenish, segments free, oblong to oval, pubescent and urticating, apex rounded; ovary ovoid, sub-cylindrical in cross section, pubescent to hispidulous-sericeous; styles 3–4 mm; joined at the base, four- to many-branched, 12-32 stigmatic branches; glandular disc annular striated, glabrous. Capsule $1.5\text{--}2.2 \times 1.1\text{--}1.3$ cm, $1.1\text{--}1.5 \times 1.1\text{--}1.2$, loculicidal and septicida, sub-pyriform, cylindrical in cross section, velutinous, urticating, apex acuminate, base of the styles persistent. Seeds $1.1\text{--}1.2 \times 0.4\text{--}0.5$ mm, oblong, concave dorsally, convex or flat ventrally, surface brown with prominent darker spots, base retuse, apex acute; caruncle $4\text{--}5 \times 3$ mm.

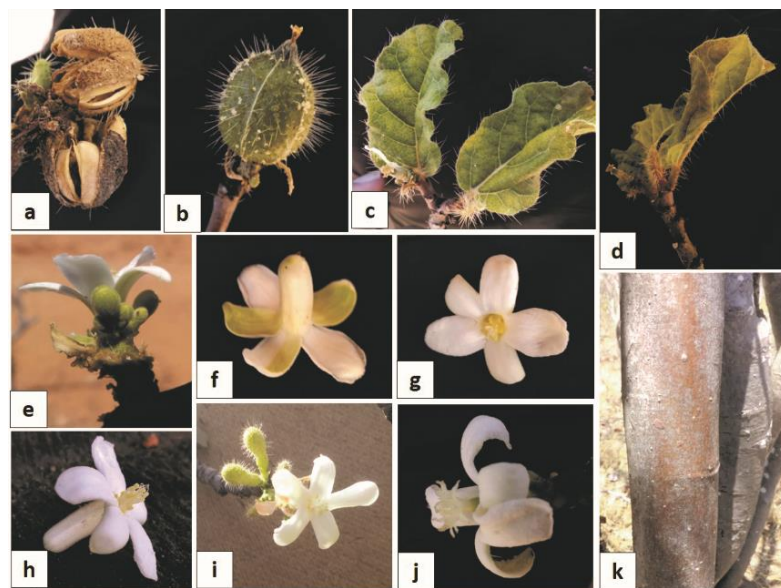


Figure 2. First record of *Cnidoscolus obtusifolius* Pohl in the estate of Paraíba, Brazil. A-B, capsule; C-D, leaves; E, inflorescence; F-I, male flower; J, female flower; K, stem.

Notes: For this genus other three species have been recorded in Paraíba: *Cnidoscolus infestus* Pax & K. Hoffm., *Cnidoscolus quercifolius* Pohl and *Cnidoscolus urens* (L.) Arthur. *Cnidoscolus obtusifolius* can be easily differentiated from these congeneric species mainly by its leaves with entire, elliptical and obovate blade, with an obtuse and sub-cordate base, its irregularly triangular, pubescent to puberulous stipules with papillose-glandulose margin, and its velutinous capsule.

Geographic range and habitat: Occurring exclusively in Brazil, this species is restricted to Seasonally Dry Tropical Forest and Seasonal Deciduous Forest being registered presence so far in the states of Minas Gerais (Southeastern Region), Alagoas, Bahia and Pernambuco (Northeastern Region) (Melo & Sales, 2008; Flora do Brasil 2020, under construction). Constituting, in this work, the first record for the State of Paraíba. Collections of this species in Paraíba were in the dry season 2019, in a fragment of Caatinga vegetation with an open physiognomy, at an elevation of 644 m a.s.l. According to World Reference Base for soil resource, soil classification prevailing soils in the study area are classified as Chromic Luvisols (IUSS Working Group WRB 2014).

The area shows signs of grazing by goats and cattle, common domestic animals and considered important drivers in the biological impoverishment of the Caatinga; which highlights the susceptibility to which *C. obtusifolius* is found. The area in which it was found is considered a fragment of shrubby Caatinga, and *C. obtusifolius* is among the five most abundant species, with random spatial distribution.

Along with *C. obtusifolius*, *Mimosa ophthalmocentra* Mart. ex Benth. (Fabaceae), *Cenostigma pyramidale* (Tul.) Gagnon & G.P. Lewis (Fabaceae), and *Aspidosperma pyriforme* Mart. & Zucc. (Apocynaceae) are most common species in the area.

Reproductive phenology: Found in flower and fruit in September and October for the observed specimens.

Examined material: BRAZIL. Paraíba: São João do Tigre municipality, (8° 10' 27.5" S, 36° 55' 00.2" W, 644 m a.s.l, 4 Oct 2019, M.B. Ramos, M.G.R. Maciel (HACAM 1978).

This new record of this specie was from in area very close to the border with Pernambuco (state where this taxon was reported before). Thus this, record underlines the importance of conducting further floristic studies in the Brazilian semiarid, since this region

has a high environmental heterogeneity, which reflect the high number of endemism, species richness and variability in the plant formations and paradoxically, the knowledge about its flora is still incipient when compared to the other world plant formations.

The finding presented here refers to a Brazilian endemic species, restricted to the semiarid environment, and thus, environmental education actions about the specie and possible management plans are needed, aimed at the conservation and maintenance of high richness of Caatinga.

References

- Alcoforado-Filho, F. G., Sampaio, E. V. D. S. B., & Rodal, M. J. N. (2003). Florística e fitossociologia de um remanescente de vegetação caducifólia espinhosa arbórea em Caruaru, Pernambuco. *Acta botanica brasílica*, 17(2), 287-303. doi: 10.1590/S0102-33062003000200011
- Alvares, C. A., Stape, J. L., Sentelhas, P. C., de Moraes Gonçalves, J. L., J. & Sparovek, G. (2013). Köppen's climate classification map for Brazil. *Meteorologische Zeitschrift*, 22, 711-728. doi: 10.1127/0941-2948/2013/0507
- Andrade, K. V. S. A., Rodal, M. J. N., Lucena, M. D. F. A., & Gomes, A. P. S. (2004). Composição florística de um trecho do Parque Nacional do Catimbau, Buíque, Pernambuco-Brasil. *Hoehnea*, 31(3), 337-348.
- APG IV (Angiosperm Phylogeny Group IV). (2016). An update of the Angiosperm Phylogeny Group classification for the orders and families of flowering plants: APG IV. *Botanical Journal of the Linnean Society*, 181, 1-20. doi: 10.1111/boj.12385
- Cronquist, A. & Takhtadzhian, A. L. (1981). An integrated system of classification of flowering plants. Columbia University Press.
- Flora do Brasil (2020) (em construção). Jardim Botânico do Rio de Janeiro. Available at: <http://floradobrasil.jbrj.gov.br>
- Heywood, V. H., Moore, D. M., Richardson, I. B. K., & Stearn, W. T. (1993). *Flowering plants of the world* (No. 582.13 F644. pp. 335). Oxford University Press.
- IUSS Working Group WRB [Word Reference Base for Soil Resources]. (2014). International soil classification system for naming soils and creating legends for soil maps. Food and Agriculture Organization of the United Nations. IUSS/ISRIC/FAO. *World Soil Resources Reports*, Rome, 106.

- JSTOR – Global Plants (2019). Available at: <https://plants.jstor.org>
- Lucena, M. F. A., & Alves, M. (2010). Notas taxonômicas para Euphorbiaceae s.l do Nordeste do Brasil. *Hoehnea*, 37 (1), 71-85. doi: 10.1590/S2236-89062010000100005
- Lucena, M.F.A. (2009). *Diversidade de Euphorbiaceae s.l. no Nordeste do Brasil*. Tese de Doutorado. Universidade Federal de Pernambuco, Recife.
- Mascarenhas, J. C., Beltrão, B. A., Souza-Junior, L. C., Morais, F., Mendes, V. A. & Miranda, J. L. F. (2005). Serviço Geológico do Brasil. Projeto cadastro de fontes de abastecimento por água subterrânea. Diagnóstico do município de São João do Tigre, Estado da Paraíba. Recife: CPRM/PRODEEM.
- Maya-Lastra, C. A. and Steinmann, V. W. (2018). A nomenclator of *Cnidoscolus* (Euphorbiaceae). *Phytotaxa*, 346, 1-30. doi: 10.11646/phytotaxa.346.1.1
- Maya-Lastra, C. A. and Steinmann, V. W. (2019). Novelties in Mexican *Cnidoscolus* sect. *Calyptrosolen* (Euphorbiaceae). *Systematic Botany*, 44 (2), 339-345. doi: 10.1600/036364419X15562052252081
- Melo, A. L. and Sales, M. F. (2008). O gênero *Cnidoscolus* Pohl (Crotonoideae-Euphorbiaceae) no Estado de Pernambuco, Brasil. *Acta Botanica Brasilica*, 22, 806-827.
- Moro, M. F., Silva, I. A., Araújo, F. S., Lughadha, E. N., Meagher, T. R. & Martins, F. R. (2015). The role of edaphic environment and climate in structuring phylogenetic pattern in seasonally dry tropical plant communities. *PLoS One*, 10(3), e0119166. doi: 10.1371/journal.pone.0119166
- Moura, L. F. W. G., da Silva Neto, J. X., Lopes, T. D. P., Benjamin, S. R., Brito, F. C. R., Magalhães, F. E. A., and Guedes, M. I. F. (2019). Ethnobotanic, phytochemical uses and ethnopharmacological profile of genus *Cnidoscolus* spp.(Euphorbiaceae): A comprehensive overview. *Biomedicine & Pharmacotherapy*, 109, 1670-1679.
- Müller Argoviensis, J. (1873). *Euphorbiaceae*. In: C.F. Martius (ed.). *Flora Brasiliensis*. In: von Martius, C. F. P. & Eichler, A.W. (Eds.) *Flora Brasiliensis*. 11(2), 1.752. F. Fleischer, Munich and Leipzig.
- Pohl, J. E. (1827). *Cnidoscolus*. In: J.E. Pohl. *Plantarum brasiliae icones et descriptiones*. Vindobanae. 1, 56-63
- Radcliffe-Smith, A. and Esser, H. J. (2001). *Genera Euphorbiacearum*. Royal Botanic Gardens, Kew.

Sátiro, L. N. e Roque, N. (2008). A família Euphorbiaceae nas caatingas arenosas do médio rio São Francisco, BA, Brasil. *Acta Botanica Brasilica*, 22, 99-118.

Satiro, L. N., & Roque, N. (2008). The family Euphorbiaceae on the sand dunes of the middle São Francisco River, Bahia State, Brazil. *Acta Botanica Brasilica*, 22(1), 99-118.

Souza, B. I., Suertegaray, D. M. A., & de Lima, E. R. V. (2009). Desertificação e seus efeitos na vegetação e solos do Cariri Paraibano (desertification and its effects over the vegetation and soils of the cariri region of Paraíba–Brazil). *Mercator*, 8(16), 217-232.

Tropicos (2019). Available at: <http://www.tropicos.org/>

Webster, G. L. (1994a). Synopsis of the genera and suprageneric taxa of Euphorbiaceae. *Annals of the Missouri Botanical Garden*, 81, 33-144.

Webster, G. L. (1994b). Classification of the Euphorbiaceae. *Annals of the Missouri Botanical Garden*, 81, 3-32.

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APÊNDICE B: Temo de consentimento livre e esclarecido utilizado para atender

TERMO DE CONSENTIMENTO LIVRE E ESCLARECIDO-TCLE

(OBS: para o caso de pessoas maiores de 18 anos e que não estejam inseridas nas hipóteses de vulnerabilidade que impossibilitam o livre discernimento com autonomia para o exercício dos atos da vida civil).

Pelo presente Termo de Consentimento Livre e Esclarecido eu, _____, em pleno exercício dos meus direitos me disponho a participar da Pesquisa **“Efeito dos impactos antrópicos na montagem de comunidades vegetais no ecossistema da Caatinga”**.

Declaro ser esclarecido e estar de acordo com os seguintes pontos:

O “ Qual o papel dos fatores antrópicos locais na montagem de comunidades vegetais da caatinga? ” tem como objetivo identificar o conhecimento e uso que os moradores de comunidade rurais da Caatinga sobre espécies vegetais tentando elucidar quais os traços funcionais preferenciais pelas populações locais.

Aos voluntários caberá a autorização para solicitamos a sua colaboração para fornecer informações sobre as plantas da região por meio de entrevistas, como também sua autorização para apresentar os resultados deste estudo em eventos da área de ciências ambientais, além de publicar em revista científicas nacionais e internacionais. Por ocasião da publicação dos resultados, seu nome será mantido em e não haverá nenhum risco ou desconforto ao voluntário.

- Ao pesquisador caberá o desenvolvimento da pesquisa de forma confidencial; entretanto, quando necessário for, poderá revelar os resultados ao médico, indivíduo e/ou familiares, cumprindo as exigências da Resolução Nº. 466/12 do Conselho Nacional de Saúde/Ministério da Saúde.
- O voluntário poderá se recusar a participar, ou retirar seu consentimento a qualquer momento da realização do trabalho ora proposto, não havendo qualquer penalização ou prejuízo para o mesmo.
- Será garantido o sigilo dos resultados obtidos neste trabalho, assegurando assim a privacidade dos participantes em manter tais resultados sem caráter confidencial.
- Não haverá qualquer despesa ou ônus financeiro aos participantes voluntários deste projeto científico e não haverá qualquer procedimento que possa incorrer em danos físicos ou

financeiro, são voluntários e, portanto, não haveria necessidade de indenização por parte da equipe científica e/ou da Instituição responsável.

- Qualquer dúvida ou solicitação de esclarecimentos, o participante poderá contatar a equipe científica no número (083) 988032622 Maiara Bezerra Ramos (aluna), Sonaly Silva da Cunha (aluna), DR. Sérgio Faria Lopes (professor).
- Ao final da pesquisa, se for do meu interesse, terei livre acesso ao conteúdo da mesma, podendo discutir os dados, como pesquisador, vale salientar que este documento será impresso em duas vias e uma delas ficará em minha posse.
- Desta forma, uma vez tendo lido e entendido tais esclarecimentos e, por estar de pleno acordo com o teor do mesmo, dato e assino este termo de consentimento livre e esclarecido.

Assinatura do pesquisador responsável

Assinatura do Participante

Assinatura Dactiloscópica do participante da pesquisa

(OBS: utilizado apenas nos casos em que não seja

Possível a coleta da assinatura do participante da pesquisa).

